

1           **Fe(III)-alginate and layered aluminum oxyhydroxide**  
2           **assisted hydro ceramsite composite for efficient removal of**  
3           **Se and As from high-concentration sulfate wastewater**

4  
5   Jingya Ren<sup>a, \*</sup>, Karen Leus<sup>b</sup>, Rino Morent<sup>b</sup>, Nathalie De Geyter<sup>b</sup>, Pascal

6                                   Van Der Voort<sup>c</sup>, Gijs Du Laing<sup>a</sup>

7   <sup>a</sup> *Center for Advanced Process Technology and Urban Resource Recovery (CAPTURE),*  
8   *Department of Green Chemistry and Technology, Ghent University, Frieda Saeynsstraat*  
9   *1, 9052 Ghent, Belgium*

10   <sup>b</sup> *RUPT-Research Unit Plasma Technology, Department of Applied Physics, Ghent*  
11   *University, Sint-Pietersnieuwstraat 41 B4, 9000 Ghent, Belgium*

12   <sup>c</sup> *COMOC-Center for Ordered Materials, Organometallics and Catalysis, Department*  
13   *of Chemistry, Ghent University, Krijgslaan 281-S3, 9000 Ghent, Belgium*

14  
15   \*Corresponding authors.

16   E-mail addresses: [Jingya.Ren@UGent.be](mailto:Jingya.Ren@UGent.be); Tel: +32-456352421

17 **ABSTRACT:** In this work, hydro ceramsite (HC) was coated by layered aluminum  
18 oxyhydroxide (LAO) with the assistance of Iron(III)-alginate (IA) to obtain HC@LAO-  
19 IA. Iron(III)-alginate (IA) and layered aluminum oxyhydroxide (LAO) assist in the  
20 improvement of the adsorption capacity of hydro ceramsite (HC) for Se(IV), Se(VI),  
21 and As(V) via hydrogen bonding, electrostatic adsorption, and inner sphere  
22 complexation, respectively. HC@LAO-IA showed superior removal of Se(IV), Se(VI),  
23 and As(V) under simulated and actual environmental conditions in the presence of  
24 multiple competing anions ( $\text{Cl}^-$ ,  $\text{NO}_3^-$ ,  $\text{CO}_3^{2-}$ ,  $\text{PO}_4^{3-}$  and  $\text{SO}_4^{2-}$ ), particularly excelling  
25 in high-concentration sulfate wastewater, and also has a wide pH application range of  
26 3-10. After four reuse cycles, HC@LAO-IA still exhibits satisfactory performance.  
27 Sand column experiments also demonstrated that HC@LAO-IA has excellent  
28 capabilities for the removal of Se(IV), Se(VI), and As(V) from real wastewater obtained  
29 from a metallurgical process plant. This work provides a promising, low-cost approach  
30 to remove Se and As from aqueous solutions and a new route for ceramsite recycling.

31 **Keywords:**

32 Hydro ceramsite, Layered aluminum oxyhydroxide, Selenate, Selenite, Arsenate

## 33 **1. Introduction**

34 Industrial activities, such as metal casting and coal mining, generate wastewater  
35 that contains toxic substances like selenium (Se) and arsenic (As), which can pose  
36 serious health risks to both aquatic animals and humans [1-4]. To mitigate these risks,  
37 industrial wastewater is often treated using chemical precipitation techniques, such as  
38 the addition of sulfate, sulfide or phosphate, to remove cationic heavy metals [5,6].  
39 However, selenate and selenite, which are oxyanions with a chemical structure similar  
40 to sulfate and sulfite, are also present in wastewater. Sorption processes are an option  
41 to remove these oxyanions. Still, they can be difficult to remove from these wastewaters  
42 through sorption due to competition for sorption sites, especially when selenate  
43 concentrations are low compared to sulfate concentrations, or when arsenate  
44 concentrations are low compared to phosphate concentrations [7-9]. Therefore, the  
45 development of a novel efficient sorption strategy is crucial for arsenic and selenium  
46 removal especially from high-sulfate wastewater in order to meet the emission or  
47 environmental quality standards set by regulatory bodies such as the World Health  
48 Organization (WHO) and the United States Environmental Protection Agency (EPA).

49 Hydro ceramsite (HC), also known as hydroton clay balls or expanded clay, is a  
50 lightweight aggregate material used in concrete, plaster, and as a soil amendment in  
51 agriculture and construction. [10-12]. Recently, the development of HC-based agents  
52 for contaminant remediation has gained attention due to their low cost, high specific  
53 surface area and ion exchange properties [13-15]. Chen et al. [16] reported a porous  
54 ceramsite produced from multiple solid wastes to remove Pb (II) by bonding with Si—  
55 O or Al—O—Si—O embedded in the matrix of ceramsite. Niu et al. [17] prepared  
56 Fe<sub>3</sub>O<sub>4</sub>-modified ceramsite to convert Cr(VI) to Cr(III) and adsorb Cr(III) by  
57 electrostatic adsorption. Nevertheless, the application of these HC-based agents still

58 presents certain limitations, such as limited adsorption capacity, low selectivity, and  
59 high production cost. Other previous studies hardly focused on the selectivity of  
60 adsorbents for Se and As in high-concentration sulfate solutions or did not evaluate the  
61 developed adsorbents using real, non-simulated Se and As contaminated real  
62 wastewater, hampering the effective valorisation and commercialisation of developed  
63 adsorbents.

64 In this work, an HC-based agent named HC@LAO-IA for highly efficient removal  
65 of Se(IV), Se(VI) or As(V) from high-concentration sulfate wastewater was fabricated.  
66 The fabrication process and mechanism involved are illustrated in Fig. 1. Raw hydro  
67 ceramsite (HC) was washed with HCl solution to remove impurities from both its  
68 surface and interior. Afterwards, HC was immersed in a mixture of layered aluminum  
69 oxyhydroxide (LAO) and sodium alginate (SA) in solution, being uniformly coated  
70 onto the surface of HC to create HC@LAO-SA. Finally, the resulting HC@LAO-SA  
71 was crosslinked in  $\text{Fe}(\text{NO}_3)_3$  solution to obtain HC@LAO-IA. The removal efficiencies  
72 of Se(IV), Se(VI) or As(V) in simulated and actual wastewater were investigated under  
73 various conditions, including the presence of competing ions. To reveal the removal  
74 mechanism, the interactions between HC@LAO-IA and Se(IV), Se(VI) or As(V) were  
75 studied. This work not only advances water treatment strategies for the removal of Se  
76 and As from high-concentration sulfate wastewater to meet stringent water quality  
77 thresholds but also presents a new direction for recycling ceramsite that has previously  
78 been used for other purposes.

## 79 **2. Materials and methods**

### 80 **2.1 Synthesis of LAO**

81 LAO, also named Ju-111 ( $\text{Al}_3\text{O}_{2.5}(\text{OH})_{3.5}\text{Cl}_{0.5}\cdot 0.9\text{H}_2\text{O}$ ), was synthesized by a  
82 simple hydrothermal method according to previous literature [18, 19]. Initially, 6.42 g

83 of CaO was dispersed in 165 mL of deionized water to form milky white suspension  
84 and stirred for 3 h. Subsequently, 27.54 g of AlCl<sub>3</sub> was added to the white suspension.  
85 All aforementioned reactions were carried out in a fume hood. After stirring the solution  
86 for 2 h, the solution gradually became transparent due to the hydrolysis reaction of  
87 AlCl<sub>3</sub>. Then, the solution was transferred to five reaction vessels (50 mL) and heated at  
88 180°C for 72 h. After cooling to room temperature, the solid product was washed 3  
89 times with 70°C deionized water and dried at 100°C. Finally, the product was taken out  
90 and ground into powder using a mortar and pestle.

## 91 **2.2 Fabrication of HC@LAO-IA**

92 Raw HC spheres (100 g) were washed with 500 mL of HCl solution (3 mol/L) to  
93 eliminate impurities from both their surface and interior. The cleaned HCs (0.4 g) were  
94 soaked in SA (2g) and LAO (0.5-2.5g) hybrid suspension (50 mL) under continuous  
95 stirring (400 rpm) for 1 h at room temperature to obtain HC@LAO-SA. Eventually, the  
96 HC@LAO-SA was immersed in Fe(NO<sub>3</sub>)<sub>3</sub> solution (0.09-0.25 mol/L) and stirred (350  
97 rpm) for 1 h, leading to the formation of HC@LAO-IA through cross-linking between  
98 SA and Fe<sup>3+</sup> [20, 21]. The resulting HC@LAO-IA was rinsed with deionized water 3  
99 times to wash away residual Fe<sup>3+</sup>. Prior to the production of HC@LAO-IA according  
100 to this procedure for use in subsequent adsorption experiments, various production  
101 conditions were tested. Results of this production process optimization experiment are  
102 presented in Fig. S1. The use of 0.4 g of HC, 50 mL of sodium alginate solution (20  
103 g/L), 2 g of LAO and 0.17 mol/L of Fe(NO<sub>3</sub>)<sub>3</sub> resulted in the highest Se(IV), Se(VI)  
104 and As(V) removal efficiency.

## 105 **2.3 Batch adsorption experiments with HC@LAO-IA**

106 In the adsorption experiments, 0.5 g of HC@LAO-IA was always added to 10 mL  
107 of Se(IV), Se(VI) or As(V) aqueous solutions. The resulting solutions were kept on a

108 shaker (200 rpm) for adsorption until equilibrium was achieved and subsequently  
109 separated using 0.45 µm membrane filters. Afterwards, the concentrations of Se(IV),  
110 Se(VI) or As(V) in solution were determined using an inductively coupled plasma-  
111 optical emission spectrometer (ICP-OES). Subsequently, the removal efficiencies of  
112 Se(IV), Se(VI) or As(V) and adsorption capacities of HC@LAO-IA (mg/g) were  
113 calculated according to equations (1) and (2):

$$114 \quad \text{RE (\%)} = (C_0 - C_t) / C_0 \times 100\% \quad (1)$$

$$115 \quad Q_e = (C_0 - C_e) \frac{V}{m} \quad (2)$$

116 where  $C_0$  and  $C_t$  are the initial and residual concentrations (mg/L).  $C_e$  is the  
117 concentration at equilibrium (mg/L), and  $Q_e$  is the adsorption capacity of HC@LAO-  
118 IA at equilibrium. The time is measured in h,  $V$  represents the volume of solution in  
119 litres, and  $m$  is the weight of HC@LAO-IA in grams. All experiments were performed  
120 in duplicate.

