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2 Life cycle cost analysis of agri-food products: A systematic review

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11

12 Abstract

13 Because of the increasing challenges the global food system is facing on a social, economic and
14 environmental level, and the need to meet the United Nations Sustainable Development Goals (SDGs)
15 by 2030, agri-food systems are increasingly required to become more sustainable. Life cycle tools, such
16 as a life cycle assessment (LCA) and life cycle cost analysis (LCC) to evaluate the environmental and
17 economic performance respectively, play an important role in sustainability research. Contrary to LCA,
18 the LCC methodology is not standardized for agri-food products. This study aims to obtain insights into
19 the use of LCC in the agri-food sector using a systematic review approach. Data related to the
20 methodology and findings of life cycle cost analyses of agri-food products were extracted from 92
21 articles, covering a wide range of products (crops: 59, food/drinks: 22, other: 11) and purposes.
22 Currently, there is no consensus about LCC type definitions and the definition of different types of
23 system boundaries amongst researchers. Furthermore, these and other methodological choices are
24 often not reported in the analyzed studies. The data collection itself can also differ across studies,

25 especially with regards to the inclusion of different cost categories. It is important to include each cost
26 category since all categories have been identified as a costs hotspot in our list of studies (inputs: 84%,
27 labor: 62%, machinery: 27%, other: 39%). Standardizing the LCC methodology is recommended to
28 ensure comparability and enhance the scientific impact of studies. Integrating LCC results with findings
29 from other life cycle tools, as done in 29 studies, can further support decision-making. The most
30 common methods for integrating results are eco-efficiency analysis and multi-criteria decision analysis
31 methods. In conclusion, it is clear that LCC is a very valuable tool, as a method on its own or
32 complemented by other life cycle tools.

33

34 **Keywords:** life cycle costing, economic sustainability, sustainability assessment, life cycle thinking,
35 sustainable agri-food systems, systematic review

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37

38 1. Introduction

39 There is a need for agri-food systems to become more sustainable on a social, economic and
40 environmental level. Considering the population growth projections and consequently the increase in
41 food demand, reaching sustainability will be even more challenging in the future Gladek et al. (2016;
42 (Godfray et al., 2010). Poverty, inequality, hunger and malnutrition, resource scarcity, ecosystem
43 degradation and climate change are the result of our unsustainable food system (FAO, 2014). To tackle
44 these challenges and meet the United Nations Sustainable Development Goals (SDGs) by 2030, the
45 transformation of the agri-food sector will be crucial (Djekic et al., 2021). Therefore, the ability to
46 measure and compare the sustainability of different food and processing options is important (Darton,
47 2015). To assess sustainability, multiple tools have been developed, including life cycle tools, which
48 seem to play an important role in sustainability research (Notarnicola et al., 2017; Sala et al., 2013a;
49 Sala et al., 2013b). Life cycle assessment (LCA), for example, is a well-established method to assess
50 environmental sustainability over the entire product's life cycle in a wide range of different sectors
51 (Guinée et al., 2011). The LCA tool has been used extensively to assess the environmental sustainability
52 of agricultural products, food products, food processing technologies, food waste and bioenergy (Gava
53 et al., 2019; Notarnicola et al., 2017; Omolayo et al., 2021; Roos and Ahlgren, 2018; Silva and Sanjuán,
54 2019). Thereby, different impact categories are often distinguished, e.g. global warming potential, land
55 use, acidification, eutrophication,... (Meier et al., 2015).

56 More recently, researchers have started to include economic sustainability in the life cycle thinking
57 framework (Norris, 2001), using life cycle cost analysis (LCC), also known as life cycle costing. This tool
58 can be used for decision-making or for identifying economic hotspots to potentially decrease the
59 product's life cycle costs (Hunkeler et al., 2008). The Society of Environmental Toxicology and
60 Chemistry (SETAC) distinguishes three different types of life cycle cost analyses: conventional LCC,
61 environmental LCC and societal LCC (Hunkeler et al., 2008). In contrast to conventional LCC, where
62 typically only the direct costs covered by one actor are taken into account, the environmental LCC

63 considers all costs over the product's life cycle for one or more actors, including the costs of
64 internalized or soon-to-be-internalized externalities (Hoogmartens et al., 2014; Hunkeler et al., 2008).
65 Societal LCC takes all costs, including externalities, for all actors that are directly or indirectly affected
66 by the production, into account (Hoogmartens et al., 2014; Hunkeler et al., 2008). Environmental LCC
67 can be used in combination with LCA, if the same system boundaries and functional unit is used
68 (Hunkeler et al., 2008; Kloepffer, 2008; Swarr et al., 2011). While the LCA methodology has been
69 standardized and can be applied to any product (ISO, 2006a; ISO, 2006b), the LCC methodology has
70 only been standardized for petroleum, gas, buildings and constructed assets (ISO, 2008; ISO 2021).
71 Currently, there is no international standard available for the life cycle cost analysis of agricultural and
72 food products. Recent reviews of life cycle costing of food waste management and life cycle costing in
73 urban agriculture have found inconsistencies within the methodological aspects of LCC (De Menna et
74 al., 2018; Peña and Rovira-Val, 2020). In conclusion, insights into the methodological framework for
75 LCC in the agri-food sector are needed.

76 Even though the - rather standardized - use of LCA for agri-food systems still faces some challenges
77 (e.g. comparability of studies, LCA does not capture all aspects necessary for sustainability,...) (Baldini
78 et al., 2017; Notarnicola et al., 2017; Schau and Fet, 2008), it is well-established, unlike the use of LCC
79 in agri-food systems. To address this gap, a systematic review of life cycle cost analyses of agri-food
80 products will be conducted. This paper aims to identify the current issues related to the LCC
81 methodology and to get a better understanding of the methodological framework of life cycle costing
82 in the agri-food sector. Furthermore, methodological suggestions for future LCC studies will be
83 formulated. To our knowledge, such a systematic review of life cycle cost analyses in the agri-food
84 sector does not exist, except for the aforementioned reviews on two specific agri-food domains, i.e.
85 food waste (De Menna et al., 2018) and urban agriculture (Peña and Rovira-Val, 2020). This systematic
86 review will further differ from existing reviews in the fact that it (1) uses a different methodological
87 approach (systematic) and (2) includes all studies conducting a life cycle cost analysis of food or
88 agricultural products instead of focusing on food waste or urban agriculture only. The next section of

89 the paper explains the methodology that is used for this systematic review. The third section focuses
90 on the results and compares them to existing literature. Finally, the last section will contain the
91 conclusions.

92 2. Material and methods

93 2.1. Article selection

94 The database Scopus and Web of Science were used to search for relevant articles in June 2022. The
95 syntax that was used to search for relevant literature consisted of two parts. The first part of the syntax
96 had to identify all life cycle costing studies and therefore consisted of LCC and all its synonyms: “Life
97 Cycle Cost*” OR “LCC” OR “production cost* analysis” OR (economic AND (“life cycle sustainability
98 assessment” OR LCSA)). The second part of the syntax had to identify all agri-food related studies,
99 including beer and other beverages: food OR feed OR agri* OR agro* OR crop* OR farm* OR livestock
100 OR beer OR beverage*. The word “beer” was added to the syntax to include a highly cited paper
101 (Amienyo and Azapagic, 2016) that was found during a preliminary search to define the study’s scope.
102 Since the aim of this systematic review is to get insights into the methodological framework of life cycle
103 cost analysis in the agri-food sector, both parts of the syntax had to be present for an article to be
104 selected. The search resulted in 1347 papers from Scopus and 1882 papers from Web of Science, which
105 were later transferred into one database using EndNote software.

106 2.2. Article screening

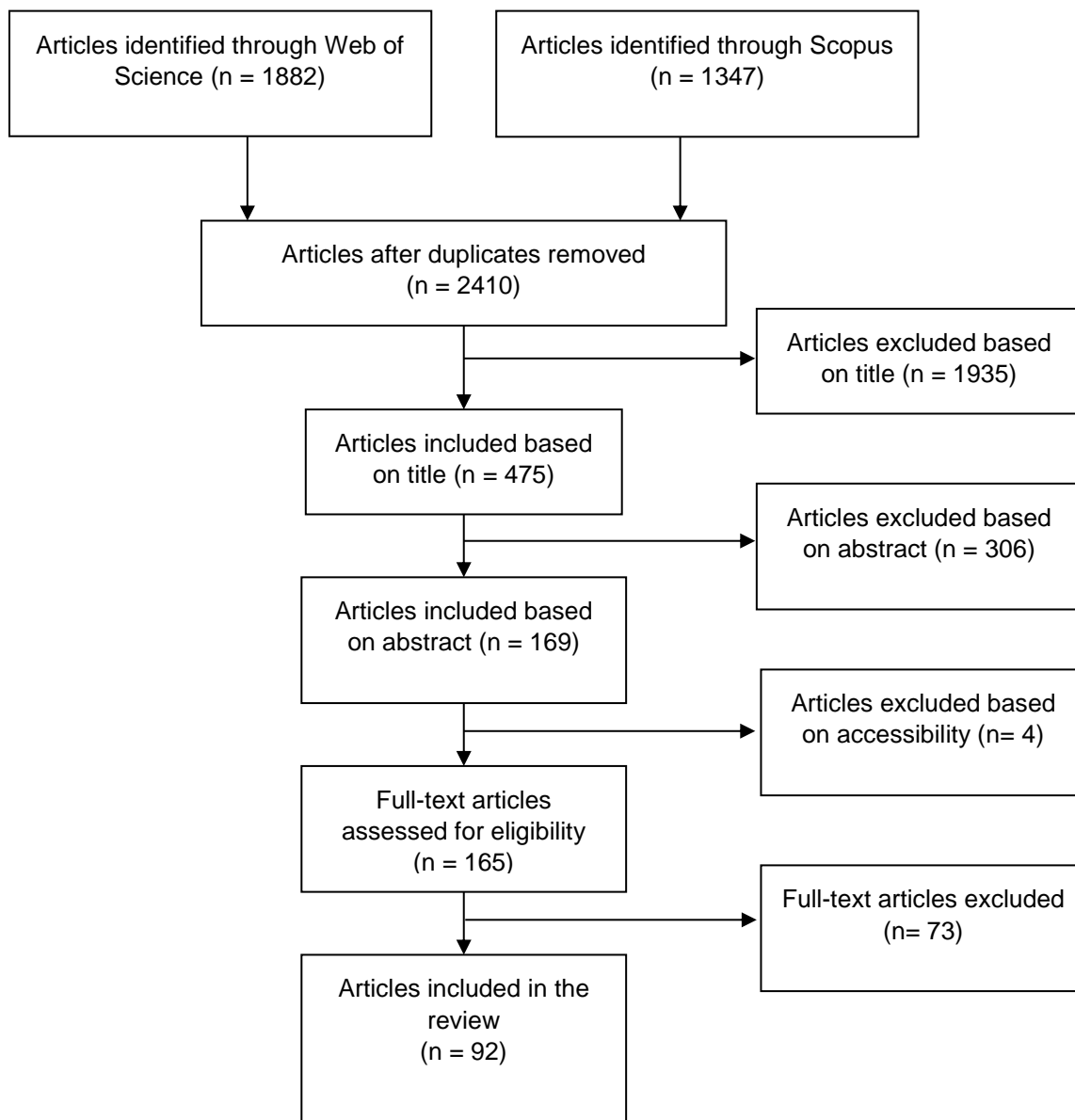
107 To identify all relevant articles, specific inclusion criteria were used. All peer-reviewed articles
108 conducting a life cycle cost analysis in the agri-food sector, either as an independent study, or in
109 conjunction with a life cycle assessment, or as part of a life cycle sustainability assessment, were
110 included. Furthermore, the article had to be written in English, accessible, the life cycle costing method
111 had to be explicitly mentioned and the article had to cover the production of food, feed, bio-energy or
112 other agricultural products. Articles published before 2000 were excluded. In addition, studies that

