

## Agricultural Land Conversion in Northwest Iran

Barati, A.A.<sup>1\*</sup>, Asadi, A.<sup>1</sup>, Kalantari, K.<sup>1</sup>, Azadi, H.<sup>2</sup> and Witlox, F.<sup>2</sup>

1- Department of Agricultural Development and Management, University of Tehran, Iran.

2- Department of Geography, Ghent University, Belgium.

### Abstract

Agricultural Land Conversion (ALC) has been introduced as one of the most important factors affecting ecosystem. This type of conversion has led to several challenges in agricultural development and human life. Monitoring ALC plays a crucial role when dealing with such challenges. The main objective of this study was to monitor the trend of ALC in the Qazvin province located in Northwest Iran from 1990 to 2010 using remote sensing data. The results showed that 44,845 ha of agricultural lands (3.03% of the total agricultural lands of the province) were converted to non-agricultural lands, of which, 32,033 and 10,243 ha (2.16% and 0.69% of total agricultural lands of the province), were respectively transformed to saline lands and urban areas and infrastructures. Our projection for 2030 shows that among other uses, the conversion of agricultural lands to the saline lands and urban areas and infrastructures will stay most likely. However, the conversion probability for irrigated and orchard lands to urban areas and infrastructures will be more than the saline lands while the conversion probability for dry and rangelands to the saline lands will be more than urban areas and infrastructures.

**Key words:** Agriculture; Land conversion; Land use change; Climate change; Iran.

### Introduction

Agricultural land conversion (ALC), as a types of land use/cover change (LUCC), is understood as one of the most important factors that can affect and be affected by climate change (Biro et al., 2013; Debolini et al., 2013; Lambin et al., 2000; Liu et al., 2012; Mahmoudi et al., 2010; Miyake et al., 2012; Mondal and Southworth, 2010; Salvati and Carlucci, 2010; Turner II, 2002, 2009; Vitousek, 1994; X Y et al., 2012). Climate change has recently been one of the most complex challenges that has affected all the countries in the world (The World Bank, 2010). Among others factors, such as emissions of CO<sub>2</sub> (Muñoz-Rojas et al., 2012) and alterations in the global nitrogen cycle (Wu et al., 2013; Yu et al., 2012), LUCC is understood as one of the most important. LUCC and specially ALC are recognized as an unavoidable phenomenon during economic development and population growth (Haque et al., 2008; Tan et al., 2009; J. J. Zhang et al., 2013b). ALC, among various types of LUCC, is the most important one in many countries in which agriculture is the major source of income. Since the world population expected to rise to about 9 billion by 2050 (FAO, 2011), demand for food and infrastructures will further increase (Dyson, 1999; Ewert et al., 2005; Johnson, 1999; Rosegrant et al., 2001) and the ALC trend will be further intensified. For example, in Mato Grosso area in Brazil the net cropped area expanded by 43% during the 2001 to 2007 (Arvor et al., 2012). But in China, since 1980, the conversion of agricultural land to non-agriculture has been widespread and intense (Ho and Lin, 2004; Su et al., 2011) and cultivated land was significantly reduced (J. Wang et al., 2012). In some European regions, urban sprawl has also affected vast agricultural areas in the past few years (Ceccarelli et al., 2014; European Environment Agency, 2006; Mazzocchi et al., 2013).

---

\* Correspondence to: A. A. Barati, Department of Agricultural Development and Management, University of Tehran, Iran. Phone: +98 26 32206825, Fax: +98 26 32206824, Email: [baratialiakbar@yahoo.com](mailto:baratialiakbar@yahoo.com)

On the other, over half of the population in the developing world, that includes 3.1 billion people (45% of all humanity), live in rural areas. Roughly, 2.5 billion of them make their livelihoods from agriculture (Hosseini Azadi and Barati, 2013). For these people land is a means to secure their food, livelihood and social status (Caldeira, 2008). That is why for many economies, especially developing countries such as Iran, monitoring ALC is of primary importance (Hosseini Azadi and Barati, 2013). In Iran, agricultural lands have more rapidly changed over the past 50 years than any time before and are expected to accelerate in the future (Bahrami et al., 2010). According to Iran's Statistical Center, agriculture is one of the most important sectors of the country's economy that currently contributes to 10% of the country's GDP and 18.2% of the total employment and agricultural products form about 30% of the country's non-oil exports (Hosseini Azadi and Barati, 2013). Nevertheless, the pace of ALC has been intensified in Iran. Recently, the lands are more fragmented and this process has aggravated the ALC (H. Azadi et al., 2011; Nejadi et al., 2012). For example, in Norway smaller farmers have encouraged many to leave agriculture (Forbord et al., 2014). According to the FAO Statistical Yearbook (FAO, 2012), Iran has one of the highest rates of the ALC. The arable land per capita during 1970-2009 has decreased 2.1% in the country compared to -1.46% as for the global rate. Iran's Agricultural Land Organization has reported that between 2005 and 2010, more than 74,755 ha of agricultural lands have changed to non-agriculture. Although other sources (such as the Agriculture Bank of Iran) have reported these changes up to 200,000 ha. Accordingly, despite the important role of the agricultural sector in the country's economy, this sector has considerably been threatened by ALC. Therefore, having a clear understanding of ALC and its drivers is vital for Iran's economy and sustainable agricultural land use management. Since monitoring ALC remains imperative and is considered as an essential first step to identify the main driving forces of ALC (Burgi et al., 2004). Using remote sensing (RS) and Geographic Information Systems (GIS) data, this paper aims to study the trend of ALC from 1990 to 2010 in Northwest Iran.

## Material and methods

The study was conducted in the Qazvin province located in Northwest Iran between the north latitude of 35°23'- 36°49' and east longitude of 48°44'-50°53' that covers 15,636 km<sup>2</sup> (Figure. 1). The province is divided to five counties, 19 localities (Bakhsh) and 46 rural districts. In 2012, the province held a population of 1.2 million, of which 72% lived in urban and 27% in rural areas. Although this province covers only about 1% of the total area of Iran, its contribution to the country's economy reaches up to 5% and more than 3% of Iran's agricultural products (from 2% to 20% of main products) are produced in this province that shows the importance of the province economy in the country.

