

Experimental accumulation and depuration kinetics and natural occurrence of microcystin-LR in basil (*Ocimum basilicum* L.)

Wannes Hugo R. Van Hassel, Mohamed F. Abdallah, Maria Gracia Guzman Velasquez, Christopher O. Miles, Ingunn A. Samdal, Julien Masquelier, Andreja Rajkovic

PII: S0269-7491(24)00429-9

DOI: https://doi.org/10.1016/j.envpol.2024.123715

Reference: ENPO 123715

To appear in: Environmental Pollution

Received Date: 13 November 2023

Revised Date: 29 February 2024

Accepted Date: 3 March 2024

Please cite this article as: Van Hassel, W.H.R., Abdallah, M.F., Gracia Guzman Velasquez, M., Miles, C.O., Samdal, I.A., Masquelier, J., Rajkovic, A., Experimental accumulation and depuration kinetics and natural occurrence of microcystin-LR in basil (*Ocimum basilicum* L.), *Environmental Pollution* (2024), doi: https://doi.org/10.1016/j.envpol.2024.123715.

This is a PDF file of an article that has undergone enhancements after acceptance, such as the addition of a cover page and metadata, and formatting for readability, but it is not yet the definitive version of record. This version will undergo additional copyediting, typesetting and review before it is published in its final form, but we are providing this version to give early visibility of the article. Please note that, during the production process, errors may be discovered which could affect the content, and all legal disclaimers that apply to the journal pertain.

© 2024 Published by Elsevier Ltd.

| 1 | Expe | rimental | accumulati | on a | nd | depuration | kir | netics |
|---|-------|----------|------------|------|------|-------------|-----|--------|
| 2 | and | natural | occurrence | of | micı | rocystin-LR | in | basil |
| 3 | (Ocir | num basi | ilicum L.) | | | | | |
| 4 | | | | | | | | |

| 5 6 | | | | | | | |
|----------|--|--|--|--|--|--|--|
| 7 | Wannes Hugo R. Van Hassel ^{1,2,3*} , Mohamed F. Abdallah ^{1,4} , Maria Gracia Guzman Velasquez ^{1,2} , | | | | | | |
| 8 | Christopher O. Miles ^{5,6} , Ingunn A. Samdal ⁶ , Julien Masquelier ² , Andreja Rajkovic ¹ | | | | | | |
| 9 | | | | | | | |
| 10 | | | | | | | |
| 11 | ¹ Department of Food Technology, Safety and Health, Faculty of Bioscience Engineering, Ghent | | | | | | |
| 12 | University, Gent 9000, Belgium. | | | | | | |
| 13 | ² Sciensano, Chemical and Physical Health Risks, Organic Contaminants and Additives, | | | | | | |
| 14 | Leuvensesteenweg 17, Tervuren 3080, Belgium. | | | | | | |
| 15 | ³ InBios- Centre for Protein Engineering, Department of Life Sciences, University of Liège, Allée du six | | | | | | |
| 16 | Août 11, Liège 4000, Belgium. | | | | | | |
| 17 | ⁴ Department of Forensic Medicine and Toxicology, Faculty of Veterinary Medicine, Assiut University, | | | | | | |
| 18 | Assiut 71515, Egypt. | | | | | | |
| 19 | ⁵ Biotoxin Metrology, National Research Council Canada, Halifax, NS B3H 3Z1, Canada. | | | | | | |
| 20 | ⁶ Norwegian Veterinary Institute, Postboks 64, 1431 Ås, Norway. | | | | | | |
| 21 | | | | | | | |
| 22 | | | | | | | |
| 23 | | | | | | | |
| 24 25 | | | | | | | |
| 25 | | | | | | | |
| 27 | | | | | | | |
| 28 | | | | | | | |
| 29 | | | | | | | |
| 30 | | | | | | | |
| 31 | | | | | | | |
| 32 | *Correspondence: <u>wannes.vanhassel@sciensano.be</u> Leuvensesteenweg 17, Tervuren 3080, Belgium. | | | | | | |

33 Abstract

34 Microcystin-LR (MC-LR) is a hepatotoxic metabolite that naturally occurs during some cyanobacterial blooms in eutrophic waterbodies, and irrigation of edible plants with MC-LR-contaminated water 35 36 causes bioaccumulation of the toxin. However, sufficient information about accumulation and 37 depuration mechanics in hydroculture-grown herb plants is still lacking. This work aimed at 1) 38 investigating bioaccumulation and depuration of MC-LR in basil, 2) verifying the possible MC-LR 39 detoxification mechanisms in the plant, and 3) detecting the natural occurrence of MC-LR in basil (n = 40 50) collected from the Belgian market. Basil plants grown in a hydroculture were exposed to MC-LR 41 (5, 20, and 50 μ g L⁻¹) spiked in a Hoagland solution for seven days. MC-LR depuration was also studied by transferring the plants to a non-contaminated Hoagland solution after exposure to MC-LR for 42 another seven days. MC-LR concentrations in Hoagland solution, basil leaves, and roots were 43 44 quantified using a validated UHPLC-MS/MS method. In addition, ELISA and LC-HRMS (only basil 45 leaves) were used for confirmation. The results showed an increase in the accumulated levels of MC-LR at higher exposure doses, with higher MC-LR levels in roots than in leaves for all the treatment 46 47 conditions. For MC-LR depuration, significant reductions were observed in all the treatment conditions for roots only. No MC-LR conjugates, potentially related to metabolism, were detected by LC–HRMS. 48 49 Finally, MC-LR was detected in one store-bought basil sample, representing the first occurrence of cyanotoxins in an edible crop from Belgium. 50

51 Graphical abstract



52

53 Keywords: Microcystins; MC-LR; basil; UHPLC-MS/MS; accumulation; depuration; food safety

54 **1. Introduction**

55 Cyanobacterial blooms are common in eutrophic waterbodies worldwide (Svirčev et al., 2019), 56 promoted by, but not limited to, increasing temperatures and other climate change phenomena 57 (O'Neil et al., 2012). During the formation of these blooms, the production and release of a wide array 58 of toxic metabolites into the water can happen (Abdallah et al., 2021; Ibelings et al., 2014). Irrigation 59 of food crops with cyanotoxin-contaminated water from the environment (e.g., basins, lakes, ponds, 60 or canals) may result in bioaccumulation and translocation of the cyanotoxins to edible parts (Xiang et al., 2019; Zhang et al., 2021). Besides their possible negative impacts on the plant by lowering yield 61 62 and quality (Peuthert et al., 2007), cyanotoxins pose a human health risk (Codd et al., 1999; 63 Tsoumalakou et al., 2021). The most common cyanotoxin is microcystin-LR (MC-LR), a hepatotoxic and possibly hepatocarcinogenic agent (group 2B) in humans (IARC, 2010; World Health Organization, 64 65 2020). Accumulation of MC-LR was primarily shown in soil-based systems for different edible crops 66 (Bittencourt-Oliveira et al., 2016; Chen et al., 2012; Codd et al., 1999; Corbel et al., 2016; Machado et 67 al., 2017). Moreover, MC accumulation has already been studied in some crop plants. Lettuce is the prime example, as MC accumulation (originating from the environment or fortified irrigation water) 68 was observed in roots and leaves (Bittencourt-Oliveira et al., 2016; Codd et al., 1999; Cordeiro-Araújo 69 70 et al., 2016; Crush et al., 2008; Hereman and Bittencourt-Oliveira, 2012; Levizou et al., 2017; 71 Mohamed and Al Shehri, 2009). Accumulation of cylindrospermopsin, another hepatotoxic 72 cyanobacteria metabolite, was also observed in lettuce (Cordeiro-Araújo et al., 2017; Llana-Ruiz-73 Cabello et al., 2019). Exposure-dependent accumulation of MCs was also established in rice (Oryza 74 sativa) and root vegetables such as carrots, radishes, and rape (Chen et al., 2012, 2004; Crush et al., 75 2008; Levizou et al., 2020; Machado et al., 2017; Wijewickrama and Manage, 2019). Other laboratory-76 based studies showed the accumulation of MCs in fruit-bearing plants, including chili, tomato, and 77 strawberries (Corbel et al., 2016; Redouane et al., 2023; Romero-Oliva et al., 2014). Yet, primary 78 accumulation in these products was observed in the roots, stems, and leaves. Most of the observed 79 accumulation of MCs in crops originated from soil-grown crops (Bittencourt-Oliveira et al., 2016; Chen 80 et al., 2004; Codd et al., 1999; Crush et al., 2008; Hereman and Bittencourt-Oliveira, 2012; Levizou et 81 al., 2017; Mohamed and Al Shehri, 2009). However, plants grown in hydroculture systems may be 82 especially vulnerable to MC-LR accumulation due to the direct contact between the contaminated 83 water and plant roots. Accumulation of MCs in hydroculture systems has already been documented 84 for several rice varieties (Wijewickrama and Manage, 2019) and leafy vegetables such as lettuce and 85 spinach (Llana-Ruiz-Cabello et al., 2019; Wijewickrama and Manage, 2019).