113 focused exclusively on waste without including the agricultural or food production phase (e.g.
114 production of biogas), were excluded. Review articles and articles that did not include a case study
115 (e.g. articles focusing on methodology) were also excluded. The identification of all relevant studies
116 was done by using a screening process based on the PRISMA flow diagram (Moher et al., 2009), as
117 depicted in Fig. 1. Initially, all papers retrieved from Scopus and Web of science were transferred into
118 one database using EndNote software, and duplicates were removed (819). Subsequently, articles
119 were screened based on their titles, which led to the exclusion of 1935 articles because they did not
120 consider the LCC of agri-food products. Next, 306 articles were excluded based on their abstracts, and
121 4 more papers were excluded based on the accessibility. The remaining 165 papers were screened in-
122 depth. Based on the inclusion criteria, 73 papers were excluded during full paper screening (e.g. no
123 case study, focus on food packaging, LCC not explicitly mentioned,...). As a result, 92 relevant papers
124 were included for data extraction.

125 2.3. Data extraction

126 All relevant information was extracted from the papers by using data extraction sheets. Three
127 extraction sheets were developed to gather information on methods, findings and quality. First,
128 general information like authors, title, publication year, database, journal, country, and product type
129 was extracted. Next, all relevant information regarding the LCC methodology was extracted. This
130 included information on the goal of the LCC, functional unit, system boundaries, included stakeholders,
131 data source, discount rate, sensitivity analysis, included costs, visual reporting, use of other economic
132 parameters, use of other life cycle tools and whether results of these tools are integrated with LCC
133 results. As for the data extraction sheet on findings, information on cost hotspots and main conclusions
134 from LCC and other life cycle tools (e.g. LCA and social LCA) were gathered. A final data extraction sheet
135 was developed to assess the quality of the selected studies. The yes/no questions in this sheet were
136 based on existing critical appraisal checklists for economic evaluation from the Joanna Briggs Institute

137 (JBI) (Gomersall et al., 2015) and the Scottish Intercollegiate Guidelines Network (SIGN) (Sutherland et
138 al., 2015).



139

140 Fig. 1. Screening process, based on the PRISMA flow diagram (Moher et al., 2009), to identify all

141

relevant studies.

142 3. Results and discussion

143 3.1. Study characteristics

144 In total, relevant data was extracted from 92 articles. The articles were published in a wide range of
145 different journals, from journals focusing directly on life cycle assessment (e.g. 6 articles in the
146 International Journal of Life Cycle Assessment) to journals with a broader scope like the Journal of
147 Cleaner Production (20 articles), Sustainability (12 articles) and Science of the Total Environment (5
148 articles). The number of published LCC studies has increased significantly over the recent years. Only
149 9 LCC studies were published from 2008 (oldest studies in dataset)-2013. During the period 2014-2016
150 15 studies were published, while 26 studies were published from 2017-2019. During the last 2 years
151 and a half (2020-2022) 42 LCC papers got published, showing the increased interest of the scientific
152 community in life cycle thinking. Most studies (66) perform a life cycle cost analysis in Europe. Life
153 cycle costing studies have been focusing mainly on Italian agri-food products (28 articles). After Italy,
154 most LCC research has been conducted in the United Kingdom (7 articles). Most of the papers (59) deal
155 with agricultural products, of which 30 articles focus on perennials, 22 on annual crops and 7 papers
156 focus on both perennials and annuals. Regarding the product type, 19 papers focus on food products,
157 2 on beverages and 1 paper focuses on a food product and beverage. The remaining papers perform
158 an LCC on livestock (8 studies) or other products (3 studies; investment in poultry shed, investment in
159 beekeeping infrastructure, feed).

160 3.2. Methodological aspects

161 3.2.1. Goal and scope of life cycle costing studies

162 Most LCC studies (54 studies) compare 2 or more different production scenarios, for example, the
163 organic versus conventional production of olives (Iofrida et al., 2020). Eight papers compare the life
164 cycle costs of different products based on specific production scenarios, for example, the organic and
165 conventional production of both oranges and lemons (Pergola et al., 2013). Fifteen papers focus on

166 the comparison of the life cycle costs of different products like the study from Wagner et al. (2019b),
167 in which biogas costs are compared to the costs of fossil fuel. Apart from this, 15 papers neither
168 compare products nor production scenarios. Following these observations, it can be stated that life
169 cycle costing can be used for different purposes, such as choosing between different options or
170 evaluating the life cycle cost of 1 specific product.

171 The type of LCC is not mentioned in 61 papers (Fig. 2). In the studies that mention the type of LCC (31,
172 one paper applies 2 different types of LCC), conventional LCC is most often referred to (21 papers),
173 followed by environmental LCC (9 papers) and finally societal LCC (2 papers). For the societal LCC, the
174 social cost of externalities, in this case from atmospheric pollutants (Blanc et al., 2019) and air, water,
175 soil pollutants, indirect land use change and transport (Albizzati et al., 2021), were added to the
176 conventional costs. The lack of studies including a societal life cycle cost analysis might be caused by
177 the difficulties associated with conducting a societal LCC, such as the risk of double counting (when
178 LCA is complemented by LCC, the costs of environmental externalities should not be included in the
179 LCC) and methodological difficulties with regards to internalizing externalities (which externalities
180 should be internalized?, how can externalities be internalized?, etc.) (Neugebauer et al., 2016).
181 According to the Society of Environmental Toxicology and Chemistry (SETAC), environmental LCC takes
182 the costs over the complete life cycle of a product for one or more actors, including the soon-to-be-
183 internalized externalities, into account and is complemented by an LCA (Hunkeler et al., 2008). In
184 contrast, a conventional LCC takes the costs over the product's life cycle, excluding the end-of-life
185 phase, for one actor into account and is not accompanied by an LCA (Hunkeler et al., 2008). The
186 definition for conventional and environmental LCC, given by e.g. Hunkeler et al. (2008), is often not
187 followed in the analyzed studies. For example, both types are often accompanied by a life cycle
188 assessment (all articles using environmental LCC and 17 out of 21 articles for conventional LCC). As a
189 result, there is no clear distinction between LCC types. This finding is in line with the results of the
190 review paper from De Menna et al. (2018). Most papers conducting an environmental LCC take the
191 consumption phase and end-of-life costs (cradle-to-grave) into account, except for Florindo et al.

192 (2017) and García-Herrero et al. (2022). On the contrary, papers conducting a conventional LCC do not
193 consider the consumption phase and end-of-life costs (cradle-to-gate), except for Albizzati et al. (2021).
194 In conclusion, there is a need for standardized definitions for the different LCC types, which researchers
195 can follow when conducting an LCC.

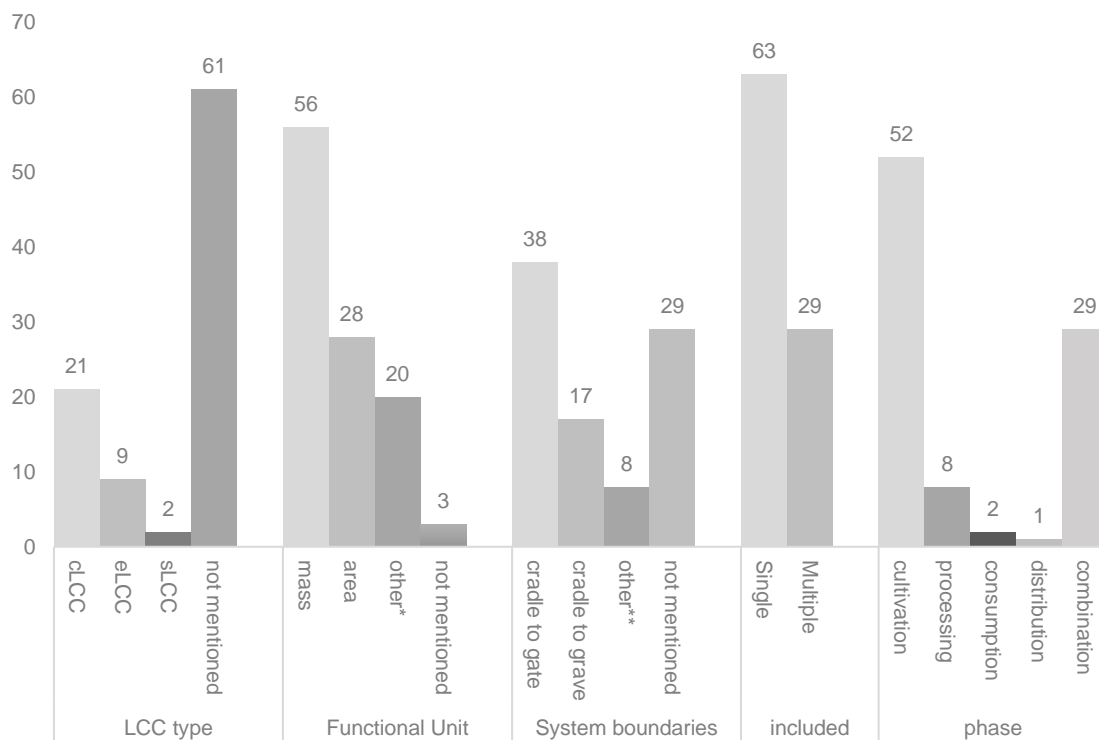
196 Three papers do not explicitly mention a functional unit (Fig. 2) but calculate the total cost of a firm
197 over the entire production facility's lifetime (Koričan et al., 2022; Le Feon et al., 2021; Strano et al.,
198 2015). In the papers that express the life cycle costs per functional unit, mass (56 papers; e.g. costs per
199 kg) and area (28 papers; e.g. costs per ha) are most often used (Fig. 2). Only 15 studies calculate the
200 life cycle costs for more than one functional unit. The functional unit should be carefully chosen since
201 the life cycle costs might differ significantly depending on the functional unit (e.g. yield impacts the life
202 cycle costs per kg, while it does not impact the life cycle costs per ha) (Fenollosa et al., 2014; Tamburini
203 et al., 2015), which is why the use of more than 1 functional unit might be recommended.