#### 121 **2.4 Experiments to assess factors affecting Se(IV), Se(VI) or As(V) removal by** 122 **HC@LAO-IA**

123 To evaluate the removal efficiency and capacity of HC@LAO-IA for removal of  
124 Se(IV), Se(VI) or As(V) from low and high-concentrate solutions at different pH levels,  
125 HC@LAO-IA (0.5 g) was added to 10 mL of Se(IV), Se(VI) or As(V) (10 mg/L and  
126 800 mg/L) aqueous solution at various pH conditions (2.0-11.0). pH values were  
127 adjusted with HCl (0.01 M to 1 M) and NaOH (0.01 M).

128 To investigate the influence of temperature and competing anions on Se(IV),  
129 Se(VI) or As(V) removal, 0.5 g of HC@LAO-IA was added to 10 mL of Se(IV), Se(VI)  
130 or As(V) (1 mg/L) aqueous solution with 500 mg/L of NaCl, Na<sub>2</sub>CO<sub>3</sub>, NaNO<sub>3</sub>, Na<sub>2</sub>SO<sub>4</sub>  
131 or Na<sub>3</sub>PO<sub>4</sub>, at different temperatures (10, 20, 30 and 40°C).

132 After 4 h, the HC@LAO-IA was taken out and the residual concentrations of

133 Se(IV), Se(VI) or As(V) were measured to calculate their removal efficiencies and  
 134 adsorption capacities of HC@LAO-IA (mg/g) by equations (1) and (2). All experiments  
 135 were performed in duplicate.

## 136 **2.5 Adsorption kinetics and isotherms of HC@LAO-IA for Se(IV), Se(VI) or As(V)**

137 To study the adsorption kinetics of HC@LAO-IA, 10 g of HC@LAO-IA was  
 138 added to 250 mL of Se(IV), Se(VI) or As(V) (10 mg/L) at an initial pH value of 6.0.  
 139 The resulting sample was placed on a shaker and 1 mL of the supernatant was aspirated  
 140 at specific time points (from 0 min to 600 min). After that, all collected supernatants  
 141 were diluted 10 times and quickly filtered through a 0.45  $\mu\text{m}$  filter. Finally, the  
 142 concentrations of Se and As were measured by ICP-OES. Pseudo-first-order and  
 143 pseudo-second-order models were employed to fit the data and analyze the kinetics.  
 144 Their linearized equations are given in Equations (3) and (4) [22, 23], respectively.

$$145 \quad \ln(q_e - q_t) = \ln q_e - k_1 t \quad (3)$$

$$146 \quad t/q_t = 1/(k_2 q_e^2) + t/q_e \quad (4)$$

147 Where  $k_1$  ( $\text{min}^{-1}$ ) and  $k_2$  ( $\text{g} \cdot \text{mg}^{-1} \cdot \text{min}^{-1}$ ) are rate constants for pseudo-first-order  
 148 and pseudo-second-order reactions, respectively.  $q_t$  and  $q_e$  (mg/g) represent adsorption  
 149 capacities at time  $t$  (min) and equilibrium, respectively.

150 HC@LAO-IA was added to a series of solutions with different concentrations (from  
 151 200 mg/L to 1500 mg/L) of Se(IV), Se(VI) or As(V) at an initial pH of 6.0, and the  
 152 resulting suspensions were shaken for 10 h. The solution concentrations before and after  
 153 adsorption were determined by ICP-OES. The Se(IV), Se(VI) or As(V) adsorption  
 154 isotherm of HC@LAO-IA was analyzed by fitting the data to the Langmuir (equation  
 155 (5)) and Freundlich (equation (6)) model [22, 24].

$$156 \quad C_e/q_e = C_e/Q_0 + 1/(Q_0 b) \quad (5)$$

$$157 \quad q_e = K_F C_e^{1/n} \quad (6)$$

158 where  $C_e$  (mg/L) represents equilibrium concentration;  $q_e$  (mg/g) is the adsorption  
159 capacity at equilibrium;  $Q_0$  (mg/g) represents the maximum adsorption capacity;  $b$   
160 (L/mg) is the Langmuir constant;  $K_F$  [mg/g (L/mg)<sup>1/n</sup>] is the Freundlich constant;  $n$  is a  
161 constant that reflects the energetic heterogeneity.

## 162 **2.6 Sand column experiment to study removal of Se(IV), Se(VI) or As(V) from** 163 **metallurgical wastewater by HC@LAO-IA**

164 Real industrial wastewater was collected from a metallurgical process plant.  
165 Firstly, the collected wastewater was filtered with filter paper (diameter 12.5 cm), and  
166 then the composition of the wastewater was analysed by ICP-OES and an Ion  
167 Chromatograph (IC). Silica sand (20-50 mesh, 50 g) was mixed with HC@LAO-IA (40  
168 g) and placed in a glass column (length 20 cm, diameter 2.5 cm), and its top was sealed  
169 with glass wool (5 g) to prevent sand loss. The inclusion of silica sand in the column  
170 serves the purpose of enhancing column stability, maintaining a consistent flow profile,  
171 and preventing channeling. The prepared wastewater (800 mL) was pumped towards  
172 the reactor system at the speed of 2.00 mL/min, resulting in an empty bed contact time  
173 of 26.55 minutes. The effluents were collected in the outlet at specific time points (from  
174 0 min to 540 min).

## 175 **3. Results and discussion**

### 176 **3.1 Characterization of HC@LAO-IA**

177 To elucidate the morphological and microstructural features of HC@LAO-IA, the  
178 HC@LAO-IA system was observed under SEM. Fig. 2a to 2b and Fig. 2d to 2e  
179 correspond to the surface and internal structure of HC, respectively, after hydrochloric  
180 acid treatment. The images reveal the presence of numerous pores within the HC  
181 ranging in size from 1 to 50  $\mu\text{m}$ . Through BET analysis, the specific surface area of HC  
182 was found to be 2.28  $\text{m}^2/\text{g}$  (Fig. S2). After HC was coated with LAO, LAO was firmly

183 embedded within the IA gel layer (Fig. 2c and 2f) and the surface pore structure was  
184 optimized. The specific surface area of HC@LAO-IA significantly increased to 11.32  
185 cm<sup>2</sup>/g (Fig. S2) and provided more Al-OH adsorption sites for Se(IV), Se(VI) and As(V)  
186 which was also confirmed by FTIR data presented in Fig. 2h. Simultaneously, the  
187 interconnected network of larger and smaller pores could facilitate the diffusion of  
188 Se(IV), Se(VI) and As(V) into the interior of HC, providing the possibility of contacting  
189 Se(IV), Se(VI) and As(V) with more amino groups on HC [25], further increasing the  
190 potential for Se(IV), Se(VI) and As(V) adsorption. In addition, the elemental mapping  
191 analysis (Fig. S3) suggested that the enhancement of Al signals on the outer surface of  
192 HC@LAO-IA, coupled with the concurrent reduction in the distinctive Si element  
193 signals inherent to HC, provides compelling evidence for the successful coating of HC  
194 by LAO-IA. In summary, this synergistic effect of structural modulation and surface  
195 optimization positions HC@LAO-IA as a highly efficient material for the adsorption of  
196 Se and As.

197 To investigate the composition and surface functional groups of the HC@LAO-IA  
198 composite material, XRD and FTIR analyses were performed. As shown in Fig. 2g, the  
199 XRD analysis revealed three distinct peaks at 24.1°, 26.7°, and 50.1°, corresponding to  
200 the mineral components of mullite calcium feldspar (CaAl<sub>2</sub>Si<sub>2</sub>O<sub>8</sub>) and quartz (SiO<sub>2</sub>) in  
201 HC [26, 27], as evidenced by the elemental mapping analysis in Fig. S3b and S3c.  
202 Besides, diffraction peaks at 11.1°, 21.1°, 24.8°, and 31.6°, corresponding to lattice  
203 planes of (010), (110), (120), and (200) of LAO [18, 19]. FTIR analysis of HC in Fig.  
204 2h indicated the presence of -NH<sub>2</sub> groups and SiO<sub>2</sub> in HC at 3386 and 1045 cm<sup>-1</sup>,  
205 corresponding to N-H and Si-O stretching vibrations [28, 29], respectively.  
206 Furthermore, the 890 cm<sup>-1</sup> and 964 cm<sup>-1</sup> peaks were identified as the Al-OH  
207 deformation mode vibration of LAO [19]. The FTIR analysis of HC@LAO-IA

208 demonstrated the presence of hydroxyl and amino groups ( $3147$  to  $3567\text{ cm}^{-1}$ ), COO-  
209 stretching vibration of IA ( $1436\text{ cm}^{-1}$ ) [30], Al-OH deformation mode vibration of LAO  
210 ( $896\text{ cm}^{-1}$  and  $967\text{ cm}^{-1}$ ), and Si-O stretching vibration of HC ( $1037\text{ cm}^{-1}$ ). These results  
211 provide clear evidence of the successful synthesis of HC@LAO-IA and its surface  
212 functionalization with LAO, amino and carboxyl groups.

### 213 **3.2 Adsorption by HC@LAO-IA as function of time and pH**

214 The performance of HC@LAO-IA in terms of the removal of Se(VI), Se(IV), or  
215 As(V) from aqueous solutions was investigated. As illustrated in Fig. 3a, the efficiency  
216 of HC@LAO-IA for the removal of Se(VI), Se(IV), and As(V) increased with time,  
217 with the removal rates of Se(VI) and Se(IV) being nearly equivalent and reaching  
218 equilibrium after 3 hours. The removal rate of As(V) was slightly slower than that of  
219 Se(VI) and Se(IV), reaching equilibrium after 4 hours. After equilibrium was achieved,  
220 the removal efficiency for all three contaminants reached 99.9%.

221 Figure 3b-d depicted the adsorption performance of Se(IV), Se(VI), or As(V) at  
222 various pH values. In low-concentration solutions ( $10\text{ mg/L}$ ) of these ions, their  
223 removal efficiency remained consistently high, reaching almost 99.9% over the pH  
224 range of 2-10. However, the removal efficiency of HC@LAO-IA for Se(IV), Se(VI),  
225 and As(V) dropped to around 85% at pH 11. This decrease could be due to the surplus  
226 hydroxide ions leading to the deprotonation of the functional groups on the surface of  
227 HC@LAO-IA, thereby impeding the adsorption of Se(IV), Se(VI), and As(V), while  
228 simultaneously damaging the coatings of IA and LAO [31]. The zeta potential  
229 measurements of HC@LAO-IA at various pH levels showed that the  $\text{pH}_{\text{zpc}}$  was around  
230 6 (Fig. S4). The excellent acid-base buffering property of HC@LAO-IA maintains the  
231 equilibrium pH values in a narrow range under 6, leading to a positively charged surface  
232 favoring anion sorption [45]. This buffering effect, observed despite initial pH levels

233 exceeding 6, contributed to the efficacy of HC@LAO-IA in removing Se(IV), Se(VI),  
234 and As(V) across a pH range of 3-10.