Landuse patterns for July 1990 and 2010 were mapped by Landsat TM images. At first, each Landsat image was georeferenced to the local coordinate system based on 1:50,000 scale topographic maps. Then the geometric, radiometric and atmospheric corrections were used. Finally, each image classified using MAXLIKE module, which is one of the supervised classification methods used to generate a map in which each pixel assigned to a class based on its multispectral composition. The classes are determined based on the spectral composition of training areas defined by the user. All these processes were carried out in the five steps as follows. Through a fieldwork nine different types of landuses were determined. Then, the areas which were later used as training sites for each landuse or cover class were defined. After digitizing the training sites (training samples), the statistical characterization of each informational class was created (signatures) (Eastman, 2003). Then, by comparing different signatures, the best bands combination to separate different landuses and covers was defined. For this case, the best composites of the bands were 3, 5, and 7. In the third step, the images were classified by the MAXLIKE procedure. The total area was divided into nine classes, including rangeland (RL), dry land (DL), irrigated land (IL), orchard land (OL), saline land (SL), forest (F), water body (WB), urban areas and infrastructures (UI) and other land covers (OC).

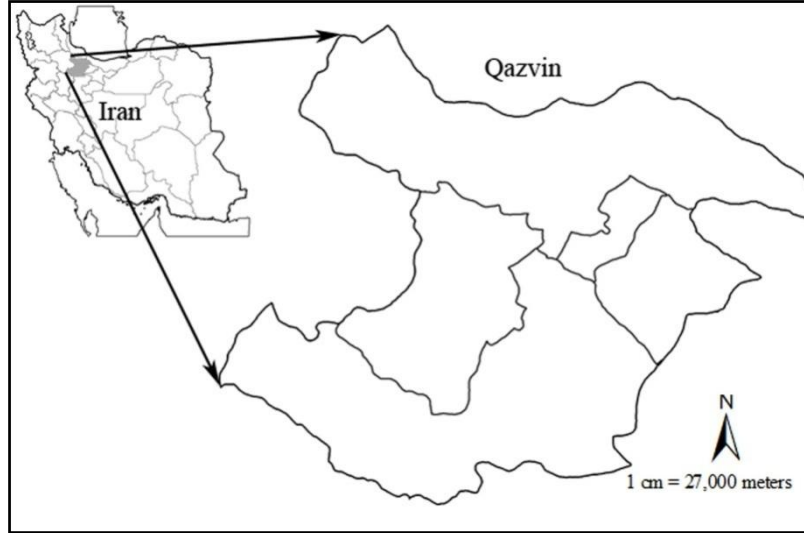


Figure. 1. Location of the Qazvin province (the study area) in Iran.

After classification, there may be many cases of isolated pixels that belong to a class which differs from the majority that surrounded them. This may be an accurate description of reality, but for mapping purposes, a very common post-processing operation is to generalize the image and remove these isolated pixels. In this study, the generalization is done by passing a median filter over the result (using the FILTER module in IDRISI). The median filter replaces each isolated pixel with the most frequent occurring class within a 3x3 window around each pixel. This effectively removes class patches of one or a few pixels and replaces them by the most common neighboring class. The final stage of the classification process usually involves an accuracy assessment that gives insight into ‘how good the classified image’ is. In fact, any map should be accompanied by an indication of the accuracy. Accuracy assessments determine the quality of the information derived from remotely sensed data. One of the main important indexes in this field is the Kappa index. For this study, the indexes, which derived from 70 actual field control points, were 0.752 for year 1990 and 0.793 for 2010. All the above procedures were performed by the IDRISI Kilimanjaro (Eastman, 2003) and ERDAS (ERDAS, 2010). Figure. 2 shows the landuse maps of the study area for the years 1990 (A) and 2010 (B).

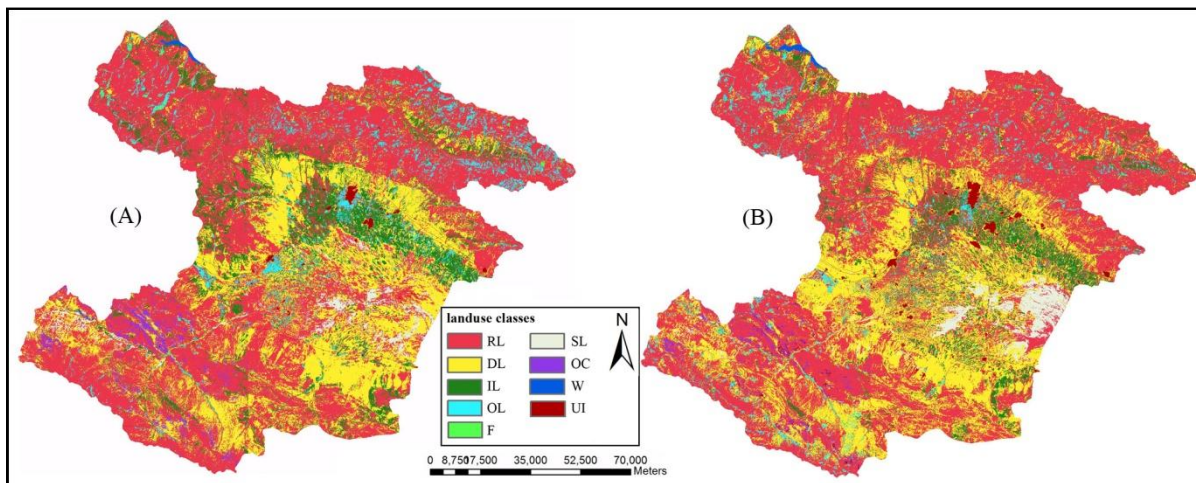


Figure. 2. The land use maps of study area: (A) 1990 and (B) 2010.