204 Surprisingly, the system boundaries used to conduct a life cycle cost analysis are often not explicitly
205 mentioned (29 papers) (Fig. 2). For the studies that mention the system boundaries, a cradle-to-gate
206 approach is most often used (38 papers), followed by a cradle-to-grave approach (17 papers) and some
207 other approaches like for example a cradle-to-fork approach. The differences between these
208 approaches are not always clearly defined in the studies. For example, a study claiming to include the
209 whole life cycle of a product (cradle-to-grave), does not include the transformation, distribution and
210 consumption phase (De Gennaro et al., 2012), in contrast to other studies. These discrepancies
211 between studies make it more difficult to compare results. Some authors use approaches similar to
212 cradle-to-gate and cradle-to-grave but name them differently, like cradle-to-use (Baquero et al., 2011),
213 cradle-to-fork (Sanyé-Mengual et al., 2018), cradle-to-consumer (Sanyé-Mengual et al., 2015), cradle-
214 to-retail (Verduna et al., 2020) and cradle-to-market (Zhen et al., 2020).

215 Finally, 63 studies consider only one actor (e.g. farmer or processor) in the value chain, as compared
216 to 29 articles that consider multiple actors (e.g. farmer + processor, processor + consumer, farmer +

processor + consumer) for life cycle cost calculation (Fig. 2). The life cycle costs assigned to the cultivation phase of agri-food products are targeted in 52 studies. Besides the cultivation phase, the costs assigned to the processing phase (8 studies), distribution phase (1 study), consumption phase (2 studies) or a combination of multiple phases in the value chain (29 studies) are considered in the literature as well. The necessity of including multiple phases depends on the goal and focus of the study. No significant differences are observed for these methodological aspects when comparing older studies with studies that have been published more recently.

224



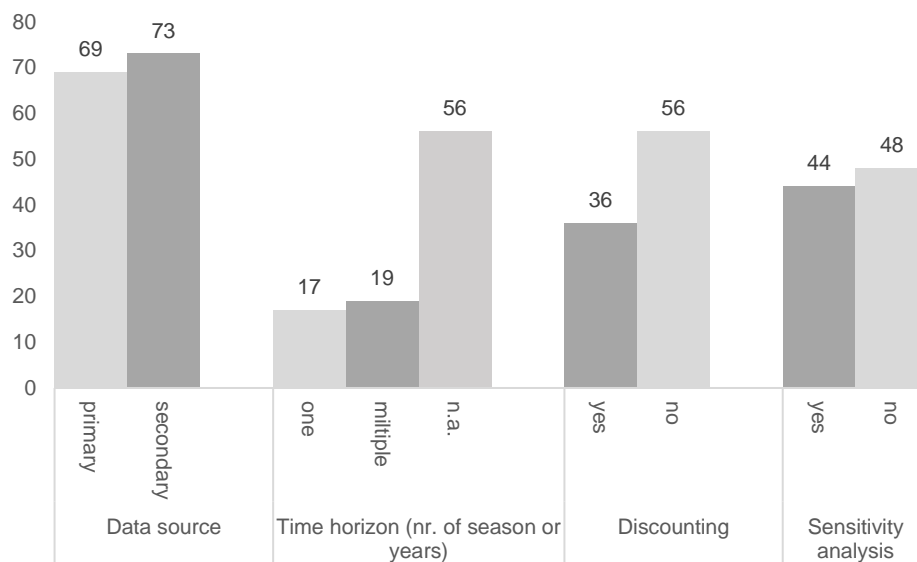
225

226 Fig. 2. Number of studies according to LCC type, functional unit, system boundaries, included number
 227 of stakeholders and included life cycle phase. *volume, 1 meal, annual consumption, energy, power
 228 needed for 1km of driving. **cradle-to-use, cradle-to-fork, cradle-to-consumer, cradle-to-market.

229 3.2.2. Life cycle costing data

230 Both primary (69 studies) and secondary (73 studies) data are used to calculate a product's life cycle
 231 cost (Fig. 3). Primary data is gathered directly from stakeholders, for example, the farmer or processor,

232 while secondary data sources consist of life cycle inventory databases like Ecoinvent, which is used in
 233 most papers, agricultural databases and data from scientific papers. In more than half of the studies
 234 (50), primary data is collected and complemented by secondary data where necessary. Concerning the
 235 primary data, a wide variety of sample sizes was used to calculate life cycle costs, ranging from 1 to
 236 213. The sampling method is not specified in most papers, except for the method in De Luca et al.
 237 (2014) (non-probability sampling), Jirapornvaree et al. (2021) (non-probability sampling), Li et al.
 238 (2021) (random sampling), Omran et al. (2021) (simple random sampling), Pergola et al. (2013)
 239 (stratified sampling) and Stillitano et al. (2016) (stratified sampling). If possible, the use of a larger
 240 sample size is more appropriate since it will give a better representation of the average life cycle cost.
 241 Additionally, data from multiple seasons/ years is preferred for a more representative average cost
 242 and to control for seasonal fluctuations in yield and external factors that influence the productivity (De
 243 Luca et al., 2014; De Luca et al., 2018b; Falcone et al., 2016). A similar number of studies consider
 244 multiple seasons or years (19 studies) and only one season or year (17), while in most papers (56) this
 245 is not mentioned or irrelevant (Fig. 3).



246
 247 Fig. 3. Number of studies based on to the used data source, time horizon, application of discounting,
 248 and application of a sensitivity analysis.

249 Less than half of the studies discount future cash flows to deal with the time value of money (Fig. 3).
250 The discount rate ranges from 1,12% to 12,5%. For a third of the studies, there is no explanation given
251 for the choice of discount rate. In general, agricultural investments are considered as low risk, which
252 is why a low discount rate is used (Mohamad et al., 2014; Stillitano et al., 2016). Higher discount rates
253 are used if the investment risk is perceived to be higher, for example for urban plant factories (Liaros
254 et al., 2016). The result of using a high discount rate is that the net present value will be lower.
255 Discounting should only be used if future cash flows are expected, which is not the case for annual
256 crops. Furthermore, papers focusing on food or beverage production without considering investment
257 costs for machinery and equipment, did not apply discounting (e.g., Amienyo and Azapagic (2016) and
258 Konstantas et al. (2019)). If investment costs are considered and cash flows are calculated for multiple
259 years or the entire lifetime of a project, discounting is applied (e.g. in Strano et al. (2015)).

260 Almost half of the studies (44) conduct a sensitivity analysis to deal with uncertain data (Fig. 3). A
261 sensitivity analysis can be done by manually varying certain parameters (e.g. yield, discount rate,...
262 (Falcone et al., 2016; Kim et al., 2018)) or by statistically running Monte Carlo simulations (Canaj et al.,
263 2021b; De Gennaro et al., 2012; Konstantas et al., 2019; Olba-Ziety et al., 2022; Roselli et al., 2020) to
264 see the effect on the final life cycle cost of the product.

265 All LCC studies included the costs of inputs, for instance raw materials, water and fuel (Fig. 4). Most
266 studies also include labor (72) and machinery (64) as costs (Fig. 4). In contrast, machinery is often not
267 included in life cycle assessment studies due to the lack of data (Silva and Sanjuán, 2019). In some
268 articles, it is not clear if labor and machinery are included because no details on the included costs are
269 given (e.g. Rivera and Azapagic (2016)). Other costs related to, for example, taxes, insurance, services,
270 interests, land capital, and food losses, are also included in most articles (67) (Fig. 4). No significant
271 differences are observed for the inclusion of the different cost categories when comparing older
272 studies with studies that have been published more recently. The cost of (some) externalities is only
273 considered in 12 studies (Albizzati et al., 2021; Blanc et al., 2018; Blanc et al., 2019; Canaj et al., 2021a;

274 Canaj et al., 2021b; Dobon et al., 2011; Moosavi-Nezhad et al., 2022; Rahmah et al., 2022; Ribeiro et
275 al., 2018; Ruviaro et al., 2020; Saber et al., 2020; Smith and Lal, 2022) (Fig. 4). Most of the papers (8)
276 only include the societal/environmental cost of atmospheric pollutants since the costs for those have
277 been quantified in previous studies and are more established (lower degree of uncertainty) than the
278 costs of other environmental indicators (Yang et al., 2022) (e.g. emission costs validated by the US
279 government (Smith and Lal, 2022)). The remaining papers (4) included more externalities in the LCC
280 than only the atmospheric pollutants, for example water pollutants, soil pollutants, externalities of
281 transport and land use change (Albizzati et al., 2021), all environmental indicators from the LCA (Canaj
282 et al., 2021a; Canaj et al., 2022) and noise (Dobon et al., 2011). To calculate the costs of these
283 externalities most studies multiplied LCA results with cost values from literature, while Blanc et al.
284 (2018) and Blanc et al. (2019) used the value of a life year (VOLY) approach to calculate the costs of
285 externalities to society. According to the review paper from Amadei et al. (2021) there is a lack of
286 recent research regarding the monetary valuation of LCA results. The monetization of LCA results can
287 be done by using different methodologies, e.g. observed preferences (market price), revealed
288 preferences, stated preferences, budget constraint and abatement cost, each having its own strengths
289 and weaknesses (Amadei et al., 2021; Arendt et al., 2020; Pizzol et al., 2015). The monetary values of
290 LCA results vary a lot across different studies (Arendt et al., 2020), leading to difficulties when
291 comparing results. If the LCC is complemented by an LCA and the costs of externalities are included,
292 there is a risk of double-counting environmental or social indicators (Hunkeler et al., 2008), which
293 should be avoided. There is no clear case of double-counting in the 12 studies including externalities.
294 More details about the included cost types for each study can be found in Table 2. When comparing
295 the included cost categories per product type, studies regarding the life cycle costs of perennial crops
296 generally include most cost types. Each cost category is included in at least 72% of the studies. For
297 studies focusing on annual crops and food products/beverages, each category is included in around
298 60-70% and 45-70% of the studies, respectively, except for the category "inputs", which is included in
299 all studies. The 8 studies concerning livestock include all cost types, except for the category "other",

300 which is only included in 6 studies. The remaining 3 studies, focusing on feed, the investment in a
301 poultry shed and beekeeping, include most cost categories.