235 In high-concentration solutions (800 mg/L) of Se(IV), Se(VI), or As(V), the  
236 adsorption capacity of HC@LAO-IA for these ions exhibited a pH-dependent trend. In  
237 the case of Se(IV), the removal capacity remained above 9.5 mg/g within the pH range  
238 of 2 to 11, with a slight decrease at pH 2. This trend could be attributed to the  
239 predominance of  $\text{HSeO}_3^-$  and  $\text{SeO}_3^{2-}$  in this pH range, both carrying a net negative  
240 charge, facilitating their adsorption onto the positively charged surface of HC@LAO-  
241 IA [32]. The highest adsorption capacity was observed at pH 8, reaching 11.67 mg/g,  
242 possibly due to the formation of surface complexes between Se(IV) and functional  
243 groups of HC@LAO-IA at this pH [33].

244 The adsorption capacities of Se(VI) and As(V) exhibited a similar trend as pH varied.  
245 In the range of pH 2-6, the adsorption capacity increased initially and reached a  
246 maximum at pH 6, which was 10.88 mg/g for Se(VI) and 5.0 mg/g for As(V). From pH  
247 6-11, the adsorption capacity of Se(VI) and As(V) showed a trend of initially decreasing,  
248 then increasing, and finally decreasing again. Therein, the increase of Se(VI) and As(V)  
249 adsorption capacities at pH 2 to 6 was probably because of the protonation of surface  
250 functional groups of HC@LAO-IA, leading to a more positive surface charge. At these  
251 pH values,  $\text{HSeO}_4^{2-}$  and  $\text{H}_2\text{AsO}_4^-$  forms are predominant Se and As species in solution,  
252 favoring increased Se(VI) and As(V) adsorption capacity [32, 34, 35]. In the pH  
253 increased from 6 to 11, Se(VI) and As(V) predominantly existed in  $\text{SeO}_4^{2-}$  and  $\text{HAsO}_4^{2-}$   
254 forms respectively [36-39]. In this pH range, the surface of HC@LAO-IA became more  
255 negatively charged due to the deprotonation of functional groups, which decreased the  
256 adsorption capacity for Se(VI) and As(V), especially in pH 11. Overall, the results  
257 suggested that HC@LAO-IA is a promising material for the effective removal of Se(IV),

258 Se(VI), or As(V) from contaminated wastewater, particularly in the pH range of 3-10.

### 259 **3.3 Removal of Se(IV), Se(VI) or As(V) by HC@LAO-IA as affected by** 260 **temperature and co-existing ions**

261 To understand the performance of HC@LAO-IA in practical applications and to  
262 provide accurate parameters and conditions for its effective use, the influences of  
263 temperature (10, 20, 30 and 40 °C) and co-existing anions, including Cl<sup>-</sup>, CO<sub>3</sub><sup>2-</sup>, NO<sub>3</sub><sup>-</sup>,  
264 SO<sub>4</sub><sup>2-</sup> or PO<sub>4</sub><sup>3-</sup>, on Se(IV), Se(VI) or As(V) removal by HC@LAO-IA were investigated.  
265 As shown in Fig. 4a and 4b, there was no significant difference in the removal efficiency  
266 of Se(IV), Se(VI) or As(V) by HC@LAO-IA at different temperatures, with removal  
267 rates all varying above 99.6%. This suggested that HC@LAO-IA exhibited good  
268 stability and consistency across various temperature conditions. At co-existing ion  
269 concentrations exceeding 500 times that of Se(IV), Se(VI), or As(V), the removal  
270 efficiency of these ions was unaffected by Cl<sup>-</sup>, CO<sub>3</sub><sup>2-</sup>, NO<sub>3</sub><sup>-</sup>, or SO<sub>4</sub><sup>2-</sup> (Fig. 4c-4d).  
271 However, when PO<sub>4</sub><sup>3-</sup> concentration exceeded 500 times that of As(V), the removal  
272 efficiency of As(V) decreased to 81.3%, indicating competitive adsorption between  
273 PO<sub>4</sub><sup>3-</sup> and As(V) [40, 41]. Overall, HC@LAO-IA demonstrated high stability and  
274 consistency in removing Se(IV), Se(VI), and As(V) under different temperatures and  
275 co-existing anion conditions, making it a promising option for practical applications in  
276 the treatment of water contaminated with these toxic substances.

277 The effectiveness of four materials (HC@LAO-CA, LAO@CA, LAO@IA and  
278 HC@LAO-IA) for the removal of Se(IV), Se(VI), or As(V) from high-concentration  
279 sulfate solutions was also examined. The synthesis methods for HC@LAO-CA,  
280 LAO@CA, and LAO@IA are given in Section S3. The results (Fig. 4e and 4f) clearly  
281 demonstrated that HC@LAO-IA exhibited the best removal efficiency for Se(IV),  
282 Se(VI), and As(V), especially for Se(VI), achieving a removal rate of 98.7%.

283 HC@LAO-CA, LAO@CA and LAO@IA showed comparatively lower removal rates  
284 of 54.7%, 24.5%, and 54.9%, respectively. The higher adsorption capacity of  
285 HC@LAO-IA for Se(VI) and As(V) in high-concentration sulfate solution was  
286 attributed to the presence of amine and carboxyl functional groups on IA and HC, which  
287 conferred better surface properties and stronger interactions with Se(VI) and As(V),  
288 compared to HC@LAO-CA, LAO@CA, and LAO@IA [31, 42]. In conclusion, the  
289 impact of high concentrations of  $\text{SO}_4^{2-}$  on Se(IV), Se(VI) and As(V) removal was  
290 negligible, demonstrating that HC@LAO-IA holds great promise for effectively  
291 removing Se(IV), Se(VI) and As(V) from sulfate-rich wastewater.

### 292 **3.4 Adsorption isotherms and kinetics of HC@LAO-IA for Se(IV), Se(VI) and** 293 **As(V) adsorption**

294 The behavior of Se(IV), Se(VI) and As(V) adsorption on HC@LAO-IA was  
295 investigated by fitting the kinetic data to pseudo-first-order and pseudo-second-order  
296 models. The data could be better fitted by the pseudo-first-order kinetic model (Fig. 5a-  
297 5c), indicating that the adsorption process was predominantly chemical for Se(IV),  
298 Se(VI), and As(V), and the rate of adsorption was proportional to the concentration of  
299 the adsorbate within a certain concentration range. The IA coating on the surface of HC  
300 enhanced the stability and activity of LAO, leading to improved adsorption  
301 performance for the target pollutants. Notably, The  $k_2$  values for Se(IV), Se(VI), and  
302 As(V) were determined to be 0.026, 0.027, and 0.009 g/(mg·min) (Table S1),  
303 respectively, with Se(IV) and Se(VI) exhibiting higher adsorption rates than As(V),  
304 consistent with the trends observed in Fig. 3a. Therefore, the adsorption of Se(IV),  
305 Se(VI), and As(V) on HC@LAO-IA was controlled by a pseudo-first-order kinetic  
306 model [43].

307 Furthermore, Langmuir and Freundlich isotherm models were utilized to further

308 examine the removal mechanism of Se(IV), Se(VI) and As(V). The fitting results (Fig.  
309 6 and Table S2) showed that the adsorption isotherms of Se(IV), Se(VI), and As(V)  
310 were accurately described by the Freundlich model, with a higher  $R^2$  value of 0.99  
311 compared to that of the Langmuir model. Therefore, the adsorption of Se(IV), Se(VI),  
312 and As(V) on HC@LAO-IA was mainly heterogeneous [44]. The thermodynamic  
313 parameters for Se(IV), Se(VI) and As(V) adsorption on HC@LAO-IA, shown in Table  
314 S3, reveal negative  $\Delta G^0$  values, confirming the feasibility of Se and As adsorption on  
315 HC@LAO-IA [45]. The positive  $\Delta S^0$  values indicate the spontaneity of the adsorption  
316 process [46]. Additionally, compared to most sorbents listed in Table S4, such as Fe-  
317 based or Al-based materials, our results show excellent removal capacity of Se(IV),  
318 Se(VI), and As(V). Although iMOF exhibits better adsorption capacity than HC@LAO-  
319 IA, it faces challenges in upscaling, recycling/reuse and toxicity.

### 320 **3.5. Performance of HC@LAO-IA for Se and As removal from real metallurgical** 321 **process wastewater**

322 To investigate the potential of treating real industrial wastewaters contaminated with  
323 selenium, arsenic, nitrate, phosphate, chloride, and sulfate ions, the concentrations of  
324 these ions, pH and redox potential of real wastewater collected at a metallurgical  
325 process plant were measured. The concentration of Se and As in this wastewater was  
326 1.18 and 2.17 mg/L, respectively (Fig. 7a-7b). Therein, the pH and redox potential were  
327 determined to be 7.05 and 0.45V, respectively.  $\text{HAsO}_4^{2-}$  (As(V)) is the predominant As  
328 form under these conditions [47]. It was worth noting that the wastewater contained  
329 high concentrations of chloride and sulfate ions, 4473 and 2577 mg/L, respectively. A  
330 column experiment was subsequently conducted to determine the efficacy of  
331 HC@LAO-IA in removing As and Se from the real wastewater. As shown in Fig 7c,  
332 after 6 hours, HC@LAO-IA demonstrated a remarkable adsorption rate of 98.9% for

333 As and a slightly reduced but still impressive rate of above 92% for Se. HC@LAO-IA  
334 also demonstrates superior adsorption performance for Se and As in the real  
335 metallurgical wastewater compared to goethite and schwertmannite (Fig. S7), and it  
336 outperforms other materials used in experiments investigating the influence of  
337 coexisting ions (Fig. 4e-4f), indicating its outstanding efficacy in removing Se and As  
338 from high-concentration sulfate real wastewater. The removal performance of  
339 HC@LAO-IA in real metallurgical wastewater after regeneration using sodium  
340 chloride decreased to 82% after the fourth cycle (Fig. S8). These findings highlight the  
341 potential of HC@LAO-IA for treating real industrial wastewater contaminated with  
342 toxic ions. However, other regeneration methods still need to be explored to further  
343 improve the reusability.

### 344 **3.6. Characterization of HC@LAO-IA after adsorbing Se(IV), Se(VI) and As(V)**

345 XRD measurements were also performed to investigate the interactions among  
346 Se(IV), Se(VI), As(V) and HC@LAO-IA. Figure 8a depicts the PXRD patterns of  
347 HC@LAO-IA both before and after Se(IV), Se(VI), or As(V) adsorption. Following  
348 adsorption, the peaks corresponding to the [010] plane were observed to shift  
349 significantly from 11.1° to 8.83°, 8.68°, and 8.67°, respectively, suggesting that larger  
350 Se(IV), Se(VI) or As(V) were able to enter the interlayer of LAO via ion exchange [18,  
351 19]. Simultaneously, after treatment of simulated and real industrial wastewater, the  
352 concentrations of Fe leached from the adsorbent were all less than 0.4 mg/L (Table S5),  
353 indicating that HC@LAO-I possesses excellent stability, especially in real industrial  
354 wastewater.