86 In Belgium, MC-LR contamination has been reported in freshwater reservoirs (Van Hassel et al., 2022a; 87 Willame et al., 2005). Recently, cyanobacterial blooms from multiple lakes, ponds, and canals were 88 sampled, and eight microcystins (MCs) were quantified in 86% (n = 79) of the samples, with concentrations ranging from 0.1 μ g L⁻¹ to 2800 μ g L⁻¹ (Van Hassel et al., 2022a). These water sources 89 90 can potentially be used for irrigation of crops. Therefore, potential MC-LR accumulation in Belgian 91 crops, including basil (Ocimum basilicum L.), is a valid concern that might contribute to a public health 92 risk. Basil is an important crop that is usually grown in hydroculture systems in several countries (Saha 93 et al., 2016). The plant is generally used as a flavoring agent and a main ingredient in Pesto Genovese, 94 a sauce used in many European dishes, especially in Italian cuisines. In addition to foods, basil is used in the cosmetic and pharmaceutical industries since it is a good source of essential oils, valuable 95 96 antioxidants, and other bioactive compounds (Trajkovska-Broach et al., 2023). In Europe, basil makes up 60–75% of the total consumed herbs, and Belgium is among the top four countries (after Germany, 97 98 the Netherlands, and France) that import fresh herbs (Centre for the Promotion of Imports from developing countries, 2020). Also, national producers contribute to the total marketed basil in 99 100 Belgium.

101 The objectives of this work were to investigate the accumulation and depuration of MC-LR in basil 102 plants grown in a hydroculture, verify the possible mechanisms involved in the depuration, and survey 103 the natural occurrence of MC-LR in basil samples collected from the Belgian market using an in-house 104 validated ultra-high performance liquid chromatography-tandem mass spectrometry (UHPLC– 105 MS/MS) method for the quantification of nine cyanotoxins. The toxins include eight MCs (MC-LR, MC-106 RR, MC-LA, MC-LF, MC-LY, MC-LW, MC-YR, and MC-WR) and a structurally related cyanotoxin, 107 nodularin-R (NOD-R).

108 **2. Materials and Methods**

109 2.1. Experimental setup

Certified basil seeds (Italian large green variety; La Belle Portagere brand, Sebio, Belgium) and potting 110 111 soil for herbs and aromatic plants (Campo brand, Brico, Belgium) were commercially obtained. After sowing, seeds germinated and grew inside a growth chamber (Mammoth Lite+ 45 type, 450 x 450 × 112 1200 mm, Mammoth net, Netherlands) where the light (around 200 μ mol m⁻² s⁻¹ for 16 hours/day), 113 temperature (24.6 \pm 1.6 °C), and relative humidity (57.8 \pm 12.0) were monitored and kept constant. 114 The plants were watered every other day and after 60 days, the plants were collected and transferred 115 116 to a Hoagland solution (Hoagland and Arnon, 1950). Each plant was placed in a 100 mL amber glass bottle containing Hoagland solution fortified with MC-LR and sealed with parafilm. Each treatment 117

group (n = 6) was exposed to 5, 20, or 50 µg L⁻¹ of MC-LR (Enzo Life Sciences, Belgium) for seven days, 118 119 in addition to an untreated control group. Fortified Hoagland was prepared before the incubation of the plants. A stock solution of 10 µg mL⁻¹ of MC-LR in ethanol was used to make a 1:200 dilution in 120 Hoagland solution to obtain the 50 ng mL⁻¹ concentration. After that, serial dilutions were used to 121 make 1 L of Hoagland solution at 50, 20, and 5 μ g L⁻¹ MC-LR. To study the depuration of MC-LR, the 122 123 same experimental setup was performed, parallel to the accumulation experiment, but the plants 124 were transferred into clean (i.e., non-spiked, MC-LR-free) Hoagland solution for another seven days 125 after exposure to MC-LR as described above. As a preliminary step to investigate any effect of MC-LR 126 on basil, the weight and length of each plant were measured before and after exposure to the toxin. Also, the plant leaves and roots were visually inspected during and after the experiment. To assess 127 128 the bioaccumulation and depuration of MC-LR, the toxin was quantified in basil leaves, roots, and the 129 Hoagland solution. The basil roots, obtained after the accumulation testing period, were dipped for 130 two seconds into distilled water and dried to ensure that no MC-LR was taken up from the Hoagland solution itself before processing, extraction, and UHPLC-MS/MS analysis. The experiment was 131 132 repeated twice. Additionally, the accumulation part of the experimental setup was tested before the 133 start of the experiment using an extract of *Microcystis* sp. ULC642 culture obtained from the 134 BCCM/ULC culture collection (Liège, Belgium). The cultures were grown in BG11 at 22 °C and 700 Lux 135 warm-led light intensity on a 10:14 day:night cycle. This extract contained a mixture of MCs, mainly 136 MC-LR and MC-YR (Table S1), and was used to fortify the Hoagland solution to contaminate six basil plants at 50, 20, and 5 μ g L⁻¹ total MCs. 137

138 2.2. Quantification of eight MCs congeners and NOD-R quantification in basil

139 Samples from Hoagland solution and basil (separated leaves and roots) were collected and kept at -20 140 °C until analysis. The extraction and UHPLC-MS/MS analysis methods for MC-LR from the different 141 matrices were previously published (Van Hassel et al., 2022b, 2022c). These methods are suitable for 142 analyzing eight MC congeners (MC-LR, MC-RR, MC-LA, MC-LF, MC-LY, MC-LW, MC-YR, MC-WR) and 143 NOD-R in fruit and vegetable matrices and drinking water. The validity of the methods for fruits and 144 vegetables was confirmed for the basil matrix, reassessing repeatability, reproducibility, recovery, specificity, linearity, limit of detection (LOD) and limit of quantification (LOQ), measurement 145 146 uncertainty (MU), and matrix effect. In brief, this method consisted of extracting the targeted toxins 147 from 0.5 g of homogenized basil material (leaves or roots) using a combination of liquid extraction (methanol-water 80:20), sonication (BRANSON 2510, Analis, Belgium), overhead mixing (Heidolph 148 149 Reax 2 Mixer, Analis, Belgium), and centrifugation (Sorvall Legend XT, Thermo Scientific, Belgium) at 150 15303 g for 15 min, after which the extracts were partially evaporated and further purified using