302 More than 65% of the studies consider revenues to calculate the profitability of the product, aside
303 from the calculation of the life cycle cost of a particular product. Thirty-eight papers calculate 1 or
304 more economic evaluation parameter(s). The net present value (NPV) is used in most cases (24),
305 followed by the internal rate of return (IRR) (14). In some studies, other economic parameters like
306 added value and payback period, are calculated.

307 3.2.3. Life cycle costing integrated with other life cycle tools

308 Out of 92 published studies reviewed in this systematic review, 72 studies concerning LCC also include
309 other life cycle tools, like LCA (71) and social LCA (sLCA) (13). For these studies, LCC, LCA and sLCA are
310 performed individually based on the same functional units and system boundaries. However, it is
311 rather uncommon that researchers integrate the results from these assessments. Only 29 papers use
312 an integrated approach to present LCA and LCC results together (Table 1). In this dataset, sLCA was
313 never conducted before 2017, while almost half of the studies did a life cycle assessment. Around 20%
314 of the studies before 2017 integrated the results from both assessments. More recently (from 2017
315 onwards), studies have increasingly conducted LCA (>80%) and sLCA (15-27%), and have started to
316 integrate the results of multiple life cycle tools (around 30%). These results show that a more
317 comprehensive sustainability assessment, focusing on the 3 pillars of sustainability, is starting to gain
318 more attention from researchers.

319 Life cycle assessment is performed for different agri-food products to evaluate the ecological impacts
320 of the product. The common goal of all the research is to find possible ways to reduce environmental
321 and economic impacts and highlight further optimization potential. Many different methods can be
322 used to integrate LCC and LCA results. Eco-efficiency analysis, which combines the economic index with
323 environmental burden indices (World Business Council for Sustainable Development, 2006) (Eco-
324 efficiency = (Product value)/(Environmental impact)), and multi-criteria decision-making methods are

325 mostly used within the list of studies (Table 1). To calculate the eco-efficiency different indicators are
 326 used for the economic (e.g. LCC, profit) and environmental (e.g. GWP, multiple indicators) part of the
 327 formula. Multi-criteria decision-making methods in combination with life cycle thinking methods have
 328 been identified as valuable in the review paper from De Luca et al. (2017). It can be concluded that the
 329 integration of economic and environmental results is still limited but can be very valuable from a
 330 research point of view but also for policy-makers and industry for decision-making.

331 Table 1. References that integrate LCC and LCA results, the methodology used for integrating and the
 332 visual reporting style for results.

Reference	Methodology for integrated results	Visual reporting style
Valente et al. (2020); Zira et al. (2021)	The sustainability framework/ LCSA framework	Radar chart
Baum and Bienkowski (2020); Canaj et al. (2021b); Sanyé-Mengual et al. (2018); Zhen et al. (2020); Laso et al. (2018); Konstantas et al. (2019); Konstantas et al. (2020); Pari et al. (2022); Pari et al. (2020); Pexas et al. (2021)	Eco-efficiency analysis	Scatter plots; Bar chart; Table; Heat map
Albizzati et al. (2021); Chen and Holden (2018)	Multi-criteria analysis – No specific method mentioned	Radar chart
De Luca et al. (2018a)	Multi-criteria analysis - Analytic Hierarchy Process (AHP)	Pie chart, Bar chart
Dobon et al. (2011)	Sustainability analysis (LCA + LCC + Willingness to Pay)	Table
Ekener et al. (2018)	Multi-criteria analysis – Multi-Attribute Value Theory	Table; bar chart
Falcone et al. (2016); Florindo et al. (2020)	Multi-criteria analysis – VIKOR method	Scale chart using Composite Index of Sustainability; Table
Le Feon et al. (2021)	Multi-criteria analysis – DEXi method	Table, radar chart
Ribeiro et al. (2018)	Social Return on Investment (SROI)	Color scale
Rivera and Azapagic (2016)	Quantitative approach	A combined heat map
Tamburini et al. (2015); Canaj et al. (2021a); Canaj et al. (2022)	Semi-quantitative approach	Table; bar chart
Wohner et al. (2020)	Technique for Order by Similarity to Ideal Solution (TOPSIS)	Bar chart and color matrix
Zortea et al. (2018)	Dashboard of Sustainability (DoS)	Windrose diagram with sustainability indicators

333

334 3.3. Critical appraisal

335 All studies have a clear research objective, with the majority (90) also having a clear economic
336 importance. For 19 studies it is not entirely clear which costs are included and which are not. For those
337 who clearly described the included costs, 57 studies included all cost categories, while 16 studies did
338 not. Discounting has been applied in 37 studies. Fifty of the remaining studies did not apply discounting
339 because there were no future cash flows (n.a. in Table 1 in appendix), while 5 studies did not consider
340 the time value of money (no in Table 1 in appendix). Almost half (44) of the studies conducted a
341 sensitivity analysis. Finally, most of the studies (77) clearly describe the economic results. The quality
342 aspect with the worst score is the sensitivity analysis, while the other aspects score relatively well.

343 Most studies (59) score at least 80% on the checklist for quality evaluation, which corresponds to 5
344 stars (Table 1 in appendix). Twenty studies score between 60 and 80% (4 stars). Twelve studies get 3
345 stars (40-60%) and finally only 1 study scores 2 stars (20-40%). It can be concluded that most studies
346 have a high quality. However, some methodological aspects, discussed in section 3.2., are also
347 important indicators for the quality of LCC studies (e.g. clear LCC type and clear system boundaries). If
348 these methodological aspects would have been included for quality appraisal, the quality of many
349 papers would have been lower. When conducting a life cycle costing study, all these aspects should be
350 considered to develop a high-quality study.

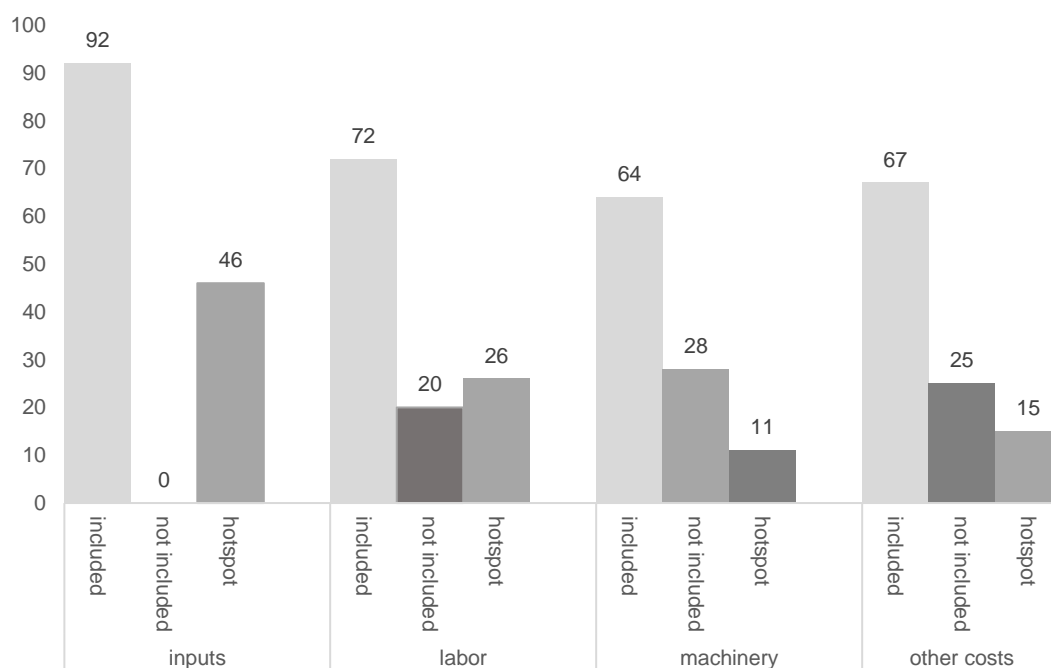
351 3.4. Reporting and findings

352 Life cycle costing is applied to a wide range of different agri-food products. For some studies, the goal
353 is to identify cost hotspots and to explore if the production is economically profitable, while other
354 studies compare products or production scenarios. Findings are most often visually reported by the
355 use of tables (59 studies), followed by column or bar charts (39). Other charts for visually representing

356 the findings are also used, for example, line charts, pie charts and scatter plots. In almost 30% of the
 357 studies, a combination of different reporting methods is used.

358 3.4.1. Economic hotspots

359 Out of 92 studies, 55 identified the economic hotspots (most significant costs) for specific agri-food
 360 products, while the remaining studies did not give any details or used a different cost typology (e.g.
 361 total cost for each activity/ stakeholder instead of dividing costs into the categories “inputs”, “labor”,
 362 “machinery” and “other”). Out of the 55 studies that identified hotspots, inputs, like raw materials,
 363 fertilizer, fuel, etc., were identified as a cost driver in 46 studies (84%; Fig. 4). Labor and machinery
 364 were identified as main cost drivers in 26 and 11 studies, respectively. Finally, 15 studies identified
 365 other main hotspots like for example costs related to land and end-of-life costs (Florindo et al., 2017;
 366 Lask et al., 2020; Ruviano et al., 2020; Stillitano et al., 2019; Strano et al., 2017; Zhen et al., 2020; Zortea
 367 et al., 2018). If labor, machinery and other costs are considered in the study and economic hotspots
 368 are identified, they are cost drivers in 62% (26 out of 42 studies), 27% (11 out of 41 studies) and 39%
 369 (15 out of 38 studies) of the studies, respectively. Considering this result, it is clear that each cost
 370 category should be considered in future life cycle costing studies of agri-food products.