355 In addition, FTIR spectroscopy was conducted to investigate the interaction  
356 between Se(IV), Se(VI) and As(V), and HC@LAO-IA. As shown in Fig 8b, after  
357 adsorption of Se(IV), Se(VI) or As(V) onto HC@LAO-IA, the smooth peak in the

358 3567-3147  $\text{cm}^{-1}$  region shifted to 3455, 3305, and 3243  $\text{cm}^{-1}$ , respectively. This is likely  
359 due to the hydrogen bonding formation between functional groups (-OH and -NH<sub>2</sub>) and  
360 adsorbed oxygenated metal anions ( $\text{SeO}_3^{2-}$ ,  $\text{SeO}_4^{2-}$  or  $\text{AsO}_4^{3-}$ ), leading to a change in  
361 the peak position [48-50]. Similarly, the peak at 1436  $\text{cm}^{-1}$  also blue-shifted to 1417,  
362 1423, and 1425  $\text{cm}^{-1}$ , which could be attributed to deprotonated -COOH groups on IA  
363 being replaced by Se(IV), Se(VI) or As(V) ions and adsorbed via electrostatic attraction  
364 [51]. Furthermore, the blueshift of the Al-OH deformation mode vibrational peak of  
365 LAO in HC@LAO-IA from 896  $\text{cm}^{-1}$  to 889, 887, and 890  $\text{cm}^{-1}$  may be due to the  
366 complexation between Al-OH and adsorbed oxygenated metal anions ( $\text{SeO}_3^{2-}$ ,  $\text{SeO}_4^{2-}$ ,  
367 or  $\text{AsO}_4^{3-}$ ), after Cl<sup>-</sup> in LAO was exchanged with these anions through ion exchange  
368 adsorption [18, 19]. Similarly, in the Raman spectra of HC@LAO-IA/Se(IV), Se(VI)  
369 or As(V) (Fig.S5), three new bands are observed at 778 and 1069  $\text{cm}^{-1}$  which can be  
370 attributed to  $\nu_1$  symmetric stretching vibration of Se-O bond in  $\text{SeO}_3^{2-}$  and  $\text{SeO}_4^{2-}$ , and  
371 asymmetric stretching vibration of As-O bond in  $\text{AsO}_4^{3-}$ , respectively [19,52]. In  
372 conclusion, the surface structure of HC@LAO-IA was found to be altered upon Se or  
373 As adsorption, as evidenced by the FTIR and Raman analysis.

374 In order to characterize the surface chemistry of HC@LAO-IA before and after the  
375 adsorption of Se(IV), Se(VI) and As(V), XPS analysis was performed. Before  
376 adsorption, the XPS survey spectrum of HC@LAO-IA in Fig 8c showed signals for Al  
377 2p, Si 2p and Fe 2p consistent with the LAO, HC and IA composition, as evidenced by  
378 the elemental mapping analysis in Fig. S3. In the high-resolution XPS spectra of the  
379 HC@LAO-IA, remarkable peaks corresponding to functional groups (-COOH and -OH  
380 on IA; -NH<sub>2</sub> on HC) were observed in Fig. S6a-c, including C=O (532.68 eV, 288.80  
381 eV), C-O (286.40 eV), O-H (531.39 eV), C-N (287.40 eV), and -NH<sub>2</sub> (399.98 eV) [48].  
382 Furthermore, characteristic peaks of Al-OH (74.4 eV) and Al-O (75.6 eV) on LAO were

383 also detected in the Al2p high-resolution spectra (Fig. S6d). Altogether, these findings  
384 provide compelling evidence for the successful immobilization of LAO and IA on HC.

385 After the adsorption of Se(IV), Se(VI) and As(V), the XPS survey spectrum in Fig.  
386 8c showed the appearance of Se and As signals, indicating the successful adsorption of  
387 these species on the HC@LAO-IA inside and surface, as evidenced by the elemental  
388 mapping analysis data in Fig. S3e to S3f and Fig. S3k to S3l. High-resolution XPS  
389 spectra of Se 3d and As 3d regions were acquired to further confirm the adsorption. The  
390 Se 3d spectrum in Fig. 8d and 8e showed two peaks at 56.19 eV and 59.61eV,  
391 corresponding to the Se 3d<sub>3/2</sub> and 3d<sub>5/2</sub> spin-orbit doublets, respectively [53]. The  
392 chemical shifts to higher binding energy suggest that the Se species were in higher  
393 oxidation states, consistent with the expected Se(IV) and Se(VI) adsorption. The As  
394 3d<sub>5/2</sub> spectrum in Fig. 8f displayed a peak at 45.38 eV, consistent with As(V) species  
395 [54,55]. The observed decrease in the Cl signal in the XPS survey spectrum after  
396 adsorption suggests the displacement of Cl<sup>-</sup> in LAO of HC@LAO-IA from the surface  
397 by Se(IV), Se(VI) and As(V) ions (Fig 8c). This displacement was attributed to the  
398 strong binding affinity of these oxyanions to the composite material surface. The shift  
399 of Al-OH and Al-O characteristic peaks towards higher binding energy, which  
400 symbolized LAO, also provided compelling evidence for the successful adsorption of  
401 some of the Se(IV), Se(VI) and As(V) ions. Additionally, the peaks of oxygen-  
402 containing groups (i.e., C=O, C-O and O-H groups) in both C 1s and O 1s spectra  
403 shifted to higher binding energies after adsorbing SeO<sub>3</sub><sup>2-</sup>, SeO<sub>4</sub><sup>2-</sup> and AsO<sub>4</sub><sup>3-</sup>. These  
404 shifts may be primarily ascribed to the decrease of electron cloud density around  
405 oxygen atoms after bonding with the Se and As species via surface complexation  
406 interactions [56]. Notably, the peak of -NH<sup>+</sup> groups (407.78 eV) in the N 1s spectrum  
407 (Fig. S6g) appeared after adsorbing Se(IV), Se(VI) and As(V) ions, primarily ascribed

408 to the decreased electron cloud intensity of nitrogen atoms after binding with Se and As  
409 species via electrostatic interactions [57, 58]. Combining the results of the above  
410 adsorption experiments, it can be concluded that hydrogen bonding, electrostatic  
411 interactions and interlayer ion exchange in LAO may play a dominant role in the  
412 removal of Se(IV), Se(VI) and As(V).

### 413 **3.7. Mechanisms involved in removal of Se(IV), Se(VI) and As(V) by HC@LAO-** 414 **IA**

415 Based on the above characterization analysis of HC@LAO-IA before and after  
416 Se(IV), Se(VI) or As(V) adsorption, it can be deduced that their removal occurs in three  
417 process steps: I. During the initial stage of sorption, Se(IV), Se(VI) or As(V) ions enter  
418 the LAO-IA layer and rapidly exchange with the free  $\text{Cl}^-$  ions present in the interlayer  
419 of LAO, and then fix onto the framework of LAO through inner-sphere complexation.  
420 II. Meanwhile, the presence of hydroxyl and deprotonated carboxyl groups on IA  
421 enables further adsorption of Se(IV), Se(VI), or As(V) through synergistic hydrogen  
422 bonding and electrostatic interactions. III. Upon entering the interior of HC, Se(IV),  
423 Se(VI), or As(V) can also be further adsorbed by the unique  $-\text{NH}_2$  groups of HC via  
424 hydrogen bonding, thereby significantly enhancing the adsorption efficiency of  
425 HC@LAO-IA for these oxyanions.

### 426 **Conclusion**

427 A novel Fe(III)-alginate and layered aluminum oxyhydroxide assisted hydro  
428 ceramsite composite (HC@LAO-IA) was prepared after HC was loaded with LAO and  
429 IA. In the pH range of 3-10, HC@LAO-IA showed excellent efficiency for Se(IV),  
430 Se(VI), or As(V) removal from contaminated wastewater. Especially, HC@LAO-IA  
431 showed high potential for removing Se and As from simulated and actual wastewater  
432 contaminated with high concentrations of sulfate ions by inner sphere complexation,

433 hydrogen bonding and electrostatic adsorption. This work provides a promising  
434 approach to remediate Se(IV), Se(VI), or As(V) contaminated wastewater, especially in  
435 presence of high concentrations of sulfate ions, which may have a potential application  
436 prospect in environmental fields.

#### 437 **Credit Author Statement**

438 Jingya Ren: Writing-Review & Editing, Supervision. Prof. Rino Morent, Prof. Nathalie  
439 De Geyter, and Dr. Karen Leus: Conducting X-ray photoelectron spectroscopy (XPS)  
440 and Scanning electron microscopy (SEM) experiments and analysis. Prof. Pascal Van  
441 Der Voort: Providing FTIR, XRD and Raman analysis. Gijs Du Laing: Formal analysis,  
442 Writing-Review & Editing, Supervision.

#### 443 **Declaration of Competing Interest**

444 The authors declare that they have no known competing financial interests or personal  
445 relationships that could have appeared to influence the work reported in this paper.

#### 446 **Acknowledgements**

447 We thank Joachim Nari for the guidance on experimental methods. J.Y.R. gratefully  
448 acknowledges the Chinese Scholarship Council (CSC) for financial support  
449 (202006500005).