RL (Range Land), DL (Dry Land), IL (Irrigated Land), OL (Orchard Land), SL (Saline Land), F (Forest), WB (Water Body), UI (Urban and Other Infrastructures) and OC (Other Land Covers)

The monitoring was performed by computing the transition areas (Tables 1 and 2) and probability matrix (Table 3) using the Markov and CrossTab modules in the IDRISI Kilimanjaro software. A Markovian process is the one in which the state of a system at the time 2 (2010) can be predicted by the state of the system at the time 1 (1990). The transition matrix records the probability that each land use category or class would change to every other category, while the transition areas matrix records the number of pixels or the areas that are expected to change from each land use type to others over the specified number of time intervals. Each row of the matrix represents the probabilities for the various kinds of landuse classes.  $P_i$  for each row at the time 1 was estimated by Eq. 1:

$$p_i = \frac{x_i}{\sum x_i} \quad (\text{Eq. 1})$$

Where:

$P_i$  is the probability of each kind of landuse class, and  
 $x_i$  is the number of pixels in cell for each row.

Furthermore, this matrix can be used for determining the probability of each land use changes over a specified time period in the future. In general, for example, the probability of changing the RL in 1990 ( $i$ ) to the DL in 2030 ( $j$ ), which is denoted by  $p_{12}^{(2)}$ , can be computed by (Eq. 2) as follows:

$$p_{12}^{(2)} = \sum p_{11}p_{12} + p_{12}p_{22} + p_{13}p_{32} + p_{14}p_{42} + p_{15}p_{52} + p_{16}p_{62} + p_{17}p_{72} + p_{18}p_{82} + p_{19}p_{92} \quad (\text{Eq. 2})$$

Generally, if a Markovchain has  $r$  states, then:

$$p_{ij}^{(n)} = \sum_{k=1}^r p_{ik}p_{kj} \quad (\text{Eq. 3})$$

In Eq. 3,  $p$  is the value of each cell within the transition matrix of a Markov chain. The  $p_{ij}^{(n)}$  gives the probability that the Markov chain, starting in the state  $i$ , will be in the state  $j$  in  $n$  steps (Ross, 1997).

## Results and Discussion

Figure 2 and Tables 1 and 2 show that in 1990, RL, DL, IL, and OL were dominant in the study area. They were respectively covering 55.22, 25.31, 10.71, and 4.59% of the total area of the study site that altogether correspond to 95.8% (about 1,498,000 ha) of the area which has been covered by agricultural lands. In 2010, the agriculture areas decreased to 94.7%, which indicates that during this period, 44,845 ha (3%) of these lands have changed to non-agricultural lands. Also, in 2010, the shares of different agricultural landuse classes have been as follows: RL (55.04%); DL (28.6%); IL (7.33%) and OL (3.74%). As shown in the tables, the percent of RL, IL and OL areas have decreased -0.18, -3.37, and -0.84%, respectively. Conversely, the share of the DL areas increased by 3.28%, from 1990 to 2010. However, among the RL, IL and OL areas, the share of the IL areas decreased more intensely (-3.37%) than the others.

According to the tables, by 2010, up to 3% (44,845 ha) of agricultural lands (including 2.3% of RL, 5.8% of DL, 0.8% of IL and 0.5% of OL) were converted to non-agricultural lands. Most of these lands (71%) changed to SL and somewhat (22.8%) to UI. It means that during this period, 32,033 ha of agricultural lands (14,075 of RL, 17,783 of DL, 149 of IL and 26 ha of OL) were converted to SL whereas 10,243 ha (4,321 of RL, 4,503 of DL, 1,050 of IL and 369 ha of OL) were changed to UI. As shown in the tables, the RL and DL have mainly converted to SL. In contrast, most of the IL and OL were converted to UI.

Furthermore, for the period of 1990 to 2010, the saline areas expanded from 1.84 to 2.86%. It means that 15,846 ha of the lands were converted to saline lands.

Table 1

The surface of different land use classes and its changes between 1990 and 2010 (transition areas matrix).

a. RL (Range Lands), DL (Dry Lands), IL (Irrigated Lands), OL (Orchard Lands), SL (Salin Lands), F (Forest), WB (Water Bodies), UI (Urban and Other Infrastructures) and OC (Other Land Covers)

|                                     |    | Landuse 2010 (hectare) <sup>a</sup> |        |        |       |      |       |       |      |       | Total   |
|-------------------------------------|----|-------------------------------------|--------|--------|-------|------|-------|-------|------|-------|---------|
|                                     |    | RL                                  | DL     | IL     | OL    | F    | SL    | OC    | WB   | UI    |         |
| Landuse 1990 (hectare) <sup>a</sup> | RL | 635205                              | 146871 | 36587  | 24543 | 56   | 14075 | 1694  | 57   | 4321  | 863408  |
|                                     | DL | 99968                               | 243730 | 24540  | 4621  | 49   | 17783 | 619   | 4    | 4503  | 395817  |
|                                     | IL | 76009                               | 40832  | 42019  | 6959  | 196  | 149   | 181   | 9    | 1050  | 167404  |
|                                     | OL | 32604                               | 7305   | 10540  | 20369 | 474  | 26    | 5     | 1    | 369   | 71693   |
|                                     | F  | 757                                 | 90     | 182    | 1730  | 675  | 1     | 0     | 0    | 5     | 3440    |
|                                     | SL | 7827                                | 7411   | 719    | 163   | 1    | 11885 | 227   | 0    | 611   | 28844   |
|                                     | OC | 8102                                | 601    | 39     | 88    | 0    | 130   | 9040  | 0    | 2132  | 20130   |
|                                     | WB | 22                                  | 11     | 11     | 43    | 0    | 0     | 0     | 2743 | 0     | 2831    |
|                                     | UI | 141                                 | 276    | 54     | 23    | 0    | 642   | 12    | 0    | 8902  | 10052   |
| Total                               |    | 860636                              | 447128 | 114690 | 58539 | 1451 | 44691 | 11778 | 2814 | 21894 | 1563620 |

Table 2

Share of different land use classes and its changes between 1990 and 2010 (transition areas matrix).