371

372 Fig. 4. Number of studies that include inputs, labor, machinery and other costs in the LCC analysis
 373 and number of studies that identified inputs, labor, machinery and other costs as an economic
 374 hotspot.

375 When comparing the cost hotspots for each product category, it is clear that inputs are an important
 376 cost driver for all products (Table 2). Inputs are identified as a hotspot in all studies focusing on annual
 377 crops, while for perennials, food products/ beverages and livestock production, input costs are a
 378 significant cost in 89%, 69% and 50% of the studies. For the remaining 3 studies, inputs are identified
 379 as a cost hotspot in only 1 study regarding the production of feed. Labor seems to be an important
 380 cost driver for agricultural products and food/ beverages (hotspot in 57-70% of the studies) and less
 381 important for livestock production (hotspot in 25% of studies). Machinery contributes less to the total
 382 life cycle cost of agricultural products, food/ beverages and livestock production (hotspot in 0-25% of
 383 studies). Finally, the cost type “other” was often (in 30-55% of the studies) identified as an economic
 384 hotspot among the different product categories.

385 Table 2. Included cost types and identified costs hotspots in the 92 LCC studies. x: cost type included,
 386 x*: cost type identified as a hotspot.

PRODUCT	INCLUDED COSTS				
	inputs	labor	machinery	other	externalities
Annual crop					
Baquero et al. (2011)	x	x	x	x	
Baum and Bienkowski (2020)	x*	x	x	x	
Brandão et al. (2010)	x*		x		
Canaj et al. (2021a)	x*				x*
Canaj et al. (2022)	x*		x		x*
Dorr et al. (2017)	x*				
Ekener et al. (2018)	x	x	x		
Escobar et al. (2022)	x*	x*	x*	x	
Fenollosa et al. (2014)	x*	x*	x*		
Holka (2020)	x	x	x	x	
Holka and Bieńkowski (2020)	x*	x	x*	x	
Hong et al. (2015)	x*	x*	x	x*	
Jirapornvaree et al. (2021)	x	x	x	x	
Kim et al. (2018)	x*		x	x	
Lask et al. (2020)	x*	x	x	x*	

Liaros et al. (2016)	x*	x	x	x	
Lokesh et al. (2019)	x*	x*			
Moosavi-Nezhad et al. (2022)	x	x	x		
Moungsree et al. (2022)	x*				
Pattanaik et al. (2020)	x*	x	x	x	
Saber et al. (2020)	x				x
Sanyé-Mengual et al. (2018)	x*				
Sanyé-Mengual et al. (2015)	x*	x*	x	x	
Schulte et al. (2021)	x				
Tamburini et al. (2015)	x*	x*	x	x	
Tziolas and Bournaris (2019)	x	x	x	x	
Venanzi et al. (2018)	x*	x*	x	x	
Zhen et al. (2020)	x*	x*	x	x*	
Zortea et al. (2018)	x*	x		x*	
Perennial crop					
Blanc et al. (2018)	x	x	x	x	x
Blanc et al. (2019)	x*	x*	x	x	x
Bosona et al. (2019)	x	x	x	x	
Brandão et al. (2010)	x*		x		
Canaj et al. (2021a)	x*				x*
Canaj et al. (2021b)	x*		x		
De Gennaro et al. (2012)	x	x			
De Luca et al. (2018a)	x*	x*	x*	x	
De Luca et al. (2014)	x	x		x	
Falcone et al. (2016)	x*	x*	x*	x*	
García-Herrero et al. (2022)	x*	x	x	x	
Hanif et al. (2016)	x*	x	x	x	
Iofrida et al. (2020)	x*	x*	x	x	
Lask et al. (2020)	x*	x	x	x*	
Lee et al. (2021)	x*	x	x	x*	
Lerkkasemsan and Achenie (2013)	x	x	x	x	
Lokesh et al. (2019)	x*	x*			
Luo et al. (2009)	x		x		
Mohamad et al. (2014)	x	x		x	
Nguyen et al. (2008)	x	x	x	x	
Olba-Ziety et al. (2022)	x*	x*	x*		
Omran et al. (2021)	x*	x*	x*	x	
Pari et al. (2022)	x	x		x	
Pari et al. (2020)	x	x		x	
Pergola et al. (2013)	x	x	x	x	
Rahmah et al. (2022)	x	x*	x	x	x
Roselli et al. (2020)	x	x		x	
Schulte et al. (2021)	x				
Smith and Lal (2022)	x	x	x	x	x
Soldatos (2015)	x	x	x	x	
Stillitano et al. (2016)	x*	x*	x*	x	
Strano et al. (2017)	x	x*	x	x*	
Styles et al. (2008)	x			x	

Tamburini et al. (2015)	x*	x*	x	x	
Tziolas and Bournaris (2019)	x	x	x	x	
Wagner et al. (2019a)	x	x	x	x	
Wagner et al. (2019b)	x*	x	x	x*	
Food/beverage					
Albizzati et al. (2021)	x	x	x	x	x
Amienyo and Azapagic (2016)	x*			x	
Cacace et al. (2020)	x*	x	x*	x*	
De Luca et al. (2018b)	x	x		x*	
Dobon et al. (2011)	x	x	x	x	x
Falcone et al. (2017)	x*	x*		x	
García-Herrero et al. (2021)	x	x*		x	
García-Herrero et al. (2019)	x	x*	x	x*	
Gosalvitr et al. (2021)	x*		x	x	
Iotti and Bonazzi (2014)	x	x	x	x	
Konstantas et al. (2019)	x*				
Konstantas et al. (2020)	x*				
Laso et al. (2018)	x			x	
Li et al. (2021)	x*	x*			
Ribeiro et al. (2018)	x*	x*	x	x	x
Rivera and Azapagic (2016)	x*				
Ruviaro et al. (2020)	x*			x*	x
Stillitano et al. (2019)	x	x	x*	x*	
Valente et al. (2020)	x	x*	x		
Verduna et al. (2020)	x*	x*	x	x	
Wohner et al. (2020)	x			x	
Zira et al. (2021)	x	x			
Livestock					
Chen and Holden (2018)	x*	x	x	x	
Diaz et al. (2021)	x	x	x	x	
Florindo et al. (2020)	x	x*	x		
Florindo et al. (2017)	x*	x	x	x*	
Geß et al. (2022)	x	x	x	x	
Koričan et al. (2022)	x	x	x*		
Le Feon et al. (2021)	x	x	x	x	
Pexas et al. (2021)	x	x	x	x	
Feed					
Roffeis et al. (2018)	x*	x*	x		
Poultry shed					
Boggia et al. (2019)	x	x	x	x	
Beekeeping					
Strano et al. (2015)	x	x	x	x	

388 3.4.2. Comparison of studies

389 To discuss LCC results, comparable studies were grouped into 4 categories: organic vs. conventional,
390 bio-energy vs. fossil fuel, innovative vs. conventional and valorization of byproducts vs. no valorization
391 (Table 3). Studies which did not fit in any of the categories (42) will not be discussed. The comparison
392 between conventional and organic production is frequently made in life cycle costing studies
393 concerning agricultural products. There is no clear consensus between studies about which production
394 technique scores best economically and environmentally. For example, Mohamad et al. (2014) found
395 that the conventional scenario for olive production had the lowest life cycle cost, while the organic
396 scenario scored best on an environmental level. Pergola et al. (2013) identified the organic production
397 of lemon and oranges as the least costly and most environmentally sustainable scenario. In the study
398 from Falcone et al. (2016) the conventional production scenario for vineyards was considered as the
399 least costly and most environmentally sustainable. The production scenario with the lowest life cycle
400 costs is not necessarily the most profitable, since there is often a higher selling price for organic
401 products. The comparison of bio-energy from different sources with fossil fuel is also regularly made
402 within the LCC studies. Despite the consensus about the environmental benefits of bio-energy
403 compared to fossil fuel, there is no agreement on the least costly option (bio-energy versus fossil fuel)
404 (e.g. contradictory results in Lerkkasemsan and Achenie (2013) and Hanif et al. (2016)). In many
405 studies, agricultural innovations or innovations in the food industry are compared with the
406 conventional production scenario. Innovations can be beneficial both economically and ecologically,
407 e.g. the use of a shelf life extension technique that reduces food loss (Falcone et al., 2017). Most
408 studies focusing on agricultural innovations or innovations in the food industry demonstrated a
409 reduction of the environmental impact compared to the conventional scenario, whereas the life cycle
410 cost was higher for the innovation in half of the studies. The valorization of by-products is considered
411 beneficial from an economic and ecological point of view, within the studied articles.

412 Table 3. Results from LCC studies in which products or production scenarios have been compared,
 413 sorted by type of comparison.