450 **References**

- 451 [1] A. Deonarine, G.E. Schwartz, L.S. Ruhl, Environmental Impacts of Coal  
452 Combustion Residuals: Current Understanding and Future Perspectives, *Environ. Sci.*  
453 *Technol.* 57(5) (2023)1855-1869. <https://doi.org/10.1021/acs.est.2c06094>
- 454 [2] C. Guo, L. Hu, L. Jiang, H. Feng, B. Hu, T. Zeng, S. Song, H. Zhang, Toxic arsenic  
455 in marketed aquatic products from coastal cities in China: Occurrence, human dietary  
456 exposure risk, and coexposure risk with mercury and selenium, *Environ. Pollut.* 295  
457 (2022) 118683. <https://doi.org/10.1016/j.envpol.2021.118683>
- 458 [3] O. Ouédraogo, M. Amyot, Mercury, arsenic and selenium concentrations in water  
459 and fish from sub-Saharan semi-arid freshwater reservoirs (Burkina Faso), *Sci. Total*  
460 *Environ.* 444 (2013) 243-254. <https://doi.org/10.1016/j.scitotenv.2012.11.095>
- 461 [4] M. ALSamman, S. Sotelo, J. Sánchez, B. Rivas, Arsenic oxidation and its  
462 subsequent removal from water: An overview, *Sep. Purif. Technol.* (2022) 123055.  
463 <https://doi.org/10.1016/j.seppur.2022.123055>
- 464 [5] J. Peng, Y. Song, P. Yuan, X. Cui, G. Qiu, The remediation of heavy metals  
465 contaminated sediment, *J. Hazard. Mater.* 161(2-3) (2009) 633-640.  
466 <https://doi.org/10.1016/j.jhazmat.2008.04.061>
- 467 [6] M. Matlock, B. Howerton, D. Atwood, Chemical precipitation of heavy metals from  
468 acid mine drainage, *Water Res.* 36(19) (2002) 4757-4764.  
469 [https://doi.org/10.1016/S0043-1354\(02\)00149-5](https://doi.org/10.1016/S0043-1354(02)00149-5)
- 470 [7] L. Pincus, H. Rudel, P. Petrović, S. Gupta, P. Westerhoff, C. Muhich, J. Zimmerman,  
471 Exploring the mechanisms of selectivity for environmentally significant oxo-anion

472 removal during water treatment: a review of common competing oxo-anions and tools  
473 for quantifying selective adsorption, *Environ. Sci. Technol.* 54(16) (2020) 9769-9790.  
474 <https://doi.org/10.1021/acs.est.0c01666>

475 [8] H. Chen, C. Shepsko, A. SenGupta, Use of a Novel Bio-Nano-IX Process to Remove  
476  $\text{SeO}_4^{2-}$  or Se (VI) from Contaminated Water in the Presence of Competing Sulfate  
477 ( $\text{SO}_4^{2-}$ ), *ACS ES&T Water*, 1(8) (2021) 1859-1867.  
478 <https://doi.org/10.1021/acsestwater.1c00126>

479 [9] L. Weerasundara, Y. Ok, J. Bundschuh, Selective removal of arsenic in water: A  
480 critical review, *Environ. Pollut.* 268 (2021) 115668.  
481 <https://doi.org/10.1016/j.envpol.2020.115668>

482 [10] E. Roces, M. Muñiz-Menéndez, J. González-Galindo, J. Estaire, Lightweight  
483 expanded clay aggregate properties based on laboratory testing, *Constr. Build Mater.*  
484 313 (2021) 125486. <https://doi.org/10.1016/j.conbuildmat.2021.125486>

485 [11] A. Rashad, Lightweight expanded clay aggregate as a building material—An  
486 overview, *Constr. Build Mater.* 170 (2018) 757-775.  
487 <https://doi.org/10.1016/j.conbuildmat.2018.03.009>

488 [12] Q. Shao, Y. Zhang, Z. Liu, L. Long, Z. Liu, Y. Chen, X. Hu, M. Lu, L. Huang,  
489 Phosphorus and nitrogen recovery from wastewater by ceramsite: Adsorption  
490 mechanism, plant cultivation and sustainability analysis, *Sci. Total Environ.* 805 (2022)  
491 150288. <https://doi.org/10.1016/j.scitotenv.2021.150288>

492 [13] J. Wang, Y. Zhao, P. Zhang, L. Yang, G. Xi, Adsorption characteristics of a novel  
493 ceramsite for heavy metal removal from stormwater runoff, *Chin. J. Chem. Eng.* 26(1)

494 (2018) 96-103. <https://doi.org/10.1016/j.cjche.2017.04.011>

495 [14] J. Qi, Z. Zhu, Z. Chen, X. Wang, Y. Zhang, H. Chen, Synthesis, Characterization  
496 and application of dewatered municipal sludge-based ceramsite and its phosphorus  
497 adsorption characteristics, *J. Clean. Prod.* 391 (2023) 136216.  
498 <https://doi.org/10.1016/j.jclepro.2023.136216>

499 [15] S. Jiang, P. Rao, H. Huan, L. Liu, Y. Luo, W. Wang, X. Du, H. Yang, J. Tian,  
500 Performance of a ceramsite-enhanced gravity-driven ceramic membrane (GDCM) for  
501 simultaneous manganese ion and ammonia removal, *J. Clean. Prod.* (2023) 136082.  
502 <https://doi.org/10.1016/j.jclepro.2023.136082>

503 [16] Y. Chen, J. Shi, H. Ron, X. Zhou, F. Chen, X. Li, T. Wang, H. Hou, Adsorption  
504 mechanism of lead ions on porous ceramsite prepared by co-combustion ash of sewage  
505 sludge and biomass, *Sci. Total Environ.* 702 (2020) 135017.  
506 <https://doi.org/10.1016/j.scitotenv.2019.135017>

507 [17] J. Niu, P. Ding, X. Jia, G. Hu, Z. Li, Study of the properties and mechanism of  
508 deep reduction and efficient adsorption of Cr (VI) by low-cost Fe<sub>3</sub>O<sub>4</sub>-modified  
509 ceramsite, *Sci. Total Environ.* 688 (2019) 994-1004.  
510 <https://doi.org/10.1016/j.scitotenv.2019.06.333>

511 [18] P. Bai, Z. Dong, S. Wang, X. Wang, Y. Li, Y. Wang, Y. Ma, W. Yan, X. Zou, J. Yu,  
512 A layered cationic aluminum oxyhydroxide as a highly efficient and selective trap for  
513 heavy metal oxyanions, *Angew. Chem. Int. Ed. Engl.* 59(44) (2020) 19539-19544.  
514 <https://doi.org/10.1002/anie.202005878>

515 [19] H. Cai, H. Bao, X. Zhang, L. Lei, C. Xiao, Highly efficient sorption of selenate

516 and selenite onto a cationic layered single hydroxide via anion exchange and inner-  
517 sphere complexation, Chem. Eng. J. 420 (2021) 129726.  
518 <https://doi.org/10.1016/j.cej.2021.129726>

519 [20] J. Wu, H. Zhen, F. Zhang, R. Zeng, B. Xin, Iron-carbon composite from  
520 carbonization of iron-crosslinked sodium alginate for Cr (VI) removal, Chem. Eng. J.  
521 362 (2019) 21-29. <https://doi.org/10.1016/j.cej.2019.01.009>

522 [21] H. Vu, A. Dwivedi, T. Le, S. Seo, E. Kim, Y. Chang, Magnetite graphene oxide  
523 encapsulated in alginate beads for enhanced adsorption of Cr (VI) and As (V) from  
524 aqueous solutions: Role of crosslinking metal cations in pH control, Chem. Eng. J. 307  
525 (2017) 220-229. <https://doi.org/10.1016/j.cej.2016.08.058>

526 [22] R. Ezzati, Derivation of pseudo-first-order, pseudo-second-order and modified  
527 pseudo-first-order rate equations from Langmuir and Freundlich isotherms for  
528 adsorption, Chem. Eng. J. 392 (2020) 123705. <https://doi.org/10.1016/j.cej.2019.123705>

529 [23] C. Yan, Z. Cheng, Y. Tia, F. Qiu, H. Chang, S. Li, Y. Cai, X. Quan, Adsorption of  
530 Ni (II) on detoxified chromite ore processing residue using citrus peel as reductive  
531 mediator: Adsorbent preparation, kinetics, isotherm, and thermodynamics analysis, J.  
532 Clean. Prod. 315 (2021) 128209. <https://doi.org/10.1016/j.jclepro.2021.128209>

533 [24] V. Ramya, D. Murugan, C. Lajapathirai, S. Meenatchisundaram, S. Arumugam, A  
534 composite adsorbent of superparamagnetic nanoparticles with sludge biomass derived  
535 activated carbon for the removal of chromium (VI), J. Clean. Prod. 366 (2022) 132853.  
536 <https://doi.org/10.1016/j.jclepro.2022.132853>

537 [25] M. Chaudhary, S. Rawat, A. Maiti, Defluoridation by bare nanoadsorbents,

538 nanocomposites, and nanoadsorbent loaded mixed matrix membranes, *Sep. Purif. Rev.*  
539 52(2) (2023) 135-153. <https://doi.org/10.1080/15422119.2022.2045610>

540 [26] C. Ou, J. Wang, W. Yang, Y. Bao, Z. Liao, J. Shi, J. Qin, Removal of ammonia  
541 nitrogen and phosphorus by porous slow-release  $\text{Ca}^{2+}$  ceramsite prepared from  
542 industrial solid wastes, *Sep. Purif. Technol.* 304 (2023) 122366.  
543 <https://doi.org/10.1016/j.seppur.2022.122366>

544 [27] N. Li, B. Lai, L. Ding, J. Li, C. Liu, L. Wu, Synchronous algae and phosphorus  
545 removal by Ceramsite@  $\text{Fe}_2\text{O}_3$  (FC) via taking the algae as crystal nuclei of  
546 hydroxylapatite, *Chem. Eng. J.* 426 (2021) 130748.  
547 <https://doi.org/10.1016/j.cej.2021.130748>

548 [28] Y. Chen, N. Wang, S. An, C. Cai, J. Peng, M. Xie, J. Peng, X. Song, Synthesis of  
549 novel hierarchical porous zeolitization ceramsite from industrial waste as efficient  
550 adsorbent for separation of ammonia nitrogen, *Sep. Purif. Technol.* 297 (2022) 121418.  
551 <https://doi.org/10.1016/j.seppur.2022.121418>

552 [29] Y. Shi, W. Guo, Y. Jia, C. Xue, Y. Qiu, Q. Zhao, D. Wang, Preparation of non-  
553 sintered lightweight aggregate ceramsite based on red mud-carbide slag-fly ash:  
554 Strength and curing method optimization, *J. Clean. Prod.* 372 (2022) 133788.  
555 <https://doi.org/10.1016/j.jclepro.2022.133788>

556 [30] B. Swamy, Y. Yun, In vitro release of metformin from iron (III) cross-linked  
557 alginate–carboxymethyl cellulose hydrogel beads, *Int. J. Biol. Macromol.* 77 (2015)  
558 114-119. <https://doi.org/10.1016/j.ijbiomac.2015.03.019>

559 [31] D. Roquero, A. Othman, A. Melman, E. Katz, Iron (III)-cross-linked alginate

560 hydrogels: A critical review, *Adv. Mater.* 3(4) (2022) 1849-1873.  
561 <https://doi.org/10.1039/D1MA00959A>

562 [32] J. Tong, J. Yang, L. Zhang, T. Liu, C. Peng, X. Ni, T. Dong, P. Mocilac, K. Shi, X.  
563 Hou, Efficient removal of Se-79 from highly acidic solution using SiO<sub>2</sub> particles  
564 functionalised with iron hydroxide, *Chem. Eng. J.* 446 (2022) 137387.  
565 <https://doi.org/10.1016/j.cej.2022.137387>

566 [33] L. Zhang, S. Jiang, Y. Guan, Efficient removal of selenate in water by cationic poly  
567 (allyltrimethylammonium) grafted chitosan and biochar composite, *Environ. Res.* 194  
568 (2021) 110667. <https://doi.org/10.1016/j.envres.2020.110667>