Agilent C18 cartridges (6 mL, 500 mg) as described by Van Hassel et al., 2022 (Van Hassel et al., 2022b, 151 152 2022c). Methanol was UPLC/MS grade (Biosolve B.V., Valkenswaard, The Netherlands). Milli-Q water 153 was produced in-house (conductivity \geq 18.2 M Ω and TOC \leq 4 ppb). After this purification step, the MCs 154 and NOD-R were separated and quantified via UHPLC-MS/MS (Llana-Ruiz-Cabello et al., 2019; Van Hassel et al., 2022c). Except for MC-RR, matrix-matched calibration curves in blank basil matrices were 155 156 prepared from 0.25 ng L⁻¹ to 50 ng L⁻¹. The latter was from 0.75 ng L⁻¹ to 50 ng L⁻¹ due to matrix interference (Table S2). The blank refers to a sample of the same food matrix free of MCs. The quality 157 control (QC) sample was a blank matrix spiked with 25 μ L of toxin standard solution (100 ng mL⁻¹) at 158 5 ng g^{-1} , giving a known toxin concentration. Both served to check the quality of the analysis. Diluted 159 160 toxin standards were prepared in MeOH and Milli-Q water (50:50) with 1% acetic acid. Blank and QC 161 samples were included in the batch to check the sensitivity of the instrument during the analysis. Validation of the method for the basil matrix followed the same methodology described in earlier 162 163 work for the fruit and vegetables (Van Hassel et al., 2022c). An MC and NOD mixture was spiked in homogenized basil leave matrix at 1, 5, and 25 ng g⁻¹ before extraction. However, the limit of 164 quantification (LOQ) was only 1 ng g⁻¹ for MC-LR and NOD-R. Therefore, additional concentration 165 levels were spiked at 5 ng g⁻¹ for MC-RR and 2.5 ng g⁻¹ for MC-YR, MC-LA, MC-LF, MC-WR, MC-LW, and 166 167 MC-LY. The LOQ was defined as the concentration for which the quantification of the compound 168 adhered to all validation parameters.

169 2.3. ELISA analysis

170 Enzyme-linked immunosorbent assay (ELISA) was used to confirm MC-LR accumulation and 171 depuration in basil leaf samples (Samdal et al., 2014). Minor adjustments to concentrations of ELISA assay reagents, such as 0.5 μ g mL⁻¹ of the plate-coating antigen, 1:4000 of antiserum 80289-5b, and 172 1:5500 of the donkey–antisheep IgG (H + L)–horseradish peroxidase conjugate (antisheep–HRP from 173 174 Agrisera antibodies (Vännäs, Sweden)) were made to optimize the ELISA. The MC-LR standard (NRC CRM-MC-LR (Lot#20070131)) in 50% MeOH (500 ng mL⁻¹) was diluted in phosphate-buffered saline 175 with Tween to give a MeOH concentration of 10% and then in a threefold dilution series in sample 176 177 buffer, resulting in a standard series of 50, 16.7, 5.56, 0.62, 0.20, 0.069, 0.023, 0.0076, and 0.0025 ng mL⁻¹. Standards and samples were analyzed with serial dilutions and in duplicate on the plate. All 178 179 incubations were performed at ~20 °C. Absorbances were measured at 450 nm using a SpectraMax i3x plate reader (Molecular Devices, Sunnyvale, CA, USA). Assay standard curves were fitted using a 4-180 181 parameter logistic treatment of the data using SoftMax Pro version 6.5.1. (Molecular Devices, 182 Sunnyvale, CA, USA). The assay working range was defined as the linear region at 20-80% of maximum 183 absorbance (A_{max}).

184 2.4. LC–HRMS analysis

185 LC-HRMS analysis was performed with a Q Exactive HF Orbitrap mass spectrometer equipped with a 186 heated electrospray ionization interface (ThermoFisher Scientific, Waltham, MA, USA) using an Agilent 187 1200 G1312B binary pump, G1367C autosampler and G1316B column oven (Agilent, Santa Clara, CA, 188 USA) connected to a Symmetry C18 column (3.5 μm, 150 × 2.1 mm; Waters, Milford, MA, USA) held at 40 °C. Analyses were performed with mobile phases A (H₂O) and B (CH₃CN), each of which contained 189 190 formic acid (0.1% v/v). Gradient elution (0.3 mL/min) was from 15–100% B over 21 min, then held at 191 100% B (6 min), then returned to 15% B over 0.1 min with a hold at 15% B (2.9 min) to equilibrate the 192 column (total run time 30 min), and the injection volume was 2 µL. Full-scan MS data were acquired 193 at 2–27 min, using alternating positive (m/z 500–1400) and negative (m/z 750–1400) ion scan modes, 194 with spray voltages of ±3.7 kV, a capillary temperature of 350 °C, sheath and auxiliary gas flow rates of 25 and 8 units, respectively, a resolution setting of 60,000, an AGC target of 1 ×10⁶, and a max IT of 195 196 100 ms. Extracted-ion chromatograms were obtained using exact m/z values with the mass tolerance 197 set to ±5 ppm.

198 2.5. Basil sample survey

Commercial samples (n = 50) were collected from Belgian food stores between February and October
 2022 to screen for the natural occurrence of MCs in basil. The samples were collected from 12
 supermarket chains covering 13 brands. More information on sample number, packaging and sample
 type, sampling date, and origin are available in Table S3.

203 2.6. Statistical analysis

204 Data processing and statistical analysis were performed with GraphPad Prism version 10.2.0 205 (GraphPad Software, Boston, Massachusetts USA). A Two-way ANOVA test was used to investigate the 206 effect of two independent variables (1- different treatments and 2- exposure and depuration) on the detected MC-LR concentrations. Multiple comparisons were performed for all the tested groups and 207 208 Tukey's test was selected as a recommended post-hoc analysis to identify differences between mean 209 values. Normal (Gaussian) distribution was verified using the Shapiro-Wilk normality test. The 210 threshold for *P*-value comparison was set at 0.05 for the significant difference. To investigate the 211 effect of different treatments on the length and weight of basil plant, a two-way repeated measures 212 ANOVA test was applied after checking the sphericity and normality. The standard deviation (SD) of 213 the measured values is presented after each value of the calculated mean for these measurements. 214 The SD represents the spread of biological repeats and thus does not constitute a measure of

- 215 uncertainty. On the other hand, the absolute error for the depuration was calculated using the square
- root of the error sum of squares and, therefore, constitutes a measure of uncertainty.

217 **3. Results**

218 **3.1. Effect of MC-LR on basil growth**

As a preliminary step to investigate whether there is an effect of MC-LR on basil growth, the weight and length of each individual plant were measured before and after the MC-LR exposure (seven days post-spiking). Also, the plant leaves and roots were visually inspected before, during and after the experiment. There were no significant differences in the average basil weight and length between the control and the treatment groups (**Figure 1**). Furthermore, there no lesions or discoloration were observed on the basil plants. This indicate that the exposure to MC-LR at 5, 20, or 50 μ g L⁻¹ for one week did not have any noticeable inhibitory effects.



226



²²⁹ Before the exposure to MC-LR, the average weight of basil plants for the untreated control was 6.0

(SD = 1.1) g, and for the 50 μ g L⁻¹ group was 7.6 (SD = 1.7) g, while the average length for the untreated

- control was 52.9 (SD = 5.1) cm and for the 50 μ g L⁻¹ group was 49.2 (SD = 5.5) cm. After the exposure
- to MC-LR, the average weights were 6.6 (SD = 1.4) g and 7.8 (SD = 1.6) g, and average lengths were
- 50.0 (SD = 1.9) and 46.5 (SD = 1.5) cm for the untreated basil plants and 50 μ g L⁻¹ group, respectively.

Averages of the weight and length of each treatment group before and after exposure to MC-LR are detailed in **Table S4**. Similarly, no significant differences were observed within and among the treatment groups during the depuration part of the experiment (data not shown). Decay of the primary root was observed after seven days for all the plants, independent of the growth condition, after they were transferred from soil to the Hoagland solutions, with multiple secondary roots replacing the primary roots.