Agricultural products: organic vs. Conventional (vs. Other)						
	<u>Lowest LCC</u>			<u>Best LCA</u>		
	Organic	Conventional	Other	Organic	Conventional	Other
De Luca et al. (2018a)			x			x
De Luca et al. (2014)		x		x		
Falcone et al. (2016)		x			x	
Fenollosa et al. (2014)		x		x		
Iofrida et al. (2020)		x		No LCA		
Jirapornvaree et al. (2021)	x			x		
Mohamad et al. (2014)		x		x		
Pergola et al. (2013)	x			x		
Rahmah et al. (2022)		x		x		
Saber et al. (2020)	x			x		
Stillitano et al. (2016)	x			No LCA		
Strano et al. (2017)	x			x		
Zhen et al. (2020)		x				x
Zira et al. (2021)		x		x	x	Depends on indicator
Bio-energy vs. Fossil fuel						
	<u>Lowest LCC</u>			<u>Best LCA</u>		
	Bio-energy	Fossil fuel	Comment	Bio-energy	Fossil fuel	Comment
Ekener et al. (2018)	x	x	Sugarcane lower, corn higher	x	x	Sugarcane lower, corn higher
Hanif et al. (2016)	x			x		
Koričan et al. (2022)	x			x		
Lask et al. (2020)	x		Maize	x		Wild plant mixtures
Lerkkasemsan and Achenie (2013)		x		x		
Luo et al. (2009)	x			x		
Nguyen et al. (2008)		x		x		
Wagner et al. (2019a)	x		Miscanthus	x		Miscanthus
Agricultural innovations and innovations in the food industry: innovative vs. Conventional scenario						
	<u>Lowest LCC</u>			<u>Best LCA</u>		
	Innovative	Conventional	Comment	Innovative	Conventional	Comment
Baquero et al. (2011)		x				No LCA
Blanc et al. (2018)		x		x		
Blanc et al. (2019)		x		x		
Boggia et al. (2019)		x		x		
Canaj et al. (2021a)		x		x		
Canaj et al. (2021b)		x		x		
Canaj et al. (2022)	x			x		

Cacace et al. (2020)			Depends on product	x		
De Gennaro et al. (2012)		x				x
De Luca et al. (2018b)	x			x		
Diaz et al. (2021)			Not clear	x		
Dobon et al. (2011)	x					No LCA
Falcone et al. (2017)	x			x		
Holka (2020)		x		x		
Holka and Bieńkowski (2020)	x	x	Reduced tillage lowest, no tillage highest	x		
Iotti and Bonazzi (2014)	x					No LCA
Lokesh et al. (2019)	x			x		
Moosavi-Nezhad et al. (2022)	x		Emission costs included	x		
Pexas et al. (2021)	x	x	Depends on strategy	x	x	Depends on indicator
Roffeis et al. (2018)		x				No LCA
Ruviaro et al. (2020)	x					No LCA
Sanyé-Mengual et al. (2015)	x			x		
Stillitano et al. (2019)		x			x	
Valente et al. (2020)	x					No LCA

Valorizing byproducts

	<u>Lowest LCC</u>			<u>Best LCA</u>		
	Valorization	No valorization	Comment	Valorization	No valorization	comment
Gosalvittr et al. (2021)	x			x	x	Depends on method
Laso et al. (2018)	x			x		
Lokesh et al. (2019)	x			x		
Venanzi et al. (2018)	x			x		

414

415 3.4.3. Results from the integration of LCC with LCA

416 only 12 studies from table 3 integrated the results from LCC and LCA data. According to Zhen et al.

417 (2020), community-supported agriculture is the most eco-efficient agricultural option compared to

418 conventional and organic. Canaj et al. (2021a) monetized the LCA results and compared the total cost

419 (internal + external) for different crops when using groundwater (conventional scenario) or reclaimed

420 water (innovative scenario) for irrigation. For most crops the innovative scenario scored best overall

421 (Canaj et al., 2021a). A similar study was done by Canaj et al. (2021b), assessing the eco-efficiency of

422 the use of reclaimed water (innovative scenario) in vineyards. In this study the innovative scenario was

423 also identified as the more eco-efficient one (Canaj et al., 2021b). Canaj et al. (2022) evaluated a smart

424 irrigation system, which was identified as being less costly, with the inclusion of external costs, than
425 farmer-led irrigation. The sustainability assessment of different olive growing systems was investigated
426 by De Luca et al. (2018a). The AHP method was used to integrate the results of LCA, LCC and social
427 LCA. Following this approach, the sustainability score of the low-dosage/no-tillage scenario was the
428 highest, in comparison to the conventional and organic scenario (De Luca et al., 2018a). Ekener et al.
429 (2018) used a multi-criteria analysis to identify fuel from sugarcane as the most sustainable option.
430 Differences were observed depending on the perspectives of the stakeholders (Ekener et al., 2018).
431 The sustainability of different wine-growing scenarios was analyzed by Falcone et al. (2016). The VIKOR
432 method was implemented to show the aggregated results of different aspects of sustainability and
433 revealed that the conventional scenario was the most sustainable (Falcone et al., 2016). Laso et al.
434 (2018), in their assessment of the fish canning industry with a focus on anchovy species, found that
435 the scenario where a circular economy approach was used, scored best in terms of the eco-efficiency
436 index, compared to other waste management scenarios (landfilling and incineration). Dobon et al.
437 (2011) used the LCA data from a previous study to integrate with the LCC data, which resulted in the
438 finding that the use of the FBBD (flexible best-before-date)-device (innovative scenario), which can
439 change the expiry date of a product based on temperature fluctuations, is the most sustainable if the
440 purchasing price does not exceed the willingness-to-pay of consumers. Different manure management
441 strategies were compared by Pexas et al. (2021). Not all innovative scenarios were more eco-efficient
442 than the conventional one (Pexas et al., 2021). Valente et al. (2020) compared the sustainability of an
443 innovative slaughter system with the conventional system on a radar chart. The innovative system was
444 more sustainable for some indicators (Valente et al., 2020). Finally, Zira et al. (2021) compared the
445 organic with the conventional pork supply chain by calculating the relative sustainability point, with
446 the conventional chain as the benchmark. The more sustainable option was different for different
447 sustainability indicators (Zira et al., 2021). It can be concluded that an integrated assessment can be
448 very useful for decision-making if LCC and LCA results are contradictory, e.g. Zhen et al. (2020).

449 3.5. Limitations and future research

450 One of the most important limitations of this systematic review is that the difference between a life
451 cycle cost analysis and other types of cost analyses is not always clear. For example, in the study of
452 Gresta et al. (2014) the term “life cycle costing” is not explicitly mentioned, even though the
453 methodology of the economic analysis is similar to life cycle costing. Papers that should have
454 mentioned LCC but did not, were excluded for analysis. Currently, there is no standardized
455 methodology for LCC, which makes it difficult to compare results between different studies. This
456 systematic review allowed to identify different methodological inconsistencies and to recommend a
457 certain approach for some methodological aspects. Future research should propose a standardized
458 framework for life cycle costing studies. Thereby, a clear distinction between LCC types and different
459 system boundaries should be made. The amount of studies that internalized externalities is still limited
460 (12) due to methodological difficulties, indicating the need for further research. In addition, a
461 standardized methodology for social life cycle assessment (sLCA), to assess the third pillar of
462 sustainability, is still missing (Onat et al., 2017). Finally, the integration of the 3 pillars of sustainability
463 through the life cycle sustainability assessment (LCSA), is also not standardized (Costa et al., 2019).
464 Future research should tackle these methodological gaps. The LCSA framework, for instance, could be
465 a useful tool for the analysis of the transformation of our food system. Currently, the use of this tool
466 in the agri-food sector is very limited and most research only focuses on the environmental impacts to
467 address increasing environmental concerns. Hence, there is a lot of potential for studies assessing the
468 3 pillars of sustainability.

469 4. Conclusions

470 This systematic review extracted data regarding the methodology and results of 92 life cycle costing
471 studies within the agri-food sector. LCC has been applied for a wide range of different products and
472 purposes, e.g. identifying cost hotspots and comparing products. An increase in LCC studies has been
473 observed in this review, underpinning the need for an in-depth insight into the LCC methodology. At

474 this moment, there is still no consensus amongst researchers about LCC definitions and system
475 boundary definitions, which leads to the lack of comparability between results. In addition, data
476 collection differs significantly between studies, for example, the inclusion or exclusion of specific cost
477 categories. Furthermore, as each cost category has been identified as a cost hotspot in our sample, it
478 is important that each category is considered in future LCC studies. The number of studies that included
479 externalities for the life cycle cost analysis is still very limited. Findings further show that LCC is often
480 applied to compare different production scenarios (e.g. conventional versus organic), innovative
481 production methods with the traditional one, bio-energy from different sources, bio-energy with fossil
482 fuel and the valorization of by-products compared to not valorizing side streams. While LCC can be
483 considered a very valuable tool to assess economic sustainability, its impact can become even more
484 far-reaching when standardized, allowing for comparability of evidence. Also the integration of LCC
485 and other life cycle tools, often analyzed through eco-efficiency and multi-criteria decision analysis,
486 has shown to be useful for decision-making, especially if the results from the analyses are
487 contradictory. This review observed that there is an increased interest for a more comprehensive
488 sustainability assessment, focusing on the 3 pillars of sustainability instead of applying only 1 life cycle
489 tool. In conclusion, this systematic review identified inconsistencies between studies and gave
490 recommendations for some methodological aspects. Using this systematic review as a tool, future
491 research should further standardize the LCC framework and tackle the identified methodological
492 issues.