a. RL (Range Lands), DL (Dry Lands), IL (Irrigated Lands), OL (Orchard Lands), SL (Saline Lands), F (Forest), WB (Water Bodies), UI (Urban and Other Infrastructures) and OC (Other Land Covers)

|                               |    | Landuse 2010 (%) <sup>a</sup> |       |       |       |       |       |       |       |       | Total |
|-------------------------------|----|-------------------------------|-------|-------|-------|-------|-------|-------|-------|-------|-------|
|                               |    | RL                            | DL    | IL    | OL    | F     | SL    | OC    | WB    | UI    |       |
| Landuse 1990 (%) <sup>a</sup> | RL | 73.57                         | 17.01 | 4.24  | 2.84  | 0.01  | 1.63  | 0.20  | 0.01  | 0.50  | 55.22 |
|                               | DL | 25.26                         | 61.58 | 6.20  | 1.17  | 0.01  | 4.49  | 0.16  | 0.00  | 1.14  | 25.31 |
|                               | IL | 45.40                         | 24.39 | 25.10 | 4.16  | 0.12  | 0.09  | 0.11  | 0.01  | 0.63  | 10.71 |
|                               | OL | 45.48                         | 10.19 | 14.70 | 28.41 | 0.66  | 0.04  | 0.01  | 0.00  | 0.51  | 4.59  |
|                               | F  | 21.99                         | 2.63  | 5.30  | 50.30 | 19.62 | 0.02  | 0.01  | 0.00  | 0.14  | 0.22  |
|                               | SL | 27.14                         | 25.69 | 2.49  | 0.56  | 0.00  | 41.20 | 0.79  | 0.00  | 2.12  | 1.84  |
|                               | OC | 40.25                         | 2.98  | 0.19  | 0.44  | 0.00  | 0.65  | 44.91 | 0.00  | 10.59 | 1.29  |
|                               | WB | 0.79                          | 0.39  | 0.39  | 1.52  | 0.00  | 0.00  | 0.00  | 96.91 | 0.00  | 0.18  |
|                               | UI | 1.41                          | 2.75  | 0.54  | 0.23  | 0.00  | 6.38  | 0.12  | 0.00  | 88.56 | 0.64  |
| Total                         |    | 55.04                         | 28.60 | 7.33  | 3.74  | 0.09  | 2.86  | 0.75  | 0.18  | 1.40  | 100   |

Table 3 indicates the probability of future changes for RL and DL to UI. As shown in the table and Figure. 3, the probability of the conversion to UI is higher than to SL while this probability for changing from IL and OL to SL is more than converting to UI.

Table 3

The transition probability matrix of different land use classes to the other classes between 2010 and 2030.

a. RL (Range Lands), DL (Dry Lands), IL (Irrigated Lands), OL (Orchard Lands), SL (Saline Lands), F (Forest), WB (Water Bodies), UI (Urban and Other Infrastructures) and OC (Other Land Covers)

|   |    | Landuse 2030 (Probability) <sup>a</sup> |       |       |       |       |       |       |       |       | Total |
|---|----|---|-------|-------|-------|-------|-------|-------|-------|-------|-------|
|   |    | RL                                      | DL    | IL    | OL    | F     | SL    | OC    | WB    | UI    |       |
| Landuse 2010 (Probability) <sup>a</sup> | RL | 0.601                                   | 0.228 | 0.078 | 0.056 | 0.000 | 0.027 | 0.004 | 0.001 | 0.004 | 1     |
|   | DL | 0.332                                   | 0.490 | 0.083 | 0.022 | 0.000 | 0.059 | 0.003 | 0.000 | 0.012 | 1     |
|   | IL | 0.507                                   | 0.226 | 0.201 | 0.053 | 0.002 | 0.001 | 0.001 | 0.002 | 0.006 | 1     |
|   | OL | 0.500                                   | 0.112 | 0.147 | 0.226 | 0.009 | 0.001 | 0.000 | 0.001 | 0.005 | 1     |
|   | F  | 0.229                                   | 0.027 | 0.072 | 0.495 | 0.175 | 0.000 | 0.000 | 0.000 | 0.001 | 1     |
|   | SL | 0.334                                   | 0.261 | 0.029 | 0.010 | 0.000 | 0.345 | 0.015 | 0.000 | 0.007 | 1     |
|   | OC | 0.486                                   | 0.041 | 0.002 | 0.006 | 0.000 | 0.010 | 0.457 | 0.000 | 0.000 | 1     |
|   | WB | 0.000                                   | 0.013 | 0.088 | 0.050 | 0.000 | 0.000 | 0.000 | 0.849 | 0.000 | 1     |
|   | UI | 0.073                                   | 0.105 | 0.012 | 0.003 | 0.000 | 0.002 | 0.000 | 0.000 | 0.805 | 1     |

### Rangelands conversion

Table 2 indicates that 73.6% of rangelands have remained unchanged during 1990-2010. About 24% of the converted RL have changed to the other agricultural landuse classes (more to DL, and less to IL and OL). This finding is consistent with studies in the Lahn-Dill Highlands (Germany) and in the Mashonaland central province (Zimbabwe) where Hietel et al. (2007) and Kamusoko et al. (2009) respectively found that land-cover changes occurred mainly between arable and grassland. The rest of the RL were mainly converted to saline lands (1.63%) or other land uses (0.2%). F. Zhang et al. (2013a) also reported the same changes in China. They showed that salinisation of grasslands has been much faster than cultivated land and more than 75% of all newly salinised lands are grasslands.

According to Table 3 and Figure. 3(a), the probability of future RL change (for 2030) to DL is more than to the other landuses and covers (0.228) and the probability of changing the OL and IL to RL is more than the others (respectively, 0.5 and 0.507). Among non-agricultural landuses, the probability of changing the RL to SL remains the most (0.027). On the other hand, the probability of changing these land uses to SL will increase from 0.016 (for 1990-2010) to 0.027 (for 1990-2030).