3.2. Validation of a UHPLC–MS/MS method for eight MC congeners and NOD-R quantification in basil matrix

242 Multiple validation parameters were checked for the eight MCs and NOD-R quantification in the basil 243 matrix. The specificity of the method was shown by a lack of MC- and NOD-specific peaks in the blank 244 samples. Moreover, the ion ratios calculated for the points in the calibration curve were consistent 245 with the parameters provided by Directive 96/23/EC of the European Commission (Table S5). The 246 coefficient of determination (R^2) of the linear calibration curve for all the MCs and NOD-R showed the appropriateness ($R^2 > 0.99$) of the linear regression during quantification (**Table S6**). Additionally, 247 comparisons of the slopes of calibration curves in the blank matrix and solvent (50:50 MeOH-H₂O, 248 249 containing 1% acidic acid), using t-tests, showed matrix effects for all MCs during quantification (Table 250 S6). Recovery, repeatability, reproducibility, and MU were calculated at three concentration levels: 1, 5, and 25 ng g^{-1} for MC-LR and NOD-R; 5, 10, and 25 ng g^{-1} for MC-RR; and 2.5, 10, and 25 ng g^{-1} for 251 252 MC-YR, MC-LA, MC-LF, MC-WR, MC-LW, MC-LY. A complete overview of the validation data (recovery, 253 repeatability, reproducibility, and MU) can be found in **Table S7**. Averages for recovery, repeatability, 254 reproducibility, and MU are reported in Table 1. Repeatability and reproducibility remained below 255 their calculated Horwitz ratios, 14.7% and 22.0%, respectively.

The LOQ value was considered to be the lowest validated concentration: 1 ng g⁻¹ for MC-LR and NOD-R; 5 ng g⁻¹ for MC-RR, and 2.5 ng g⁻¹ for MC-YR, MC-LA, MC-LF, MC-WR, MC-LW, MC-LY. The signal-tonoise (S/N) for the quantifier and qualifier analyte peak was always > 10. To determine the LODs, the quantifier and qualifier peak should have S/N > 3. The lowest point in the calibration curve for each MC was selected as the LOD, as these were the lowest concentrations at which S/N was determined.

| Validation | Validation | | | | | | | | |
|----------------------------|------------|-------|-------|-------|------------|--------|-------|-------|-----------|
| parameter | | | | Cyan | opacterial | luxins | | | |
| | MC-RR | MC-LA | MC-LF | MC-LR | MC-LY | MC-LW | MC-YR | MC-WR | NOD- R |
| Repeatability (%) | 6.2 | 5.0 | 6.0 | 4.4 | 4.9 | 4.2 | 6.4 | 7.6 | 5.5 |
| Reproducibility (%) | 9.7 | 21.2 | 20.4 | 17.1 | 20.7 | 20.5 | 9.6 | 14.0 | 12.1 |
| MU (%) | 19.4 | 42.4 | 40.8 | 34.1 | 41.4 | 41.0 | 19.7 | 28.0 | 24.2 |
| Recovery (%) | 78.7 | 89.7 | 81.3 | 90.0 | 89.3 | 83.3 | 83.0 | 82.0 | 82.7 |
| LOD (µg kg ⁻¹) | 4.5 | 1.5 | 1.5 | 0.6 | 1.5 | 1.5 | 1.5 | 1.5 | 0.6 |
| LOQ (µg kg ⁻¹) | 5.0 | 2.5 | 2.5 | 1.0 | 2.5 | 2.5 | 2.5 | 2.5 | 1.0 |

Table 1. Averages for recovery, repeatability, reproducibility, measurement uncertainty (MU), LODs, and LOQs
 for MCs and NOD-R in basil leaf matrix.

264

265 **3.3. Accumulation and depuration of MC-LR in basil**

Exposure of basil plant to higher doses of MC-LR resulted in more accumulation in basil roots and 266 267 leaves (i.e., higher accumulation occurred with a higher concentration of MC-LR in the Hoagland 268 solution). Moreover, MC-LR was detected in higher concentrations in roots than in leaves for all three treatment groups (Figure 2). The mean values of the MC-LR concentration in basil leaves were 1.6 (SD 269 = 0.8), 5.7 (SD = 2.0), and 24 (SD = 15) ng g⁻¹ fresh weight (fw) for the 5, 20, and 50 μ g L⁻¹ treatment 270 groups, respectively. In roots, the MC-LR concentrations were 8.5 (SD = 3.1), 17.3 (SD = 3.1), and 49 271 (SD=17) ng g^{-1} fw for the 5 μ g L⁻¹, 20 μ g L⁻¹, and 50 μ g L⁻¹ treatment groups, respectively. During the 272 273 preliminary experiment with the culture extract, both MC-LR and MC-YR accumulated in the roots and leaves, similar to their ratio in the original extract (Table S1). This preliminary test shows that MCs 274 275 other than MC-LR could also accumulate in plants, which is essential because multiple MCs are often 276 present in natural cyanobacterial blooms.

To investigate the depuration potential of basil plants, the depuration percentage was calculated usinga simple formula:

| 279 | $MC - LR$ concentration after depuration period $\times 100$ |
|-----|---|
| | MC - LR concentration after accumulation period |
| 280 | The percentage of MC-LR depuration from the leaves at lower MC-LR concentrations was higher than |
| 281 | the percentage depurated at higher concentrations (5 μ g L ⁻¹ > 20 μ g L ⁻¹ > 50 μ g L ⁻¹). For the 5 μ g L ⁻¹ |
| 282 | treatment group, no MC-LR was detectable (LOD = 0.6 ng g^{-1}) after seven days in leave samples, the |
| 283 | depuration was \geq 62 ± 11%. The depuration percentages were 34 ± 8% (mean MC-LR was 4 (SD = 2) ng |
| 284 | g^{-1} fw) and 32 ± 14% (mean MC-LR 16 (SD = 14) ng g^{-1} fw) for the 20 and 50 µg L^{-1} treatment groups, |

respectively (**Table S8**). MC-LR concentrations in leaves after the depuration testing period were lower, but not significantly different, than MC-LR concentrations after the accumulation period for the three treatment groups (**Figure 2**).

288 On the other hand, the average MC-LR concentrations in roots were significantly lower after 289 depuration than in the corresponding treatment group after the accumulation period (**Figure 2**). The 20 μ g L⁻¹ treatment group showed the lowest depuration percentage (53 ± 19%), with a mean of 8.0 291 (SD = 4.5) ng g⁻¹ fw MC-LR after depuration, followed by a depuration percentage of 57 ± 13% for the 292 5 μ g L⁻¹ treatment group, with a mean value of MC-LR at 3.7 (SD = 0.6) ng g⁻¹ fw (**Table S8**). The 293 depuration period for the 50 μ g L⁻¹ treatment group resulted in 68 ± 29% depuration and a mean 294 concentration of MC-LR of 15.5 (SD = 6.2) ng g⁻¹ fw.



295 296

Figure 2. MC-LR Concentrations in basil (leaves and roots) after the accumulation and depuration periods. Different letters
 above the boxes indicate statistically significant differences according to a two-way ANOVA test followed by Tukey's multiple comparison test as a post-hoc analysis.

- The MC-LR in the Hoagland solution was also quantified after the accumulation and depuration periods. The mean concentrations of MC-LR after accumulation were 2.1 (SD=1.2), 7.6 (SD=3.9), and
- 302 25.0 (SD=9.1) μ g L⁻¹ for the 5 μ g L⁻¹, 20 μ g L⁻¹, and 50 μ g L⁻¹ treatment groups, respectively (**Figure S1**).

303 Interestingly, MC-LR levels were below the LOD value (0.1 μ g L⁻¹) in the Hoagland solutions used for 304 the depuration experiment (Van Hassel et al., 2022b).