493

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496

497 References

- 498 Albizzati PF, Tonini D, Astrup TF. A Quantitative Sustainability Assessment of Food Waste
499 Management in the European Union. *Environmental Science & Technology* 2021; 55: 16099-
500 16109. 10.1021/acs.est.1c03940
- 501 Amadei AM, De Laurentiis V, Sala S. A review of monetary valuation in life cycle assessment: State of
502 the art and future needs. *Journal of Cleaner Production* 2021; 329: 129668.
503 10.1016/j.jclepro.2021.129668
- 504 Amienyo D, Azapagic A. Life cycle environmental impacts and costs of beer production and
505 consumption in the UK. *International Journal of Life Cycle Assessment* 2016; 21: 492-509.
506 10.1007/s11367-016-1028-6
- 507 Arendt R, Bachmann TM, Motoshita M, Bach V, Finkbeiner M. Comparison of Different Monetization
508 Methods in LCA: A Review. *Sustainability* 2020; 12. 10.3390/su122410493
- 509 Baldini C, Gardoni D, Guarino M. A critical review of the recent evolution of Life Cycle Assessment
510 applied to milk production. *Journal of Cleaner Production* 2017; 140: 421-435.
511 10.1016/j.jclepro.2016.06.078
- 512 Baquero G, Esteban B, Riba JR, Rius A, Puig R. An evaluation of the life cycle cost of rapeseed oil as a
513 straight vegetable oil fuel to replace petroleum diesel in agriculture. *Biomass & Bioenergy*
514 2011; 35: 3687-3697. 10.1016/j.biombioe.2011.05.028
- 515 Baum R, Bienkowski J. Eco-Efficiency in Measuring the Sustainable Production of Agricultural Crops.
516 *Sustainability* 2020; 12. 10.3390/su12041418
- 517 Blanc S, Accastello C, Girgenti V, Brun F, Mosso A. Innovative Strategies for the Raspberry Supply
518 Chain: An Environmental and Economic Assessment. *Quality-Access to Success* 2018; 19: 139-
519 142.
- 520 Blanc S, Massaglia S, Brun F, Peano C, Mosso A, Giuggioli NR. Use of Bio-Based Plastics in the Fruit
521 Supply Chain: An Integrated Approach to Assess Environmental, Economic, and Social
522 Sustainability. *Sustainability* 2019; 11: 2475. 10.3390/su11092475
- 523 Boggia A, Paolotti L, Antegiovanni P, Fagioli FF, Rocchi L. Managing ammonia emissions using no-litter
524 flooring system for broilers: Environmental and economic analysis. *Environmental Science &
525 Policy* 2019; 101: 331-340. 10.1016/j.envsci.2019.09.005
- 526 Bosona T, Gebresenbet G, Dyjakon A. Implementing life cycle cost analysis methodology for
527 evaluating agricultural pruning-to-energy initiatives. *Bioresource Technology Reports* 2019;
528 6: 54-62. 10.1016/j.biteb.2019.02.006
- 529 Brandão M, Clift R, Milà LC, Basson L. A life-cycle approach to characterising environmental and
530 economic impacts of multifunctional land-use systems: An integrated assessment in the UK.
531 *Sustainability* 2010; 2: 3747-3776. 10.3390/su2123747
- 532 Cacace F, Bottani E, Rizzi A, Vignali G. Evaluation of the economic and environmental sustainability of
533 high pressure processing of foods. *Innovative Food Science & Emerging Technologies* 2020;
534 60. 10.1016/j.ifset.2019.102281
- 535 Canaj K, Mehmeti A, Berbel J. The Economics of Fruit and Vegetable Production Irrigated with
536 Reclaimed Water Incorporating the Hidden Costs of Life Cycle Environmental Impacts.
537 *Resources-Basel* 2021a; 10. 10.3390/resources10090090
- 538 Canaj K, Morrone D, Roma R, Boari F, Cantore V, Todorovic M. Reclaimed Water for Vineyard
539 Irrigation in a Mediterranean Context: Life Cycle Environmental Impacts, Life Cycle Costs, and
540 Eco-Efficiency. *Water* 2021b; 13. 10.3390/w13162242
- 541 Canaj K, Parente A, D'imperio M, Boari F, Buono V, Toriello M, et al. Can precise irrigation support the
542 sustainability of protected cultivation? A life-cycle assessment and life-cycle cost analysis.
543 *Water (Switzerland)* 2022; 14. 10.3390/w14010006
- 544 Chen W, Holden NM. Tiered life cycle sustainability assessment applied to a grazing dairy farm.
545 *Journal of Cleaner Production* 2018; 172: 1169-1179. 10.1016/j.jclepro.2017.10.264

546 Costa D, Quinteiro P, Dias AC. A systematic review of life cycle sustainability assessment: Current
547 state, methodological challenges, and implementation issues. *Science of The Total*
548 *Environment* 2019; 686: 774-787. 10.1016/j.scitotenv.2019.05.435

549 Darton RC. Chapter 14 - Setting a policy for sustainability: The importance of measurement. In:
550 Klemeš JJ, editor. *Assessing and Measuring Environmental Impact and Sustainability*.
551 Butterworth-Heinemann, Oxford, 2015, pp. 479-496. 10.1016/B978-0-12-799968-5.00014-2

552 De Gennaro B, Notarnicola B, Roselli L, Tassielli G. Innovative olive-growing models: an
553 environmental and economic assessment. *Journal of Cleaner Production* 2012; 28: 70-80.
554 10.1016/j.jclepro.2011.11.004

555 De Luca AI, Falcone G, Stillitano T, Iofrida N, Strano A, Gulisano G. Evaluation of sustainable
556 innovations in olive growing systems: A Life Cycle Sustainability Assessment case study in
557 southern Italy. *Journal of Cleaner Production* 2018a; 171: 1187-1202.
558 10.1016/j.jclepro.2017.10.119

559 De Luca AI, Falcone G, Stillitano T, Strano A, Gulisano G. Sustainability assessment of quality-oriented
560 citrus growing systems in Mediterranean area. *Quality - Access to Success* 2014; 15: 103-108.

561 De Luca AI, Iofrida N, Leskinen P, Stillitano T, Falcone G, Strano A, et al. Life cycle tools combined with
562 multi-criteria and participatory methods for agricultural sustainability: Insights from a
563 systematic and critical review. *Science of The Total Environment* 2017; 595: 352-370.
564 10.1016/j.scitotenv.2017.03.284

565 De Luca AI, Stillitano T, Falcone G, Squeo G, Caponio F, Strano A, et al. Economic and environmental
566 assessment of extra virgin olive oil processing innovations. *Chemical Engineering*
567 *Transactions* 2018b; 67: 133-138. 10.3303/CET1867023

568 De Menna F, Dietershagen J, Loubiere M, Vittuari M. Life cycle costing of food waste: A review of
569 methodological approaches. *Waste Management* 2018; 73: 1-13.
570 10.1016/j.wasman.2017.12.032

571 Diaz F, Vignati JA, Marchi B, Paoli R, Zanoni S, Romagnoli F. Effects of Energy Efficiency Measures in
572 the Beef Cold Chain: A Life Cycle-based Study. *Environmental and Climate Technologies* 2021;
573 25: 343-355. 10.2478/rtuct-2021-0025

574 Djekic I, Batlle-Bayer L, Bala A, Fullana-i-Palmer P, Jambrak AR. Role of the Food Supply Chain
575 Stakeholders in Achieving UN SDGs. *Sustainability* 2021; 13: 9095. 10.3390/su13169095

576 Dobon A, Cordero P, Kreft F, Ostergaard SR, Antvorskov H, Robertsson M, et al. The sustainability of
577 communicative packaging concepts in the food supply chain. A case study: part 2. Life cycle
578 costing and sustainability assessment. *International Journal of Life Cycle Assessment* 2011;
579 16: 537-547. 10.1007/s11367-011-0291-9

580 Dorr E, Sanye-Mengual E, Gabrielle B, Grard BJP, Aubry C. Proper selection of substrates and crops
581 enhances the sustainability of Paris rooftop garden. *Agronomy for Sustainable Development*
582 2017; 37. 10.1007/s13593-017-0459-1

583 Ekener E, Hansson J, Larsson A, Peck P. Developing Life Cycle Sustainability Assessment methodology
584 by applying values-based sustainability weighting - Tested on biomass based and fossil
585 transportation fuels. *Journal of Cleaner Production* 2018; 181: 337-351.
586 10.1016/j.jclepro.2018.01.211

587 Escobar N, Bautista I, Pena N, Fenollosa ML, Osca JM, Sanjuan N. Life Cycle Thinking for the
588 environmental and financial assessment of rice management systems in the Senegal River
589 Valley. *Journal of Environmental Management* 2022; 310. 10.1016/j.jenvman.2022.114722

590 Falcone G, De Luca AI, Stillitano T, Iofrida N, Strano A, Piscopo A, et al. Shelf life extension to reduce
591 food losses: The case of Mozzarella Cheese. *Chemical Engineering Transactions* 2017; 57:
592 1849-1854. 10.3303/CET1757309

593 Falcone G, De Luca AI, Stillitano T, Strano A, Romeo G, Gulisano G. Assessment of Environmental and
594 Economic Impacts of Vine-Growing Combining Life Cycle Assessment, Life Cycle Costing and
595 Multicriterial Analysis. *Sustainability* 2016; 8: 793. 10.3390/su8080793

596 FAO. *Building a Common Vision for Sustainable Food and Agriculture. Principles and Approaches*,
597 Rome, 2014.

598 Fenollosa ML, Ribal J, Lidon A, Bautista I, Juraske R, Clemente G, et al. Influence of Management
599 Practices on Economic and Environmental Performance of Crops. A Case Study in Spanish
600 Horticulture. *Agroecology and Sustainable Food Systems* 2014; 38: 635-659.
601 10.1080/21683565.2014.896302

602 Florindo TJ, Florindo G, Ruviaro CF, Pinto AT. Multicriteria decision-making and probabilistic weighing
603 applied to sustainable assessment of beef life cycle. *Journal of Cleaner Production* 2020; 242.
604 10.1016/j.jclepro.2019.118362

605 Florindo TJ, Florindo G, Talamini E, da Costa JS, Ruviaro CF. Carbon footprint and Life Cycle Costing of
606 beef cattle in the Brazilian midwest. *Journal of Cleaner Production* 2017; 147: 119-129.
607 10.1016/j.jclepro.2017.01.021

608 García-Herrero L, Brenes-Peralta L, Leschi F, Vittuari M. Integrating Life Cycle Thinking in a policy
609 decision tool: Its application in the pineapple production in Dominican Republic. *Journal of*
610 *Cleaner Production* 2022; 360. 10.1016/j.jclepro.2022.132094

611 García-Herrero L, Costello C, De Menna F, Schreiber L, Vittuari M. Eating away at sustainability. Food
612 consumption and waste patterns in a US school canteen. *Journal of Cleaner Production* 2021;
613 279. 10.1016/j.jclepro.2020.123571

614 Garcia-Herrero L, De Menna F, Vittuari M. Food waste at school. The environmental and cost impact
615 of a canteen meal. *Waste Management* 2019; 100: 249-258. 10.1016/j.wasman.2019.09.027

616 Gava O, Bartolini F, Venturi F, Brunori G, Zinnai A, Pardossi A. A Reflection of the Use of the Life Cycle
617 Assessment Tool for Agri-Food Sustainability. *Sustainability* 2019; 11: 71.
618 10.3390/su11010071

619 Geß A, Tolsdorf A, Ko N. A life cycle perspective of lamb meat production systems from Turkey and
620 the EU. *Small Ruminant Research* 2022; 208. 10.1016/j.smallrumres.2022.106637

621 Gladek E, Fraser M, Roemers G, Sabag Muñoz O, Kennedy E, Hirsch P. The global food system: an
622 analysis. *Metabolic*, Amsterdam 2016.