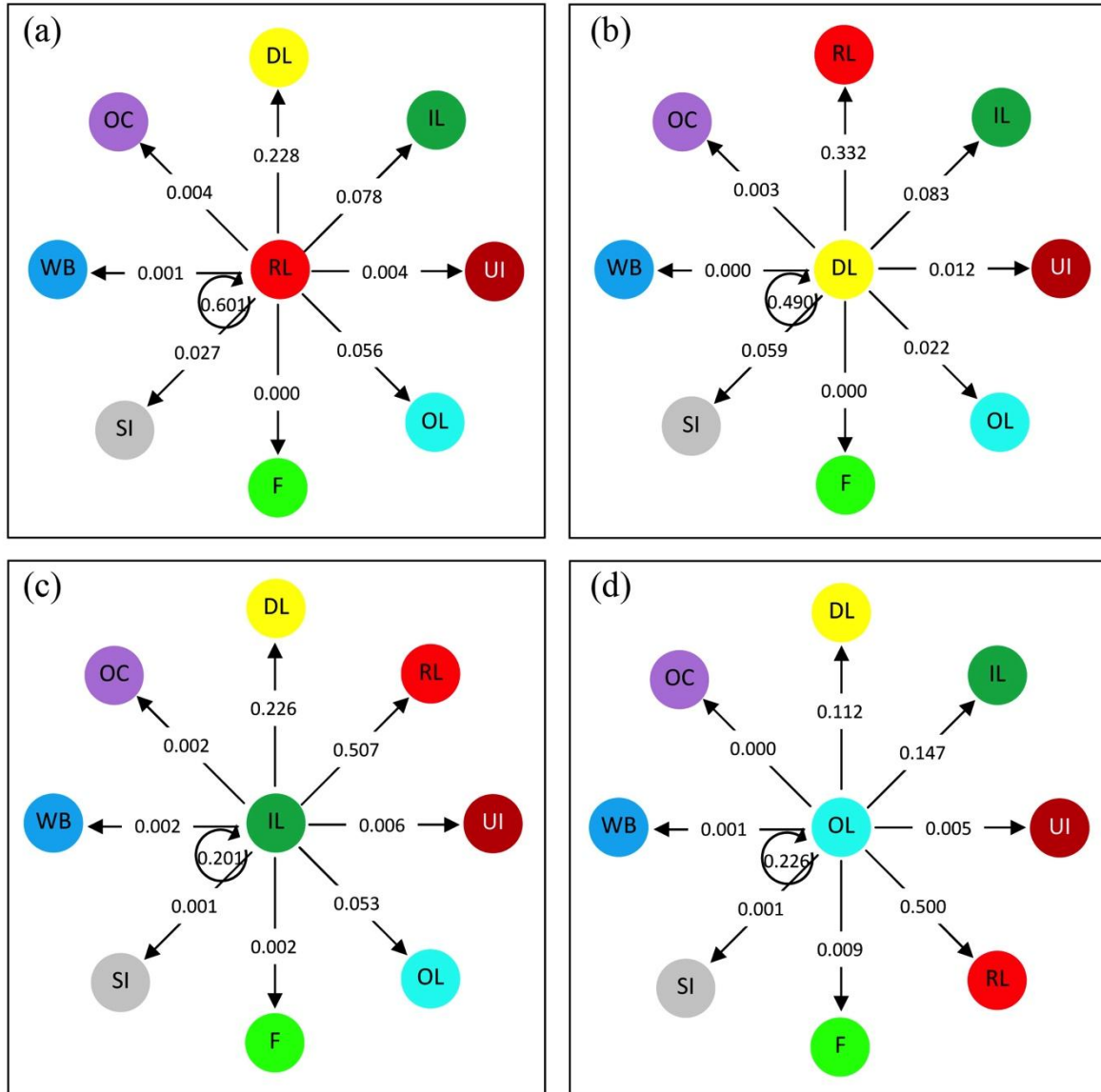


Figure. 3. Diagram of the future changing probability of agricultural lands between 2010-2030. RL (Range Lands), DL (Dry Lands), IL (Irrigated Lands), OL (Orchard Lands), SL (Saline Lands), F (Forest), WB (Water Bodies), UI (Urban and Other Infrastructures) and OC (Other Land Covers)

### Dry lands conversion

As shown in Table 2, until 2010, 38.4% of the DL have changed to other landuses and covers, in which 32.6% of these lands are converted to the other agricultural lands (25.3% to RL, 6.2% to IL and 1.2% to OL), and 6.2% of them have changed to non-agricultural lands, especially to SL (4.5%). During the same period, approximately 1% (1,584 ha) of the IL and 1.2% (875 ha) of the OL are converted to non-agricultural lands. The same changes among arable land, grassland and fallow land are reported by Hietel et al. (2007) and Kamusoko et al. (2009). Furthermore, Table 3 and Figure. 3(b) show that, by 2030, the probability of changing DL to SL is more than other non-agricultural landuses and covers (0.059). Among the agricultural lands, the probability of changing SL to DL is more than the others (0.228). Instead, among the non-agricultural lands, the probability of future changes of SL to DL remains the most (0.261).

### Irrigated lands conversion

According to Table 2, by 2010, about 75% of the IL is converted to the other landuses. The majority (74%) of these lands have changed to the other agricultural lands (45.4% to RL, 24.4% to DL and 4.2% to OL), and about 1% of them changed to non-agricultural lands, mostly to UIs (0.63%). Instead, from 1990 to 2010, around 4.2% of the RL and 6.2% of the DL and 14.7% of the OL were converted to the IL. This is in line with findings of Ho and Lin (2004) and Wang et al. (2012) who found that agricultural land during the past few decades has been converted to non-agriculture, respectively in China and Indonesia. Table 3 and Figure. 3(c) also show that in the future (by 2030), among the agricultural lands, the probability of the IL conversion to RLs is more than to the DL or OL (0.453). But for the non-agricultural lands, this probability, especially to UI, is more than the others (0.006). Also, like DL, the expansion probability of the SL within IL will increase by 2030 compared to 1990-2010. Although Ho and Lin (2004) indicated that the conversion of agricultural land to non-agricultural land has been widespread and intense in China, they did not specify the conversion types of agricultural to non-agricultural lands.

### **Orchard lands conversion**

Between 1990 to 2010, 71.5% of OL have changed (Table 2), of which, 70.4% occurred among the agricultural lands (45.5% to RL, 10.2% to DL and 14.7% to IL). The rest (1.1%) changed to non-agricultural lands; mainly to the UI. In contrast, by 2010, 2.8% of RL and 1.2% of DL and 4.2% of IL have changed to OL.

The probability of future OL changes to non-agricultural lands is 0.007 (Figure. 3(d) and Table3). Within the agricultural lands group, this probability for RL is higher than the others (0.5). This means that the chance of conversion for OL to RL is more than to DL or IL. However, among the non-agricultural lands, this chance for UI is more (0.005) than the other non-agricultural lands (i.e., SL, OC and WB). According to Azadi et al. (2011), Fukamachi et al. (2001) in Japan, Hitel et al. (2007) in Germany and Lichtenberg and Ding (Lichtenberg and Ding, 2008) in China, the development of infrastructures (such as road construction), also contributes significantly to ALC.