569 [34] K. Kalaitzidou, A. Nikolettopoulos, N. Tsiftsakis, F. Pinakidou, M. Mitrakas,  
570 Adsorption of Se (IV) and Se (VI) species by iron oxy-hydroxides: Effect of positive  
571 surface charge density, *Sci. Total Environ.* 687 (2019) 1197-1206.  
572 <https://doi.org/10.1016/j.scitotenv.2019.06.174>

573 [35] Y. Wu, C. Guan, N. Griswold, L. Hou, X. Fang, A. Hu, Z. Hu, C. Yu, Zero-valent  
574 iron-based technologies for removal of heavy metal (loid) s and organic pollutants from  
575 the aquatic environment: Recent advances and perspectives, *J. Clean. Prod.* 277 (2020)  
576 123478. <https://doi.org/10.1016/j.jclepro.2020.123478>

577 [36] M. Li, C. Wang, M. O'Connell, C. Chan, Carbon nanosphere adsorbents for  
578 removal of arsenate and selenate from water, *Environ. Sci. Nano*, 2(3) (2015) 245-250.  
579 <https://doi.org/10.1039/C4EN00204K>

580 [37] C. Tabelin, R. Sasaki, T. Igarashi, I. Park, S. Tamoto, T. Arima, M. Ito, N. Hiroyoshi,  
581 Simultaneous leaching of arsenite, arsenate, selenite and selenate, and their migration

582 in tunnel-excavated sedimentary rocks: II. Kinetic and reactive transport modeling,  
583 Chemosphere, 188 (2017) 444-454. <https://doi.org/10.1016/j.chemosphere.2017.08.088>

584 [38] J. Yamani, A. Lounsbury, J. Zimmerman, Towards a selective adsorbent for  
585 arsenate and selenite in the presence of phosphate: Assessment of adsorption efficiency,  
586 mechanism, and binary separation factors of the chitosan-copper complex, Water Res.  
587 88 (2016) 889-896. <https://doi.org/10.1016/j.watres.2015.11.017>

588 [39] L. Camacho, R. Parra, S. Deng, Arsenic removal from groundwater by  $\text{MnO}_2^-$   
589 modified natural clinoptilolite zeolite: Effects of pH and initial feed concentration, J.  
590 Hazard. Mater. 189(1-2) (2011) 286-293. <https://doi.org/10.1016/j.jhazmat.2011.02.035>

591 [40] Y. Han, J. Park, Y. Min, D. Lim, Competitive adsorption between phosphate and  
592 arsenic in soil containing iron sulfide: XAS experiment and DFT calculation  
593 approaches, Chem. Eng. J. 397 (2020) 125426. <https://doi.org/10.1016/j.cej.2020.125426>

594 [41] X. Li, L. Yan, W. Zhong, M. Kersten, C. Jing, Competitive arsenate and phosphate  
595 adsorption on  $\alpha\text{-FeOOH}$ ,  $\text{LaOOH}$ , and nano- $\text{TiO}_2$ : Two-dimensional correlation  
596 spectroscopy study, J. Hazard. Mater. 414 (2021) 125512.  
597 <https://doi.org/10.1016/j.jhazmat.2021.125512>

598 [42] S. Anita, F. Ardiati, M. Oktaviani, F. Sari, O. Nurhayat, K. Ramadhan, D. Yanto,  
599 Immobilization of laccase from *Trametes hirsuta* EDN 082 in light expanded clay  
600 aggregate for decolorization of Remazol Brilliant Blue R dye, Bioresour. Technol. Rep.  
601 12 (2020) 100602. <https://doi.org/10.1016/j.biteb.2020.100602>

602 [43] O. Kazak, A. Tor, In situ preparation of magnetic hydrochar by co-hydrothermal  
603 treatment of waste vinasse with red mud and its adsorption property for Pb (II) in

604 aqueous solution, *J. Hazard Mater.* 393 (2020) 122391.  
605 <https://doi.org/10.1016/j.jhazmat.2020.122391>

606 [44] I. Mohammed, A. Jawad, A. Abdulhameed, M. Mastuli, Physicochemical  
607 modification of chitosan with fly ash and tripolyphosphate for removal of reactive red  
608 120 dye: statistical optimization and mechanism study, *Int. J. Biol. Macromol.* 161  
609 (2020) 503-513. <https://doi.org/10.1016/j.ijbiomac.2020.06.069>

610 [45] N. Jain, A. Maiti, Fe-Mn-Al metal oxides/oxyhydroxides as As (III) oxidant under  
611 visible light and adsorption of total arsenic in the groundwater environment, *Sep. Purif.*  
612 *Technol.* 302 (2022) 122170. <https://doi.org/10.1016/j.seppur.2022.122170>

613 [46] N. Jain, A. Maiti, Arsenite Oxidation and Arsenic Adsorption Strategy Using  
614 Developed Material from Laterite and Ferromanganese Slag: Electron Paramagnetic  
615 Resonance Spectroscopy Analysis, *Ind. Eng. Chem. Res.* 62(38) (2023) 15600-15612.  
616 <https://doi.org/10.1021/acs.iecr.3c01377>

617 [47] M. Charykova, V. Krivovichev, Mineral systems and thermodynamic stability of  
618 arsenic minerals in the environment, In *Processes and Phenomena on the Boundary*  
619 *Between Biogenic and Abiogenic Nature*, (2020) 259-276. [https://doi.org/10.1007/978-](https://doi.org/10.1007/978-3-030-21614-6_15)  
620 [3-030-21614-6\\_15](https://doi.org/10.1007/978-3-030-21614-6_15)

621 [48] F. Ahmadpoor, S. Shojaosadati, S. Mousavi, Magnetic silica coated iron  
622 carbide/alginate beads: synthesis and application for adsorption of Cu (II) from aqueous  
623 solutions, *Int. J. Biol. Macromol.* 128 (2019) 941-947.  
624 <https://doi.org/10.1016/j.ijbiomac.2019.01.173>

625 [49] Y. Zhang, X. Xu, C. Yue, L. Song, Y. Lv, F. Liu, A. Li, Insight into the efficient co-

626 removal of Cr (VI) and Cr (III) by positively charged UiO-66-NH<sub>2</sub> decorated  
627 ultrafiltration membrane, Chem. Eng. J. 404 (2021) 126546.  
628 <https://doi.org/10.1016/j.cej.2020.126546>

629 [50] Z. Lu, J. Yu, H. Zeng, Q. Liu, Polyamine-modified magnetic graphene oxide  
630 nanocomposite for enhanced selenium removal, Sep. Purif. Technol. 183 (2017) 249-  
631 257. <https://doi.org/10.1016/j.seppur.2017.04.010>

632 [51] J. Wu, R. Zeng, In situ preparation of stabilized iron sulfide nanoparticle-  
633 impregnated alginate composite for selenite remediation, Environ. Sci. Technol. 52(11)  
634 (2018) 6487-6496. <https://doi.org/10.1021/acs.est.7b05861>

635 [52] Y. Penke, A. Yadav, I. Malik, A. Tyagi, J. Ramkumar, K. Kar, Insights of arsenic  
636 (III/V) adsorption and electrosorption mechanism onto multi synergistic (redox-  
637 photoelectrochemical-ROS) aluminum substituted copper ferrite impregnated rGO,  
638 Chemosphere, 267 (2021) 129246. <https://doi.org/10.1016/j.chemosphere.2020.129246>

639 [53] D. Han, B. Batchelor, A. Abdel-Wahab, XPS analysis of sorption of selenium (IV)  
640 and selenium (VI) to mackinawite (FeS), Environ. Prog. Sustain. Energy, 32(1) (2013)  
641 84-93. <https://doi.org/10.1002/ep.10609>

642 [54] H. Viltres, O. Odio, L. Lartundo-Rojas, E. Reguera, Degradation study of arsenic  
643 oxides under XPS measurements, Appl. Surf. Sci. 511 (2020) 145606.  
644 <https://doi.org/10.1016/j.apsusc.2020.145606>

645 [55] N. Jain, A. Maiti, Arsenic adsorbent derived from the ferromanganese slag,  
646 Environ. Sci. Pollut. Res. 28 (2021) 3230-3242. [https://doi.org/10.1007/s11356-020-](https://doi.org/10.1007/s11356-020-10745-9)  
647 10745-9

- 648 [56] A. Hassan, A. Abdel-Mohsen, H. Elhadidy, Adsorption of arsenic by activated  
649 carbon, calcium alginate and their composite beads, *Int. J. Biol. Macromol.* 68 (2014)  
650 125-130. <https://doi.org/10.1016/j.ijbiomac.2014.04.006>
- 651 [57] B. Rathi, P. Kumar, A review on sources, Identification and treatment strategies for  
652 the removal of toxic Arsenic from water system, *J. Hazard Mater.* 418 (2021) 126299.  
653 <https://doi.org/10.1016/j.jhazmat.2021.126299>
- 654 [58] P. Fan, Y. Sun, B. Zhou, X. Guan, Coupled effect of sulfidation and ferrous dosing  
655 on selenate removal by zerovalent iron under aerobic conditions, *Environ. Sci. Technol.*  
656 53(24) (2019) 14577-14585. <https://doi.org/10.1021/acs.est.9b04956>

657 **Figure captions:**

658 **Figure 1.** Schematic illustration of the fabrication process of HC@LAO-IA

659 **Figure 2.** SEM images of (a, b) HC surface and (d, e) HC interior, and (c, f) LAO-IA;  
660 FTIR spectra (g) and XRD patterns (h) of HC@LAO-IA.

661 **Figure 3.** (a) Influence of time on removal efficiency of Se(IV), Se(VI) and As(V) by  
662 HC@LAO-IA at 25°C (initial concentration 10 mg/L); (b,c,d) Influence of pH on  
663 removal efficiency and adsorption capacity of Se(IV) (b), Se(VI) (c) and As(V) (d) by  
664 HC@LAO-IA at 25°C (initial concentrations 10 mg/L and 800 mg/L).

665 **Figure 4.** Influence of (a and b) temperature at pH 6 on Se(IV), Se(VI), and As(V)  
666 removal efficiency of HC@LAO-IA; (c) Effect of co-existing anions on Se(IV), Se(VI),  
667 and As(V) (d) removal efficiency of HC@LAO-IA. The effectiveness of four materials  
668 in removing (e) Se(IV), Se(VI), or (f) As(V) from high-concentration sulfate solutions  
669 (initial concentrations: 1 mg/L of Se(IV), Se(VI), or As(V) and 500 mg/L of SO<sub>4</sub><sup>2-</sup>).