To further confirm the results of the accumulation and depuration of MC-LR in basil leaves, four 305 306 samples (two samples from each experiment from the 50 μ g L⁻¹ treatment group) were analyzed by two alternative analytical approaches, ELISA and LC-HRMS, at two different laboratories. Table 3 307 308 shows the results of the analyzed samples and the analysis of the same samples using UHPLC–MS/MS. 309 For the basil samples from the accumulation period, the MC-LR concentrations by ELISA were higher 310 than those detected by UHPLC–MS/MS. However, due to matrix effects, the MC-LR concentrations 311 determined by ELISA in the two basil samples from the depuration period were lower than the LOQ 312 value (0.2 ng g^{-1}). It is important to mention that the ELISA determines the total concentration of MCs 313 and could include other metabolites and conjugates than the concentration of MC-LR alone, resulting

- 314 in reporting a total MCs.
- 315 Table 3. Mean values of total MCs measured by ELISA and MC-LR measured by UHPLC–MS/MS in two samples 316 of basil leaves in the 50 μ g L⁻¹ treatment group after the accumulation and depuration periods.

| Sample | Sample number (n) | Detection method (ng g^{-1} fw) | | | | |
|--------------------------------|----------------------|--|-------------------------------|-----------------------------------|--|--|
| | | Mean MCs by ELISA | Mean MC-LR by UHPLC– MS/MS | Detection of MC-LR by LC— HRMS | | |
| Basil leaves (accumulation) | 2 | 62.3 | 47.4 | Detected | | |
| Basil leaves (depuration) | 2 | <loq (0.2)<="" td=""><td>8.0</td><td>Detected</td></loq> | 8.0 | Detected | | |

317

319 significantly lower peak area compared to the semi-quantified MC-LR in the basil samples analyzed

320 after the accumulation period (Figure 3). No conjugates of MC-LR with glutathione (GSH) or cysteine

³¹⁸ The analysis of the same samples using LC–HRMS showed that the MC-LR was depurated to give a

321 (Cys), potentially originating from metabolism, were detected during the LC–HRMS screening of the

322 samples. The identity of MC-LR in basil was further confirmed by its isotopic pattern (**Figure S2**).

323



Figure 3. Extracted-ion (*m/z* 995.5560) chromatograms of MC-LR by LC–HRMS in positive ionization mode in
 extracts of basil leaves after accumulation and depuration.

326 **3.4. Occurrence of MC-LR in basil collected from Belgian markets**

327 The UHPLC–MS/MS analysis of the basil samples (n = 50) collected from the Belgian market showed the contamination of one sample (sample ID: S22FD02911) with MC-LR at a level higher than LOD but 328 329 lower than the LOQ value (*i.e.*, between 0.6 and 1.0 ng g^{-1}). Other MC congeners and NOD-R were 330 under the LOD levels of the method in all the analyzed samples. Figure 4 shows an extracted-ion 331 chromatogram for MC-LR in the contaminated samples and the matrix-matched calibration curve. It 332 is worth noting that this market sample was collected in September 2022 (Supplementary data, Table S3), the end of the cyanobacterial bloom season in Belgium. The results further indicate that the basil 333 334 was irrigated with water contaminated with MC-LR of 'environmental' origin during its growth.



Figure 4: Extracted-ion chromatograms of an MC-LR standard in a QC sample and the commercial sample contaminated with
 MC-LR. A) MC-LR quantifier ion; B) MC-LR qualifier ion; C) MC-LR quantifier ion in the 5 ng L⁻¹ MC-LR standard in a QC sample,
 and D) MC-LR qualifier ion in the 5 ng L⁻¹ MC-LR. For each chromatogram, the toxin name, precursor- and product-ion,
 retention time, and area under the peak are described from top to bottom.

340 4. Discussion

335

341 **4.1.** Effect of MC-LR on basil growth and its accumulation and depuration potentials

Several plant species readily take up MCs, accumulating them in leaves, stems, and roots (Abdallah et al., 2021; Bittencourt-Oliveira et al., 2016; Peuthert et al., 2007; Xiang et al., 2019). This accumulation depends on various factors, including the concentration of the toxin in water or soil, the plant species, and the stage of plant growth (Xiang et al., 2019). Plants are generally more susceptible to MC uptake during their vegetative growth stage when they actively take up water and nutrients from the soil. The results obtained from the preliminary assessment of the plant length and weight before and after MC-

348 LR treatment showed that MC-LR had no detectable inhibitory effects on basil growth. Therefore, no further work or assessment was done to investigate the impact of the toxin on the plant. However, 349 350 the primary basil roots did decay after the transfer to the Hoagland solution, while smaller secondary 351 roots developed. This change in root structure probably resulted from the manipulation, as it was also 352 observed in the untreated control group. Previous work reported different potential negative or 353 positive effects of MCs (including MC-LR on plant root growth depending on the toxin concentration 354 and the plant species (Chen et al., 2012; Freitas et al., 2015; McElhiney et al., 2001; Pflugmacher et al., 2001). The effects of MC exposure on plant growth have been observed in tomatoes (Corbel et al., 355 356 2015), potato shoots and mustard seedlings (McElhiney et al., 2001), wheat, carrots, cucumber, beans, and spinach (Bittencourt-Oliveira et al., 2016; Llana-Ruiz-Cabello et al., 2019; Machado et al., 2017; 357 358 Mohamed et al., 2022; Peuthert et al., 2007).

359 The results obtained during this study clearly show that basil can accumulate MC-LR when the toxin is 360 available at relevant concentrations during naturally occurring cyanobacterial blooms. Accumulation 361 of MCs, including MC-LR, through water irrigation was reported in a range of edible crops (Abdallah et al., 2021; Hereman and Bittencourt-Oliveira, 2012; Machado et al., 2017). In general, this matches 362 363 the results obtained in the current work. Peuthert et al. investigated MC accumulation both in the 364 shoots and roots of several agricultural plants (e.g., soybean, lentil, alfalfa, pea, and wheat) using ELISA (Peuthert et al., 2007). The roots of these plants accumulated MC-LR at higher concentrations than 365 366 leaves. The MCs concentration in the roots and leaves, detected with ELISA, ranged from 13 to 127 µg kg^{-1} fw and 3.1 to 31.1 μ g kg⁻¹ fw, respectively. Moreover, foliar bioaccumulation of MCs in lettuce 367 (8.3 to 178 μ g kg⁻¹ fw) was shown to be linearly proportional to the treatment concentration (0.6 to 368 12.5 μ g L⁻¹) (Hereman and Bittencourt-Oliveira, 2012), which is inconsistent with our findings in basil 369 370 roots and leaves.

371 Accumulation of MC-LR in crops could also depend on the irrigation method used (Machado et al., 372 2017). Earlier ELISA research established that spray irrigation with water contaminated with MC-LR 373 results in accumulation in leafy crops (Bittencourt-Oliveira et al., 2016; Codd et al., 1999; Crush et al., 374 2008). Our study showed that MC-LR can also be taken up through the roots and transported to the leaves. In previous research, lettuce and spinach were exposed to a mixture of MCs and 375 376 cylindrospermopsin, each at 5 and 25 μ g L⁻¹, using hydroculture, resulting in 0.2±0.1 to 1.3±0.1 μ g kg⁻¹ fw of MCs in roots and no detected MCs (LOD = 0.06 μ g kg⁻¹ fw) in leaves while using UHPLC-MS/MS 377 378 (Llana-Ruiz-Cabello et al., 2019). Similarly, MC-LR was quantified with UHPLC-photodiode-array (PDA) in edible parts of rice and water spinach (Ipomoea aquatica), 429.8±4.4 to 567.5±4.9 μ g kg⁻¹ fw and 379 350.8 \pm 2.9 µg kg⁻¹ fw respectively, after exposure to cyanobacteria-contaminated water through 380

hydroculture (Wijewickrama and Manage, 2019). Since plants are constantly in contact with the enriched solution, the bioavailability of MCs in hydroculture cultures might be higher than in soilbased systems because, in soil, the toxins can be adsorbed to clay minerals and organic matter, flow to deeper layers, or be degraded by microorganisms (Zhang et al., 2021).