### **Grabbing agricultural lands by urban areas and infrastructures (UIs)**

Tables 1 and 2 indicate that during the 20 years (1990-2010), UI areas increased from 0.64% (10,052 ha) to 1.4% (21,894 ha) in the study site. This means that UI areas increased to 11,842 ha, of which, 10,243 ha were the result of ALC. These lands include 2.78% of total agricultural lands, of which, 1.77% (5,553 ha) were farm lands (IL and DL). As other researchers (H. Azadi et al., 2011; Fukamachi et al., 2001; Ho and Lin, 2004; Lichtenberg and Ding, 2008; Schulz et al., 2010; Tan et al., 2009) have reported, it seems that UI has expanded within the agricultural lands significantly. And as mentioned by Shrestha et al. (2012) in Southwest America, Su et al. (2011) in Hang-Jia-Hu region of China and EEA (2006) in Europe, rapid urbanization leads to agricultural land fragmentation. Furthermore, based on our study and the probability of the future UI conversion (Table 3 and Figure. 2), the urbanization trend will further be intensified by 2030 (from 0.098 to 0.127).

### **Conclusions**

This paper monitored the agricultural land conversion during the past two decades (from 1990 to 2010) using RS data and their probability of future change (by 2030). As showed by Mazzocchi et al. (2013), FAO (2012), Behnassi and Yaya (2011), IFAD (2010) and Azadi et al. (2011), our study also revealed that the surface of agricultural lands has decreased significantly. More specifically, the share of the IL areas decreased more intensely (-3.37%) than the other agricultural land uses, and this decrease will be more intensified by 2030. However, based on our study, the future probability of conversion for IL and OL is more than RL and DL. Since agriculture has an important role in food security and development (Vermeulen et al., 2012; World Bank, 2008) and access to adequate food is one of the most important aspects of food security, this decrease in IL and OL can be a serious threat to food security.

In line with Leh et al. (2013), Azadi et al. (2011; 2013), Shrestha et al. (2012) and EEA (2006), this study also showed that urban sprawl is one of the main threatening drivers of the ALC. However, the threat of UI for IL and OL has been realized much more. Since the role of the IL and OL in rural and agricultural



economy is vital, and because about one third of the population of the study area live in rural areas, conversion and degradation of these lands can be a serious threat for rural development (Kamusoko et al., 2009).

As discussed by Zhang et al. (2013a; 2007) and Xiujun (2000) in China, or Mahmoudi et al. (2010) in Karkhe sub-regions, our study also showed that the saline land expansion (salinization) has been the main threat for DLs and RLs, and this expansion will be continuing by 2030. Given that the salinization is a major process of land degradation and it is also considered to be one of the main drivers of the loss of agricultural land and crop yield (Thomas and Middleton, 1993), the saline expansion should be cautioned as a main threat for food security in agri-rural development. In addition, increasing the dust emission from the saline land (Chen et al., 2009) could worsen this situation.

Given the fact that most of the ALC has occurred within different types of agricultural lands, as a conclusion, we recommend monitoring not only the conversion of agricultural land to non-agricultural land, but also monitoring the conversion of different types of agricultural lands from one to another. The study of Williams and Schirmer (2012) in south-eastern Australia also showed that the most widespread of the land use changes was growth in farm lands.

## References

- Arvor, Damien, Meirelles, Margareth, Dubreuil, Vincent, Bégué, Agnès, & Shimabukuro, Yosio E. (2012). Analyzing the agricultural transition in Mato Grosso, Brazil, using satellite-derived indices. *Applied Geography*, 32(2), 702-713. doi: <http://dx.doi.org/10.1016/j.apgeog.2011.08.007>
- Azadi, H., Ho, P., & Hasfiati, L. (2011). Agricultural land conversion drivers: A comparison between less developed, developing and developed countries. *Land Degradation & Development*, 22(6), 596-604. doi: 10.1002/ldr.1037
- Azadi, Hossein, & Barati, Ali Akbar. (2013). Agricultural Land Conversion Drivers in Northeast Iran LDPI Working Papers. United Kingdom: The Land Deal Politics Initiative, Working Paper 36.
- Bahrami, Amir, Emadodin, Iraj, Ranjbar Atashi, Maryam, & Rudolf Bork, Hans (2010). Land-use change and soil degradation: A case study, North of Iran. *Agriculture and Biology Journal of North America*, 1(4), 600-605.
- Behnassi, M., & Yaya, Sanni. (2011). Land Resource Governance from a Sustainability and Rural Development Perspective. In M. Behnassi, S. A. Shahid & J. D'Silva (Eds.), *Sustainable Agricultural Development: Recent Approaches in Resources Management and Environmentally-Balanced Production Enhancement* (pp. 3-24). London: Springer.
- Biro, K., Pradhan, B., Buchroithner, M., & Makeschin, F. (2013). Land use/land cover change analysis and its impact on soil properties in the northern part of Gadarif region, Sudan. *Land Degradation & Development*, 24(1), 90-102. doi: 10.1002/ldr.1116
- Burgi, M., Hersperger, A. M., & Schneeberger, N. . (2004). Driving forces of landscape change current and new directions. *Landscape Ecology*, 19, 857-868.
- Caldeira, Rute. (2008). 'My land, your social transformation': Conflicts within the landless people movement (MST), Rio de Janeiro, Brazil. *Journal of Rural Studies*, 24(2), 150-160. doi: <http://dx.doi.org/10.1016/j.jrurstud.2007.12.001>
- Ceccarelli, T, Bajocco, S, Perini, L Luigi, & Salvati, L Luca. (2014). Urbanisation and Land Take of High Quality Agricultural Soils-Exploring Long-term Land Use Changes and Land Capability in Northern Italy. *International Journal of Environmental Research*, 8(1).
- Chen, Bing, Kitagawa, Hiroyuki, Hu, Ke, Jie, Dongmei, Yang, Junpeng, & Li, Jingmin. (2009). Element and mineral characterization of dust emission from the saline land at Songnen Plain, Northeastern China. *Journal of Environmental Sciences*, 21(10), 1363-1370. doi: [http://dx.doi.org/10.1016/S1001-0742\(08\)62427-4](http://dx.doi.org/10.1016/S1001-0742(08)62427-4)
- Debolini, Marta, Schoorl, Jeroen M., Temme, Arnaud, Galli, Mariassunta, & Bonari, Enrico. (2013). Changes in agricultural land use affecting future soil redistribution patterns: a case study in southern Tuscany (Italy) *Land Degradation & Development*, n/a-n/a. doi: 10.1002/ldr.2217