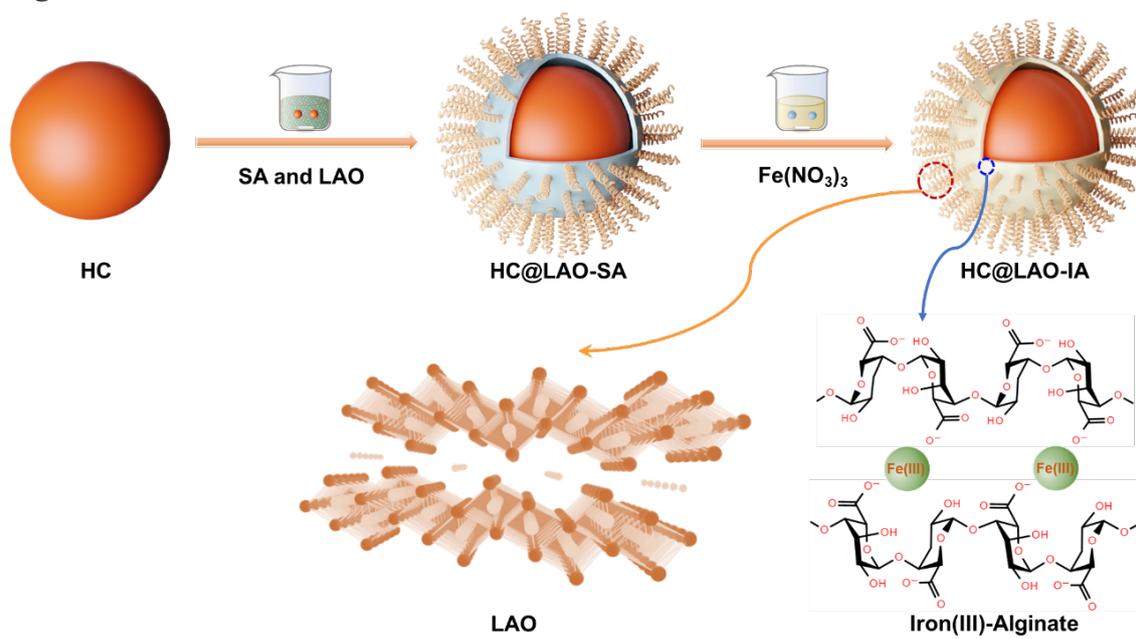
670 **Figure 5.** Adsorption kinetics plots for adsorption of (a) Se(IV), (b) Se(VI) and (c) As(V)  
671 onto HC@LAO-IA, respectively (C<sub>0</sub> of Se(IV), Se(VI) and As(V): 10 mg/L each, pH  
672 6.0, 25 °C, 200 rpm).

673 **Figure 6.** Equilibrium isotherms plots for adsorption of (a) Se(IV), (b) Se(VI), and (c)  
674 As(V) onto HC@LAO-IA (C<sub>0</sub> of Se(IV), Se(VI) and As(V): 200 mg/L-1500 mg/L each,  
675 pH 6.0, 25 °C, 200 rpm).

676 **Figure 7.** (a-b) The concentration of various anions in actual metallurgical process  
677 wastewater; (c) Column experiment of HC@LAO-IA in actual metallurgical process  
678 wastewater.

679 **Figure 8.** (a) XRD patterns, (b) FTIR spectra, (c) full-range XPS spectra and (d-f)  
680 Se3d/As3d XPS spectra of HC@LAO-IA before and after Se(IV), Se(VI), or As(V)  
681 adsorption.