385 The depuration of natural toxins by plants is a well-known phenomenon. A few studies have 386 documented this for MCs, including MC-LR in edible crops. One study showed that 25% of the accumulated MC-LR in lettuce leaves could be depurated after seven days of exposure (10 μ g L⁻¹ MC-387 388 LR), followed by seven days of irrigation with clean water, while full detoxification of the toxin was 389 estimated to occur after 37 days (Cordeiro-Araújo et al., 2016). The study also reported that the 390 depuration was less efficient at higher levels or concentrations of exposure. In a second study, MC-LR depuration rates of 9.5 and 8.1 µg kg⁻¹ dw day⁻¹ were reported over 12 days for lettuce and spinach 391 392 leaves, respectively (Cao et al., 2019). Those depuration rates were obtained after a 12-day bioaccumulation period of irrigation with water containing MC-LR at 10 μ g L⁻¹ preceded the 393 394 depuration period of equal length (Cao et al., 2019). These two studies are consistent with our 395 findings, with the observed trends in the accumulation and the depuration of MC-LR in basil plants 396 depending on the MC concentration in the Hoagland solution. However, further investigation of the 397 MC-LR depuration in edible crops, including basil, should be considered under other lab and field conditions. 398

399 Different metabolic pathways may be relevant for the derivatization of MCs in plants (Cao et al., 2019; 400 Chen et al., 2012). However, in our study, no known conjugates of MC-LR were detected during LC-401 HRMS screening of the basil leaf samples. An explanation could be that conjugates might be present 402 at ultra-low concentrations in the basil samples and, therefore, not possible to detect with our current 403 methods. Metabolic pathways should be examined in the future with other approaches. For instance, 404 radio-labeled MCs could be used to explore the mechanics and kinetics of accumulation and 405 depuration, as was previously applied to study MC-LR accumulation in tomatoes (Corbel et al., 2016). 406 Crop type should also be taken into account when studying metabolic mechanisms. Finally, 407 understanding the depuration dynamics for different crops might allow the development of 408 approaches to salvage crop harvests exposed to irrigation water contaminated with MCs.

409 **4.2.** Natural occurrence of MC-LR in basil samples collected from the Belgian market

410 The detection of MC-LR in basil leaves during this (limited) survey further showed that human 411 exposure to MC-LR through food crops is a growing concern, even with the presence of MC-LR in only 412 one basil sample. The concentration of MC-LR in the sample can be estimated between 0.6 (LOD) and

1 ng g^{-1} (LOQ), which would probably not constitute a health risk. However, it does confirm that water 413 containing naturally occurring toxic cyanobacterial blooms is being used to irrigate crops in Belgium. 414 415 Until now, two studies found higher MC accumulation in multiple field samples in Asia irrigated with 416 contaminated water (Wijewickrama and Manage, 2019; Xiang et al., 2019). These field samples 417 consisted of various crops (e.g., rice, beans, fruits, leafy and root vegetables). Rice and water spinach 418 contained 21 and 133 µg kg⁻¹ fw MC-LR, respectively, as quantified with UHPLC- PDA (Wijewickrama and Manage, 2019). Leafy vegetables, including lettuce, celery, cabbage, and spinach, contained the 419 420 highest concentration of MCs (mean values from 9.2 to 118 μ g kg⁻¹ fw) compared to fruit (mean values 421 from 1.4 to 47 μ g kg⁻¹ fw) and root vegetables (mean values from 4.9 to 17 μ g kg⁻¹ fw) (Xiang et al., 2019). For some samples, the content of MCs exceeded the tolerable daily intake (i.e., maximum 422 423 allowed daily exposure to a compound before it becomes harmful) (World Health Organization (WHO), 2020; Xiang et al., 2019). The current work shows that MC-LR accumulation in basil is concentration-424 425 dependent. Similar results were observed as vegetables cultivated around two Chinese lakes with higher concentrations of MCs (mean values from 66 to 276 μ g L⁻¹ fw) accumulated more MCs 426 compared to vegetables associated with a third lake with lower a lower concentration of MCs (mean 427 values from mean 7.7 to 53 μ g L⁻¹) (Xiang et al., 2019). Consequently, higher concentrations of MC-LR 428 429 in the irrigation water will probably lead to a significantly higher level of bioaccumulated MC-LR in the 430 edible parts of the plant. This scenario is also likely in Belgium due to prevalent cyanobacterial blooms 431 throughout Belgium and the dwindling of available water sources during climate-change-driven hot 432 summers. Ideally, the occurrence of cyanotoxins in crops on the Belgian market should be regularly monitored in the future. Additionally, processed foods based on basil leaves, where their water 433 434 content is reduced (e.g., pesto sauce and herbs), could lead to a concentration of the toxin, resulting 435 in potential health risks. The effect of storage, cooling, and heating on MC-LR stability during the 436 preparation of basil-based foods and herbs should be considered for future studies.

437 **5. Conclusions**

Accumulation of MCs in food crops is becoming more relevant since it poses an environmental and 438 public health risk. The current study showed that basil can bioaccumulate MC-LR in roots and leaves 439 440 when grown in a hydroculture system with MC-LR-contaminated water. More MC-LR concentrations 441 were detected in basil roots and leaves at higher MC-LR exposure doses. The bioaccumulations were 442 significantly higher in roots than in leaves for all the treatment conditions. Partial depuration was 443 achieved when basil plants were placed in a clean Hoagland solution. The detected MC-LR levels after 444 the depuration were significantly lower than the accumulated MC-LR levels in roots but not in leaves. 445 Both accumulation and depuration of MC-LR were analyzed using UHPLC-MS/MS, ELISA, and LC- 446 HRMS. Basil plants did not show any obvious signs of phytotoxicity since no significant differences in the weight and length were found between the different basil groups before and after the treatment. 447 448 As the LC-HRMS analysis did not show any MC-LR conjugates formed by the plant, more research is 449 required to unravel the depuration mechanism of MC-LR in basil. Although the collection of basil 450 samples from the market showed a contamination of only one sample with MC-LR, this confirms that 451 contamination via cyanotoxin-contaminated irrigation water is a fact in Belgium. Given that the effects 452 of climate change are getting more serious, the occurrence of cyanotoxins in crops might increase. This study describes the first detection of MC-LR in basil and food crops from the Belgian market. 453 454 Therefore, more research on MC accumulation in basil is needed, and yearly monitoring of human 455 food crops for MCs is recommended.

456 **CRediT authorship contribution statement**

457 Wannes H. R. Van Hassel conceptualization, conducting LC-MS/MS analysis, data processing, and 458 writing the original draft. Mohamed F. Abdallah conceptualization, data processing, writing the 459 original draft, and study supervision. Maria Gracia Guzman Velasquez conducted experiments, LC-460 MS/MS analysis, data processing, and writing the original draft. Christopher O. Miles conducted the LC–HRMS analysis, data processing, and manuscript revising. Ingunn A. Samdal conducted the ELISA 461 462 analysis, data processing, and manuscript revising. Julien Masquelier supervised the LC-MS/MS analysis and manuscript revising. Andreja Rajkovic conceptualization, funding acquisition, providing 463 464 consumables, manuscript revising, project coordination, and study supervision.

465 **Funding:**

This work received funding from the Cyantir project, funded by the Federal Public Service (Health,
Food Chain Safety and Environment), and the EU IMPTOX project (grant agreement No 965173).

468 Acknowledgments

Wannes Hugo R. Van Hassel received a full PhD scholarship within the Cyantir project. Mohamed F.
Abdallah received a postdoctoral mandate from the Ghent University Special Research Fund (BOF),
grant number BOF 01P03220. The ULC/BCCM provided the *Microcystis* ULC642 strain used to produce
the MCs mixture (https://bccm.belspo.be/about-us/bccm-ulc).