- Dyson, T. (1999). World food trends and prospects to 2025. *PNAS*, 96, 5929–5936.
- Eastman, J. Ronald (2003). *IDRISI Kilimanjaro Guide to GIS and Image Processing*. USA: Clark Labs.
- ERDAS. (2010). *ERDAS Field Guide*. USA: ERDAS, Inc.
- European Environment Agency. (2006). *Urban Sprawl in Europe, the Ignored Challenge* (pp. 56). Luxembourg: Office for Official Publications of the European Communities.
- Ewert, F., Rounsevell, M. D. A., Reginster, I., Metzger, M. J., & Leemans, R. (2005). Future scenarios of European agricultural land use: I. Estimating changes in crop productivity. *Agriculture, Ecosystems & Environment*, 107(2–3), 101–116. doi: <http://dx.doi.org/10.1016/j.agee.2004.12.003>
- FAO. (2011). *The State of the World's Land and Water Resources for Food and Agriculture* (FAO Ed.). Rome: FAO.
- FAO. (2012). *FAO Statistical Yearbook 2012*. Rome: FAO.
- Forbord, Magnar, Bjørkhaug, Hilde, & Burton, Rob J. F. (2014). Drivers of change in Norwegian agricultural land control and the emergence of rental farming. *Journal of Rural Studies*, 33(0), 9–19. doi: <http://dx.doi.org/10.1016/j.jrurstud.2013.10.009>
- Fukamachi, Katsue, Oku, Hirokazu, & Nakashizuka, Tohru. (2001). The change of a satoyama landscape and its causality in Kamiseya, Kyoto Prefecture, Japan between 1970 and 1995. *Landscape ecology*, 16(8), 703–717.
- Haque, A, Alam, JB, Shaha, NK, & Raihan, F. (2008). Study on Land use Pattern Change and Its Causes. *International Journal of Environmental Research*, 2(2).
- Hietel, Elke, Waldhardt, Rainer, & Otte, Annette. (2007). Statistical modeling of land-cover changes based on key socio-economic indicators. *Ecological Economics*, 62(3–4), 496–507. doi: 10.1016/j.ecolecon.2006.07.011
- Ho, Samuel PS, & Lin, George. (2004). Converting Land to Nonagricultural Use in China's Coastal Provinces Evidence from Jiangsu. *Modern China*, 30(1), 81–112.
- IFAD. (2010). *Land tenure security and poverty reduction*. International Fund for Agriculture and Development. Rome: IFAD.
- Johnson, D.G. (1999). The growth of demand will limit output growth for food over the next quarter century. *PNAS*, 96, 5915–5920.
- Kamusoko, Courage, Aniya, Masamu, Adi, Bongo, & Manjoro, Munyaradzi. (2009). Rural sustainability under threat in Zimbabwe – Simulation of future land use/cover changes in the Bindura district based on the Markov-cellular automata model. *Applied Geography*, 29(3), 435–447. doi: 10.1016/j.apgeog.2008.10.002
- Lambin, E. F., Rounsevell, M. D. A., & Geist, H. J. (2000). Are agricultural land-use models able to predict changes in land-use intensity? *Agriculture, Ecosystems & Environment*, 82(1–3), 321–331. doi: 10.1016/S0167-8809(00)00235-8
- Lichtenberg, Erik, & Ding, Chengri. (2008). Assessing farmland protection policy in China. *Land Use Policy*, 25(1), 59–68.
- Liu, Z., Yao, Z., Huang, H., Wu, S., & Liu, G. (2012). Land use and climate changes and their impacts on runoff in the Yarlung Zangbo river basin, China. *Land Degradation & Development*, n/a–n/a. doi: 10.1002/ldr.1159
- Mahmoudi, B, Bakhtiari, F, Hamidifar, M, & Danehkar, A. (2010). Effects of Land use Change and Erosion on Physical and Chemical Properties of Water (Karkhe watershed). *International Journal of Environmental Research*, 4(2).
- Mazzocchi, Chiara, Sali, Guido, & Corsi, Stefano. (2013). Land use conversion in metropolitan areas and the permanence of agriculture: Sensitivity Index of Agricultural Land (SIAL), a tool for territorial analysis. *Land Use Policy*, 35(0), 155–162. doi: <http://dx.doi.org/10.1016/j.landusepol.2013.05.019>
- Miyake, Saori, Renouf, Marguerite, Peterson, Ann, McAlpine, Clive, & Smith, Carl. (2012). Land-use and environmental pressures resulting from current and future bioenergy crop expansion: A