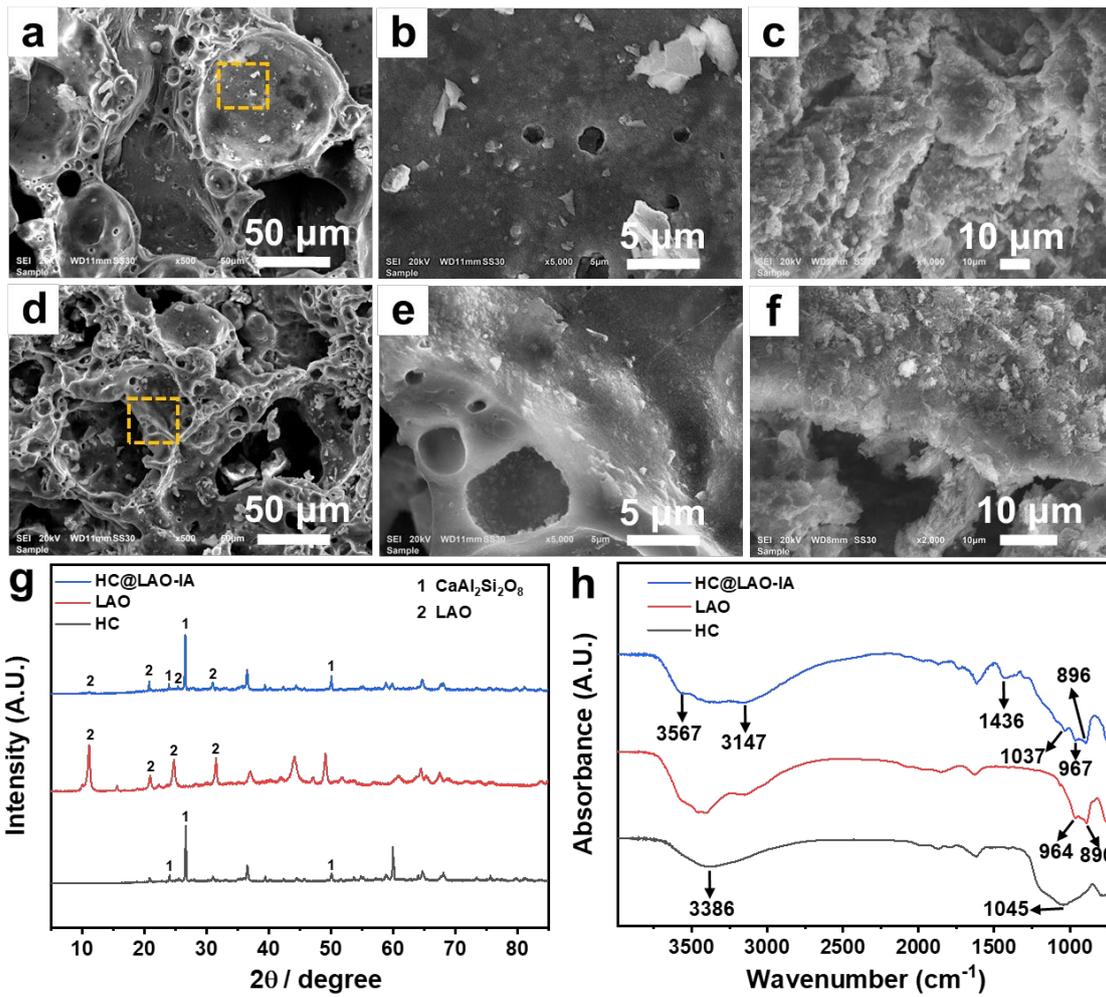
682 **Figures**



683

684

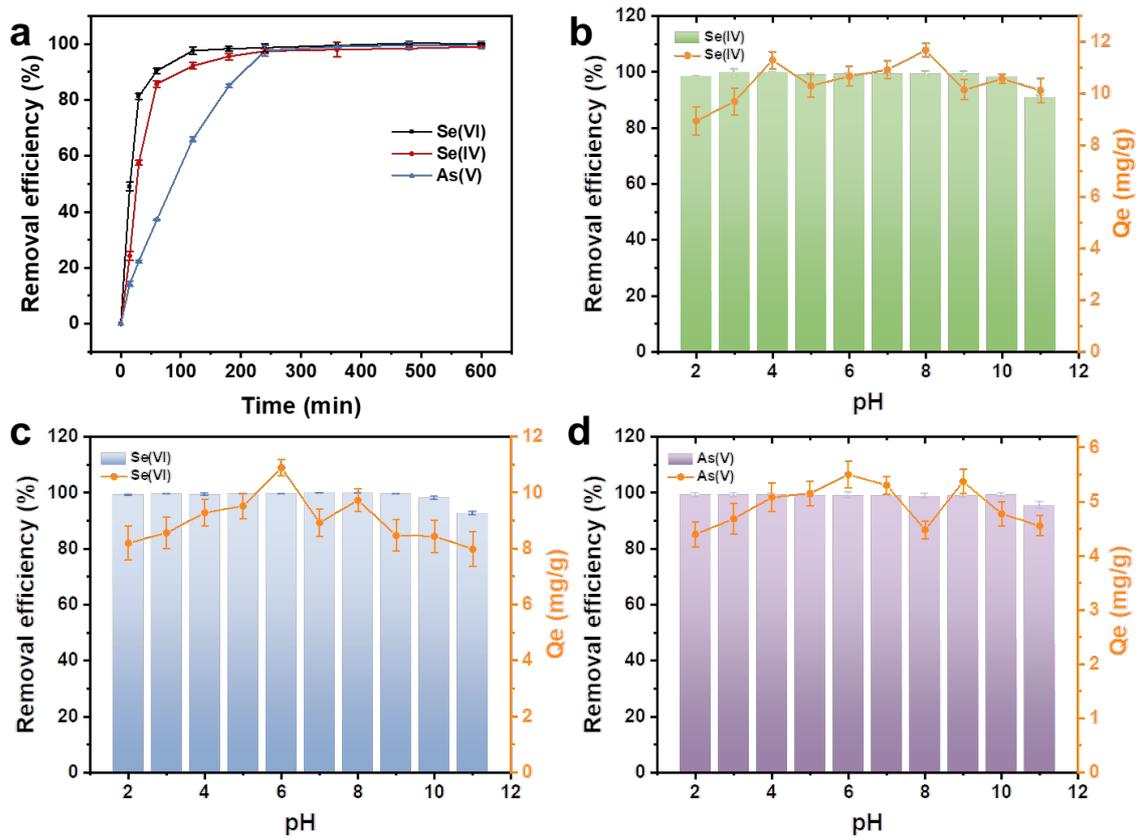
Figure 1



685

686

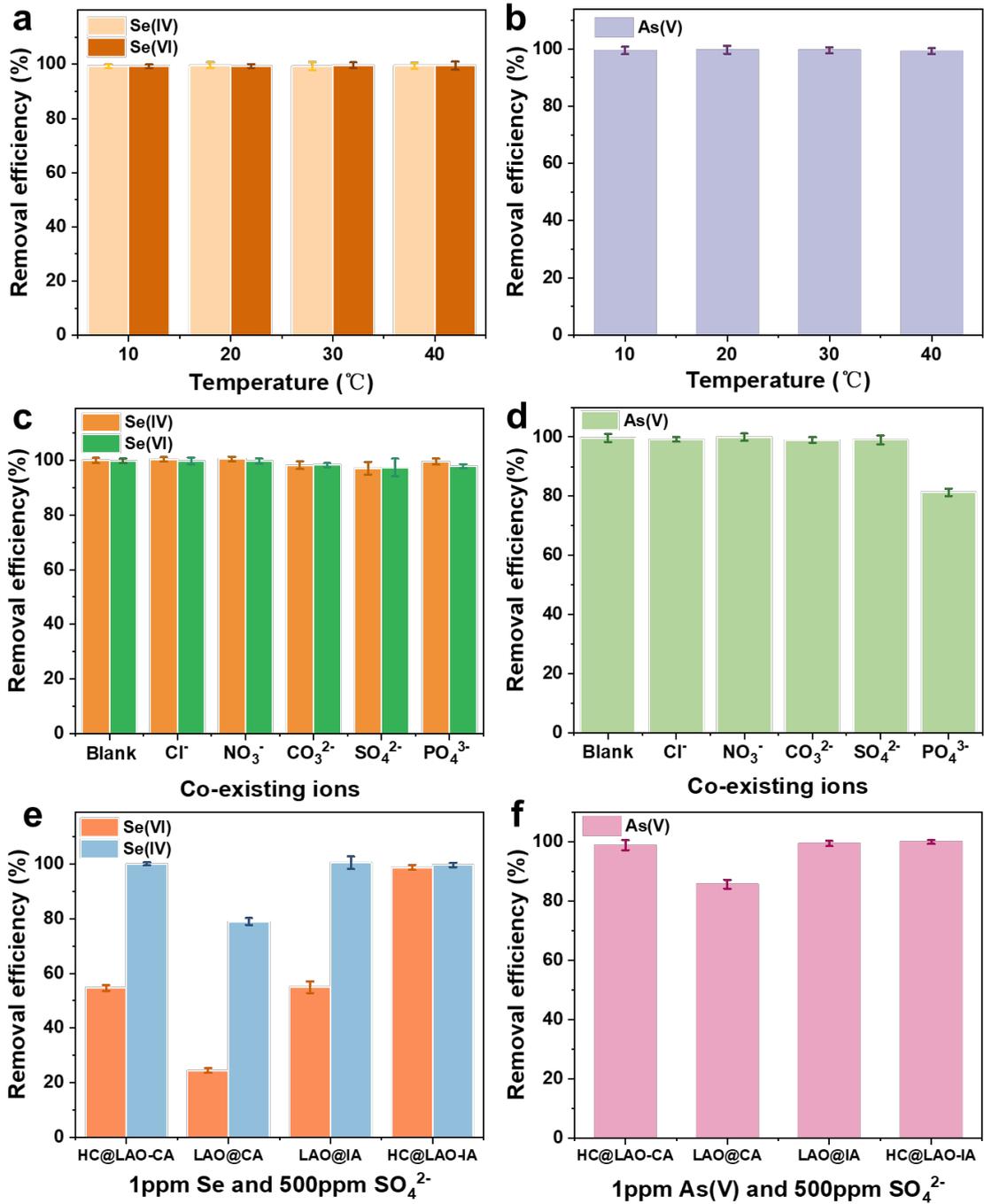
Figure 2



687

688

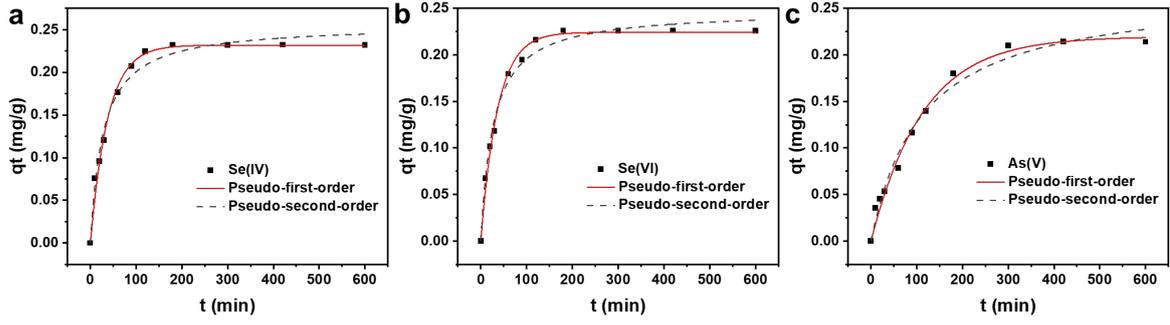
Figure 3



689

690

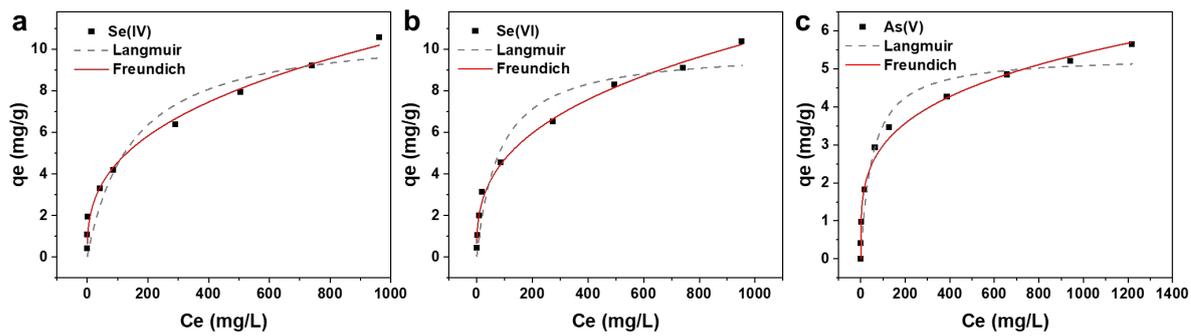
Figure 4



691

692

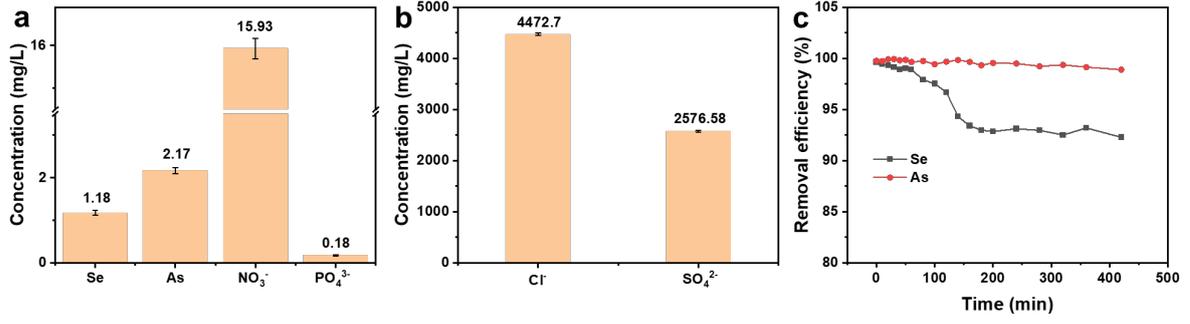
Figure 5



693

694

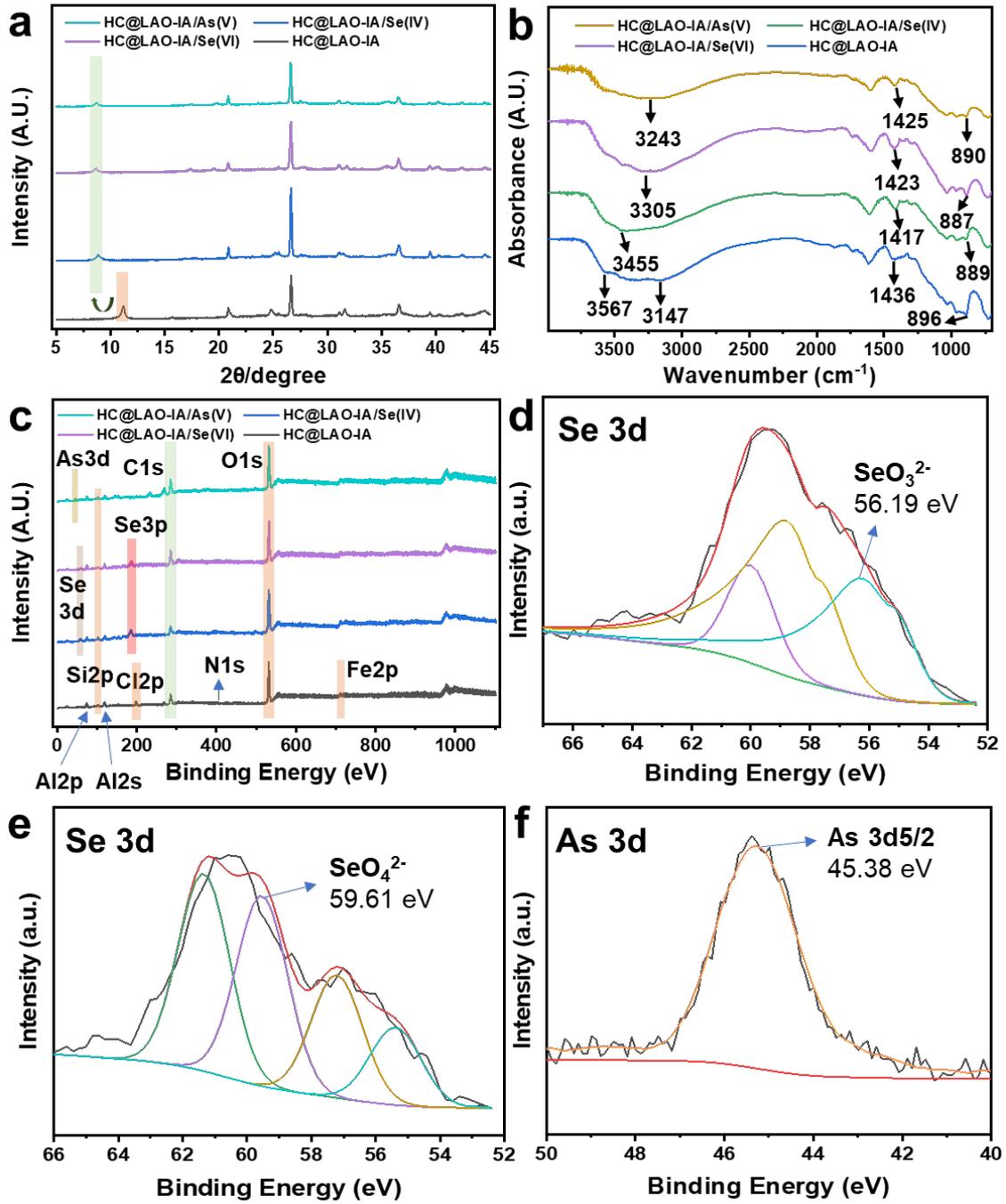
Figure 6



695

696

Figure 7



697

698

Figure 8