473 **Declaration of Competing Interest**

- 474 The authors declare that they have no known competing financial interests or personal relationships
- that could have appeared to influence the work reported in this paper.

476 Supplementary material

477 Tables S1–7. Figures S1–2.

478 Data availability

479 Data will be made available on request.

480 **References**

- Abdallah, M.F., Van Hassel, W.H.R., Andjelkovic, M., Wilmotte, A., Rajkovic, A., 2021. Cyanotoxins
 and food contamination in developing countries: Review of their types, toxicity, analysis,
- 483 occurrence and mitigation strategies. Toxins (Basel) 13, 786.
- 484 https://doi.org/10.3390/toxins13110786
- Bittencourt-Oliveira, M. do C., Cordeiro-Araújo, M.K., Chia, M.A., Arruda-Neto, J.D. de T., Oliveira,
 Ê.T. de, Santos, F. dos, 2016. Lettuce irrigated with contaminated water: Photosynthetic
 effects, antioxidative response and bioaccumulation of microcystin congeners. Ecotoxicol
 Environ Saf 128, 83–90. https://doi.org/10.1016/j.ecoenv.2016.02.014
- Cao, Q., Liu, W., Jiang, W., Shu, X., Xie, L., 2019. Glutathione biosynthesis plays an important role in
 microcystin-LR depuration in lettuce and spinach. Environ Pollut 253, 599–605.
 https://doi.org/10.1016/j.envpol.2019.07.064
- 492 Centre for the Promotion of Imports from developing countries, (CBI), 2020. The European market
 493 potential for fresh herbs. https://www.cbi.eu/market-information/fresh-fruit-vegetables/fresh 494 herbs/market-potential (accessed 01.02.2024).
- Chen, J., Han, F.X., Wang, F., Zhang, H., Shi, Z., 2012. Accumulation and phytotoxicity of microcystinLR in rice (*Oryza sativa*). Ecotoxicol Environ Saf 76, 193–199.
 https://doi.org/10.1016/j.ecoenv.2011.09.022
- Chen, J., Song, L., Dai, J., Gan, N., Liu, Z., 2004. Effects of microcystins on the growth and the activity
 of superoxide dismutase and peroxidase of rape (*Brassica napus* L.) and rice (*Oryza sativa* L.).
 Toxicon 43, 393–400. https://doi.org/10.1016/J.TOXICON.2004.01.011
- Codd, G.A., Metcalf, J.S., Beattie, K.A., 1999. Retention of Microcystis aeruginosa and microcystin by
 salad lettuce (*Lactuca sativa*) after spray irrigation with water containing cyanobacteria.
 Toxicon 37, 1181–1185. https://doi.org/10.1016/S0041-0101(98)00244-X
- 504 Corbel, S., Bouaïcha, N., Nélieu, S., Mougin, C., 2015. Soil irrigation with water and toxic
 505 cyanobacterial microcystins accelerates tomato development. Environ Chem Lett 13, 447–452.
 506 https://doi.org/10.1007/s10311-015-0518-2
- 507 Corbel, S., Mougin, C., Nélieu, S., Delarue, G., Bouaïcha, N., 2016. Evaluation of the transfer and the
 508 accumulation of microcystins in tomato (*Solanum lycopersicum* cultivar MicroTom) tissues
 509 using a cyanobacterial extract containing microcystins and the radiolabeled microcystin-LR
- 510 (14C-MC-LR). Sci Total Environ 541, 1052–1058.
- 511 https://doi.org/10.1016/j.scitotenv.2015.10.004

Cordeiro-Araújo, M.K., Chia, M.A., Arruda-Neto, J.D. de T., Tornisielo, V.L., Vilca, F.Z., Bittencourt Oliveira, M. do C., 2016. Microcystin-LR bioaccumulation and depuration kinetics in lettuce and
 arugula: Human health risk assessment. Sci Total Environ 566–567, 1379–1386.

515 https://doi.org/10.1016/j.scitotenv.2016.05.204

516 Cordeiro-Araújo, M.K., Chia, M.A., Bittencourt-Oliveira, M. do C., 2017. Potential human health risk
517 assessment of cylindrospermopsin accumulation and depuration in lettuce and arugula.
518 Harmful Algae 68, 217–223. https://doi.org/10.1016/j.hal.2017.08.010

- Crush, J.R., Briggs, L.R., Sprosen, J.M., Nichols, S.N., 2008. Effect of irrigation with lake water
 containing microcystins on microcystin content and growth of ryegrass, clover, rape, and
 lettuce. Environ Toxicol 23, 246–252. https://doi.org/10.1002/tox.20331
- Freitas, M., Azevedo, J., Pinto, E., Neves, J., Campos, A., Vasconcelos, V., 2015. Effects of microcystin LR, cylindrospermopsin and a microcystin-LR/cylindrospermopsin mixture on growth, oxidative
 stress and mineral content in lettuce plants (*Lactuca sativa* L.). Ecotoxicol Environ Saf 116, 59–
 67. https://doi.org/10.1016/j.ecoenv.2015.02.002
- Hereman, T.C., Bittencourt-Oliveira, M. do C., 2012. Bioaccumulation of microcystins in lettuce. J
 Phycol 48, 1535–1537. https://doi.org/10.1111/jpy.12006
- Hoagland, D.R., Arnon, D.I., 1950. The water-culture method for growing plants without soil. Circ Calif Agric Exp Stn 347, 1–32.
- IARC Working Group on the Evaluation of Carcinogenic Risks to Humans, 2010. IARC monographs on
 the evaluation of carcinogenic risks to humans. Ingested nitrate and nitrite, and cyanobacterial
 peptide toxins., IARC Monogr Eval Carcinog Risks Hum 94, v–vii, 1–412.
- Ibelings, B.W., Backer, L.C., Kardinaal, W.E.A., Chorus, I., 2014. Current approaches to cyanotoxin risk
 assessment and risk management around the globe. Harmful Algae 40, 63–74.
 https://doi.org/10.1016/j.hal.2014.10.002
- Levizou, E., Papadimitriou, T., Papavasileiou, E., Papadimitriou, N., Kormas, K.A., 2020. Root
 vegetables bioaccumulate microcystins-LR in a developmental stage-dependent manner under
 realistic exposure scenario: The case of carrot and radish. Agric Water Manag 240, 106274.
 https://doi.org/10.1016/J.AGWAT.2020.106274
- Levizou, E., Statiris, G., Papadimitriou, T., Laspidou, C.S., Kormas, K.A., 2017. Lettuce facing
 microcystins-rich irrigation water at different developmental stages: Effects on plant
 performance and microcystins bioaccumulation. Ecotoxicol Environ Saf 143, 193–200.
 https://doi.org/10.1016/J.ECOENV.2017.05.037
- Llana-Ruiz-Cabello, M., Jos, A., Cameán, A., Oliveira, F., Barreiro, A., MacHado, J., Azevedo, J., Pinto,
 E., Almeida, A., Campos, A., Vasconcelos, V., Freitas, M., 2019. Analysis of the use of
 cylindrospermopsin and/or microcystin-contaminated water in the growth, mineral content,
 and contamination of *Spinacia oleracea* and *Lactuca sativa*. Toxins (Basel) 11, 624.
 https://doi.org/10.3390/toxins11110624
- Machado, J., Campos, A., Vasconcelos, V., Freitas, M., 2017. Effects of microcystin-LR and
 cylindrospermopsin on plant-soil systems: A review of their relevance for agricultural plant
 quality and public health. Environ Res. https://doi.org/10.1016/j.envres.2016.09.015