- review. *Journal of Rural Studies*, 28(4), 650-658. doi: <http://dx.doi.org/10.1016/j.jrurstud.2012.09.002>
- Mondal, Pinki, & Southworth, Jane. (2010). Evaluation of conservation interventions using a cellular automata-Markov model. *Forest Ecology and Management*, 260(10), 1716-1725. doi: 10.1016/j.foreco.2010.08.017
- Muñoz-Rojas, M., Jordán, A., Zavala, L. M., La Rosa, D. De, Abd-Elmabod, S. K., & Anaya-Romero, M. (2012). IMPACT OF LAND USE AND LAND COVER CHANGES ON ORGANIC CARBON STOCKS IN MEDITERRANEAN SOILS (1956–2007). *Land Degradation & Development*, n/a-n/a. doi: 10.1002/ldr.2194
- Nejadi, Athareh, Jafari, HR, Makhdoum, MF, & Mahmoudi, M. (2012). Modeling Plausible Impacts of land use change on wildlife habitats, Application and validation: Lissar protected area, Iran. *International Journal of Environmental Research*, 6(4).
- Rosegrant, M.W., Paisner, M.S., Meijer, S., & Witcover, J. (2001). *Global Food Projections to 2020. Emerging Trends and Alternative Future*. Washington, DC: International Food Policy Research Institute.
- Ross, Sheldon M. (1997). *Introduction to probability models (Vol. 6)*: Academic Press San Diego, California.
- Salvati, Luca, & Carlucci, Margherita. (2010). Estimating land degradation risk for agriculture in Italy using an indirect approach. *Ecological Economics*, 69(3), 511-518. doi: <http://dx.doi.org/10.1016/j.ecolecon.2009.08.025>
- Schulz, Jennifer J., Cayuela, Luis, Echeverria, Cristian, Salas, Javier, & Rey Benayas, José María. (2010). Monitoring land cover change of the dryland forest landscape of Central Chile (1975–2008). *Applied Geography*, 30(3), 436-447. doi: <http://dx.doi.org/10.1016/j.apgeog.2009.12.003>
- Shrestha, Milan K., York, Abigail M., Boone, Christopher G., & Zhang, Sainan. (2012). Land fragmentation due to rapid urbanization in the Phoenix Metropolitan Area: Analyzing the spatiotemporal patterns and drivers. *Applied Geography*, 32(2), 522-531. doi: <http://dx.doi.org/10.1016/j.apgeog.2011.04.004>
- Su, Shiliang, Jiang, Zhenlan, Zhang, Qi, & Zhang, Yuan. (2011). Transformation of agricultural landscapes under rapid urbanization: A threat to sustainability in Hang-Jia-Hu region, China. *Applied Geography*, 31(2), 439-449. doi: <http://dx.doi.org/10.1016/j.apgeog.2010.10.008>
- Tan, Rong, Beckmann, Volker, van den Berg, Leo, & Qu, Futian. (2009). Governing farmland conversion: Comparing China with the Netherlands and Germany. *Land Use Policy*, 26(4), 961-974.
- The World Bank. (2010). *World Development Report 2010: Development and Climate Change* (T. W. Bank Ed.). Washington DC: The International Bank for Reconstruction and Development / The World Bank
- Thomas, D. S. G., & Middleton, N. J. (1993). Salinization: new perspectives on a major desertification issue. *Journal of Arid Environments*, 24(1), 95-105. doi: <http://dx.doi.org/10.1006/jare.1993.1008>
- Turner II, Billy Lee. (2002). *Toward Integrated Land-Change Science: Advances in 1.5 Decades of Sustained International Research on Land-Use and Land-Cover Change*. Paper presented at the Challenges of a Changing Earth: Proceedings of the Global Change Open Science Conference, Amsterdam, NL, 10-13 July.
- Turner II, Billy Lee. (2009). *Land Change Science*. In K. Rob & T. Nigel (Eds.), *International Encyclopedia of Human Geography* (pp. 107-111). Oxford: Elsevier.
- Vermeulen, S. J., Aggarwal, P. K., Ainslie, A., Angelone, C., Campbell, B. M., Challinor, A. J., . . . Wollenberg, E. (2012). Options for support to agriculture and food security under climate change. *Environmental Science & Policy*, 15(1), 136-144. doi: <http://dx.doi.org/10.1016/j.envsci.2011.09.003>
- Vitousek, Peter M. (1994). Beyond Global Warming: Ecology and Global Change. *Ecology*, 75(7), 1861-1876. doi: 10.2307/1941591

- Wang, Chunmei, & Maclaren, Virginia. (2012). Evaluation of economic and social impacts of the sloping land conversion program: A case study in Dunhua County, China. *Forest Policy and Economics*, 14(1), 50-57. doi: 10.1016/j.forpol.2011.06.002
- Wang, Jing, Chen, Yongqi, Shao, Xiaomei, Zhang, Yanyu, & Cao, Yingui. (2012). Land-use changes and policy dimension driving forces in China: Present, trend and future. *Land Use Policy*, 29(4), 737-749. doi: <http://dx.doi.org/10.1016/j.landusepol.2011.11.010>
- Williams, Kathryn J. H., & Schirmer, Jacki. (2012). Understanding the relationship between social change and its impacts: The experience of rural land use change in south-eastern Australia. *Journal of Rural Studies*, 28(4), 538-548. doi: <http://dx.doi.org/10.1016/j.jrurstud.2012.05.002>
- World Bank. (2008). *World development report 2008: Agriculture for development*. Washington DC, USA: World Bank.
- Wu, J. P., Liu, Z. F., Sun, Y. X., Zhou, L. X., Lin, Y. B., & Fu, S. L. (2013). Introduced eucalyptus urophylla plantations change the composition of the soil microbial community in subtropical China. *Land Degradation & Development*, 24(4), 400-406. doi: 10.1002/ldr.2161
- X Y, Feng, G P, Luo, & C E, Li. (2012). Dynamics of Ecosystem Service Value Caused by Land use Changes in Manas River of Xinjiang, China. *International Journal of Environmental Research*, 6(2).
- Xiujun, Li. (2000). The Alkali-saline land and agricultural sustainable development of the Western Songnen plain in China. *Scientia Geographica Sinica*, 20(1), 51-55.
- Yu, B., Stott, P., Di, X. Y., & Yu, H. X. (2012). Assessment of land cover changes and their effect on soil organic carbon and soil total nitrogen in daqing prefecture, China. *Land Degradation & Development*, n/a-n/a. doi: 10.1002/ldr.2169
- Zhang, F., Tiyip, T., Feng, Z. D., Kung, H. T., Johnson, V. C., Ding, J. L., . . . Gui, D. W. (2013a). Spatio-temporal patterns of land use/cover changes over the past 20 years in the middle reaches of the Tarim river, Xinjiang, China. *Land Degradation & Development*, n/a-n/a. doi: 10.1002/ldr.2206
- Zhang, J. J., Fu, M. C., Zeng, H., Geng, Y. H., & Hassani, F. P. (2013b). Variations in ecosystem service values and local economy in response to land use: a case study of Wu'an, China. *Land Degradation & Development*, 24(3), 236-249. doi: 10.1002/ldr.1120