- McElhiney, J., Lawton, L.A., Leifert, C., 2001. Investigations into the inhibitory effects of microcystins
 on plant growth, and the toxicity of plant tissues following exposure. Toxicon 39, 1411–1420.
 https://doi.org/10.1016/S0041-0101(01)00100-3
- Mohamed, Z., Bakr, A., Campos, A., Vasconcelos, V., Nasr, S.A.M., 2022. Growth inhibition and
 microcystin accumulation in bush bean (*Phaseolus vulgaris* L.) plant irrigated with water
 containing toxic *Chrooccocus minutus*. Agric Water Manag 261, 107381.
- 558 https://doi.org/10.1016/j.agwat.2021.107381
- Mohamed, Z.A., Al Shehri, A.M., 2009. Microcystins in groundwater wells and their accumulation in
 vegetable plants irrigated with contaminated waters in Saudi Arabia. J Hazard Mater 172, 310–
 315. https://doi.org/10.1016/j.jhazmat.2009.07.010
- 562 O'Neil, J.M., Davis, T.W., Burford, M.A., Gobler, C.J., 2012. The rise of harmful cyanobacteria blooms:
 563 The potential roles of eutrophication and climate change. Harmful Algae 14, 313–334.
 564 https://doi.org/10.1016/j.hal.2011.10.027
- Peuthert, A., Chakrabarti, S., Pflugmacher, S., 2007. Uptake of microcystins-LR and -LF
 (cyanobacterial toxins) in seedlings of several important agricultural plant species and the
 correlation with cellular damage (lipid peroxidation). Environ Toxicol 22, 436–442.
 https://doi.org/10.1002/tox.20266
- Pflugmacher, S., Wiegand, C., Beattie, K.A., Krause, E., Steinberg, C.E.W., Codd, G.A., 2001. Uptake,
 effects, and metabolism of cyanobacterial toxins in the emergent reed plant *Phragmites australis* (CAV.) Trin. Ex Steud. Environ Toxicol Chem 20, 846–852.
 https://doi.org/10.1002/etc.5620200421
- Redouane, E.M., Tazart, Z., Lahrouni, M., Mugani, R., Elgadi, S., Zine, H., Zerrifi, S.E.A., Haida, M.,
 Martins, J.C., Campos, A., Oufdou, K., Vasconcelos, V., Oudra, B., 2023. Health risk assessment
 of lake water contaminated with microcystins for fruit crop irrigation and farm animal drinking.
 Environ Sci Pollut Res 30, 80234–80244. https://doi.org/10.1007/s11356-023-27914-1
- Romero-Oliva, C.S., Contardo-Jara, V., Block, T., Pflugmacher, S., 2014. Accumulation of microcystin
 congeners in different aquatic plants and crops—A case study from lake Amatitlán, Guatemala.
 Ecotoxicol Environ Saf 102, 121–128. https://doi.org/10.1016/j.ecoenv.2014.01.031
- Saha, S., Monroe, A., Day, M.R., 2016. Growth, yield, plant quality and nutrition of basil (*Ocimum basilicum* L.) under soilless agricultural systems. Ann Agric Sci 61, 181–186.
 https://doi.org/10.1016/j.aoas.2016.10.001
- Samdal, I.A., Ballot, A., Løvberg, K.E., Miles, C.O., 2014. Multihapten approach leading to a sensitive
 ELISA with broad cross-reactivity to microcystins and nodularin. Environ Sci Technol 48, 8035–
 8043. https://doi.org/10.1021/es5012675
- Svirčev, Z., Lalić, D., Bojadžija Savić, G., Tokodi, N., Drobac Backović, D., Chen, L., Meriluoto, J., Codd,
 G.A., 2019. Global geographical and historical overview of cyanotoxin distribution and
 cyanobacterial poisonings. Arch Toxicol. https://doi.org/10.1007/s00204-019-02524-4
- Trajkovska-Broach, A., Anka, |, Petkoska, T., 2023. Mediterranean herbs, spices, and medicinal
 plants—natural remedies and rich sources of bioactive compounds. JSFA Reports 3, 4–12.
 https://doi.org/10.1002/JSF2.96

- Tsoumalakou, E., Papadimitriou, T., Berillis, P., Kormas, K.A., Levizou, E., 2021. Spray irrigation with
 microcystins-rich water affects plant performance from the microscopic to the functional level
 and food safety of spinach (*Spinacia oleracea* L.). Sci Total Environ 789, 147948.
 https://doi.org/10.1016/J.SCITOTENV.2021.147948
- Van Hassel, W.H.R., Andjelkovic, M., Durieu, B., Marroquin, V.A., Masquelier, J., Huybrechts, B.,
 Wilmotte, A., 2022a. A summer of cyanobacterial blooms in Belgian waterbodies: Microcystin
 quantification and molecular characterizations. Toxins (Basel) 14, 61.
- 599 https://doi.org/10.3390/toxins14010061
- Van Hassel, W.H.R., Huybrechts, B., Masquelier, J., Wilmotte, A., Andjelkovic, M., 2022b.
 Development, validation and application of a targeted LC–MS method for quantification of
 microcystins and nodularin: Towards a better characterization of drinking water. Water (Basel)
 14, 1195. https://doi.org/10.3390/w14081195
- Van Hassel, W.H.R., Masquelier, J., Andjelkovic, M., Rajkovic, A., 2022c. Towards a better
 quantification of cyanotoxins in fruits and vegetables: Validation and application of an UHPLC MS/MS-based method on Belgian products. Separations 9, 319.
- 607 https://doi.org/10.3390/separations9100319
- Wijewickrama, M.M., Manage, P.M., 2019. Accumulation of microcystin-LR in grains of two rice
 varieties (*Oryza sativa* L.) and a leafy vegetable, *Ipomoea aquatica*. Toxins (Basel) 11.
 https://doi.org/10.3390/toxins11080432
- Willame, R., Jurczak, T., Iffly, J.F., Kull, T., Meriluoto, J., Hoffmann, L., 2005. Distribution of
 hepatotoxic cyanobacterial blooms in Belgium and Luxembourg. Hydrobiologia 551, 99–117.
 https://doi.org/10.1007/S10750-005-4453-2/METRICS
- World Health Organization, 2020. Cyanobacterial toxins: Microcystins. Background document for
 development of WHO Guidelines for drinking-water quality and Guidelines for safe recreational
 water environments. WHO, Geneva, Switzerland.
- Kiang, L., Li, Y.W., Liu, B.L., Zhao, H.M., Li, H., Cai, Q.Y., Mo, C.H., Wong, M.H., Li, Q.X., 2019. High
 ecological and human health risks from microcystins in vegetable fields in southern China.
 Environ Int 133, 105142. https://doi.org/10.1016/j.envint.2019.105142
- Zhang, Y., Whalen, J.K., Sauvé, S., 2021. Phytotoxicity and bioconcentration of microcystins in
 agricultural plants: Meta-analysis and risk assessment. Environ Pollut 272, 115966.
 https://doi.org/10.1016/j.envpol.2020.115966

623

Experimental accumulation and depuration kinetics and natural occurrence of microcystin-LR in basil

Highlights:

- Microcystin-LR accumulates dose-dependently in basil grown on a hydroculture medium. ٠
- Basil plants significantly depurated microcystin-LR after one week of no toxin exposure. •
- The possible detoxification mechanisms in the plant are yet to be verified.
- The occurrence of microcystin-LR was first-time detected in basil intended for human consumption. •

Declaration of interests

□ The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

☑ The authors declare the following financial interests/personal relationships which may be considered as potential competing interests:

Wannes Hugo R. Van Hassel reports financial support was provided by Federal Public Service Health Food Chain Safety and Environment. Mohamed F. Abdallah reports financial support was provided by Ghent university special research fund. Andreja Rajkovic reports financial support was provided by EU Framework Programme for Research and Innovation Euratom. If there are other authors, they declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

Journal Prerk