623 Godfray HCJ, Beddington JR, Crute IR, Haddad L, Lawrence D, Muir JF, et al. Food Security: The
624 Challenge of Feeding 9 Billion People. *Science* 2010; 327: 812-818. 10.1126/science.1185383

625 Gomersall JS, Jadotte YT, Xue Y, Lockwood S, Riddle D, Preda A. Conducting systematic reviews of
626 economic evaluations. *JBI Evidence Implementation* 2015; 13: 170-178.
627 10.1097/xeb.0000000000000063

628 Gosalvitr P, Cuellar-Franca RM, Smith R, Azapagic A. Integrating process modelling and sustainability
629 assessment to improve the environmental and economic sustainability in the cheese
630 industry. *Sustainable Production and Consumption* 2021a; 28: 969-986.
631 10.1016/j.spc.2021.07.022

632 Gresta F, De Luca AI, Strano A, Falcone G, Santonoceto C, Anastasi U, et al. Economic and
633 environmental sustainability analysis of guar (*Cyamopsis tetragonojoba* L.) farming process in
634 a Mediterranean area: two case studies. *Italian Journal of Agronomy* 2014; 9: 20-24.
635 10.4081/ija.2014.565

636 Guinée JB, Heijungs R, Huppes G, Zamagni A, Masoni P, Buonamici R, et al. Life Cycle Assessment:
637 Past, Present, and Future. *Environmental Science & Technology* 2011; 45: 90-96.
638 10.1021/es101316v

639 Hanif M, Mahlia TMI, Aditiya HB, Chong WT, Nasruddin. Techno-economic and environmental
640 assessment of bioethanol production from high starch and root yield Sri Kanji 1 cassava in
641 Malaysia. *Energy Reports* 2016; 2: 246-253. 10.1016/j.egyr.2016.03.004

642 Holka M. Assessment of carbon footprint and life cycle costs of winter wheat (*Triticum Aestivum* L.)
643 production in different soil tillage systems. *Applied Ecology and Environmental Research*
644 2020; 18: 5841-5855. 10.15666/aeer/1804_58415855

645 Holka M, Bieńkowski J. Carbon footprint and life-cycle costs of maize production in conventional and
646 non-inversion tillage systems. *Agronomy* 2020; 10. 10.3390/agronomy10121877

647 Hong JM, Zhou J, Hong JL. Environmental and economic impact of furfuralcohol production using
648 corncob as a raw material. *International Journal of Life Cycle Assessment* 2015; 20: 623-631.
649 10.1007/s11367-015-0854-2

650 Hoogmartens R, Van Passel S, Van Acker K, Dubois M. Bridging the gap between LCA, LCC and CBA as
651 sustainability assessment tools. *Environmental Impact Assessment Review* 2014; 48: 27-33.
652 10.1016/j.eiar.2014.05.001

653 Hunkeler D, Lichtenvort K, Rebitzer G. *Environmental Life Cycle Costing*: CRC Press, 2008.

654 Iofrida N, Stillitano T, Falcone G, Gulisano G, Nicolo BF, De Luca AI. The socio-economic impacts of
655 organic and conventional olive growing in Italy. *New Medit* 2020; 19: 117-131.
656 10.30682/nm2001h

657 Iotti M, Bonazzi G. The application of Life Cycle Cost (LCC) approach to quality food production: A
658 comparative analysis in the parma pdo ham sector. *American Journal of Applied Sciences*
659 2014; 11: 1492-1506. 10.3844/ajassp.2014.1492.1506

660 ISO. ISO 14040: Environmental management—Life cycle assessment—Principles and framework.
661 *Environmental Management* 2006a; 3: 28.

662 ISO. ISO 14044: 2006-environmental management-life cycle assessment-requirements and
663 guidelines. Genf, 2006b.

664 ISO. ISO 15686–5: Buildings and constructed assets-service-life planning-Part 5: Life-cycle costing,
665 2008.

666 ISO. ISO 15663 Petroleum, petrochemical and natural gas industries—Life cycle costing. International
667 Organization for Standardization (ISO) Geneva, Switzerland, 2021.

668 Jirapornvaree I, Suppadit T, Kumar V. Assessing the economic and environmental impact of jasmine
669 rice production: Life cycle assessment and Life Cycle Costs analysis. *Journal of Cleaner*
670 *Production* 2021; 303. 10.1016/j.jclepro.2021.127079

671 Kim E, Jung J, Hapsari G, Kang S, Kim K, Yoon S, et al. Economic and environmental sustainability and
672 public perceptions of rooftop farm versus extensive garden. *Building and Environment* 2018;
673 146: 206-215. 10.1016/j.buildenv.2018.09.046

674 Kloepffer W. Life cycle sustainability assessment of products. *The International Journal of Life Cycle*
675 *Assessment* 2008; 13: 89. 10.1065/lca2008.02.376

676 Konstantas A, Stamford L, Azapagic A. Economic sustainability of food supply chains: Life cycle costs
677 and value added in the confectionary and frozen desserts sectors. *Science of the Total*
678 *Environment* 2019; 670: 902-914. 10.1016/j.scitotenv.2019.03.274

679 Konstantas A, Stamford L, Azapagic A. A framework for evaluating life cycle eco-efficiency and an
680 application in the confectionary and frozen-desserts sectors. *Sustainable Production and*
681 *Consumption* 2020; 21: 192-203. 10.1016/j.spc.2019.12.006

682 Koričan M, Perčić M, Vladimir N, Soldo V, Jovanović I. Environmental and economic assessment of
683 mariculture systems using a high share of renewable energy sources. *Journal of Cleaner*
684 *Production* 2022; 333. 10.1016/j.jclepro.2021.130072

685 Lask J, Martínez Guajardo A, Weik J, von Cossel M, Lewandowski I, Wagner M. Comparative
686 environmental and economic life cycle assessment of biogas production from perennial wild
687 plant mixtures and maize (*Zea mays* L.) in southwest Germany. *GCB Bioenergy* 2020.
688 10.1111/gcbb.12715

689 Laso J, Garcia-Herrero I, Margallo M, Vazquez-Rowe I, Fullana P, Bala A, et al. Finding an economic
690 and environmental balance in value chains based on circular economy thinking: An eco-
691 efficiency methodology applied to the fish canning industry. *Resources Conservation and*
692 *Recycling* 2018; 133: 428-437. 10.1016/j.resconrec.2018.02.004

693 Le Feon S, Dubois T, Jaeger C, Wilfart A, Akkal-Corfini N, Bacenetti J, et al. DEXiAqua, a Model to
694 Assess the Sustainability of Aquaculture Systems: Methodological Development and
695 Application to a French Salmon Farm. *Sustainability* 2021; 13. 10.3390/su13147779

696 Lee ZY, Liew PY, Woon KS, Tan LS, Tamunaidu P, Klemeš JJ. Life-Cycle Environmental and Cost
697 Analysis of Palm Biomass-based Bio-Ethanol Production in Malaysia. *Chemical Engineering*
698 *Transactions* 2021; 89: 85-90. 10.3303/CET2189015

699 Lerkkasemsan N, Achenie LEK. Life cycle costs and life cycle assessment for the harvesting,
700 conversion, and the use of switchgrass to produce electricity. *International Journal of*
701 *Chemical Engineering* 2013. 10.1155/2013/492058

702 Li J, Li W, Wang L, Jin B. Environmental and cost impacts of food waste in university canteen from a
703 life cycle perspective. *Energies* 2021; 14. 10.3390/en14185907

704 Liaros S, Botsis K, Xydis G. Technoeconomic evaluation of urban plant factories: The case of basil
705 (*Ocimum basilicum*). *Science of the Total Environment* 2016; 554: 218-227.
706 10.1016/j.scitotenv.2016.02.174

707 Lokesh K, West C, Kuylenstierna JC, Fan JJ, Budarin V, Prielcel P, et al. Economic and agronomic impact
708 assessment of wheat straw based alkyl polyglucoside produced using green chemical
709 approaches. *Journal of Cleaner Production* 2019; 209: 283-296.
710 10.1016/j.jclepro.2018.10.220

711 Luo L, van der Voet E, Huppes G. Life cycle assessment and life cycle costing of bioethanol from
712 sugarcane in Brazil. *Renewable & Sustainable Energy Reviews* 2009; 13: 1613-1619.
713 10.1016/j.rser.2008.09.024

714 Meier MS, Stoessel F, Jungbluth N, Juraske R, Schader C, Stolze M. Environmental impacts of organic
715 and conventional agricultural products – Are the differences captured by life cycle
716 assessment? *Journal of Environmental Management* 2015; 149: 193-208.
717 10.1016/j.jenvman.2014.10.006

718 Mohamad RS, Verrastro V, Cardone G, Bteich MR, Favia M, Moretti M, et al. Optimization of organic
719 and conventional olive agricultural practices from a Life Cycle Assessment and Life Cycle
720 Costing perspectives. *Journal of Cleaner Production* 2014; 70: 78-89.
721 10.1016/j.jclepro.2014.02.033

722 Moher D, Liberati A, Tetzlaff J, Altman DG. Preferred reporting items for systematic reviews and
723 meta-analyses: the PRISMA statement. *BMJ* 2009; 339: b2535. 10.1136/bmj.b2535

724 Moosavi-Nezhad M, Salehi R, Aliniaiefard S, Winans KS, Nabavi-Pelesaraei A. An analysis of energy
725 use and economic and environmental impacts in conventional tunnel and LED-equipped
726 vertical systems in healing and acclimatization of grafted watermelon seedlings. *Journal of
727 Cleaner Production* 2022; 361. 10.1016/j.jclepro.2022.132069

728 Mounsree S, Neamhom T, Polprasert S, Patthanaissaranukool W. Carbon footprint and life cycle
729 costing of maize production in Thailand with temporal and geographical resolutions.
730 *International Journal of Life Cycle Assessment* 2022. 10.1007/s11367-022-02021-4

731 Neugebauer S, Forin S, Finkbeiner M. From Life Cycle Costing to Economic Life Cycle Assessment—
732 Introducing an Economic Impact Pathway. *Sustainability* 2016; 8: 428. 10.3390/su8050428

733 Nguyen TLT, Gheewala SH, Bonnet S. Life cycle cost analysis of fuel ethanol produced from cassava in
734 Thailand. *International Journal of Life Cycle Assessment* 2008; 13: 564-573. 10.1007/s11367-
735 008-0035-7

736 Norris GA. Integrating life cycle cost analysis and LCA. *The International Journal of Life Cycle
737 Assessment* 2001; 6: 118-120. 10.1007/BF02977849

738 Notarnicola B, Sala S, Anton A, McLaren SJ, Saouter E, Sonesson U. The role of life cycle assessment in
739 supporting sustainable agri-food systems: A review of the challenges. *Journal of Cleaner
740 Production* 2017; 140: 399-409. 10.1016/j.jclepro.2016.06.071

741 Olba-Ziety E, Stolarski MJ, Krzyzaniak M, Roj E, Tyskiewicz K, Luczynski MK. Supercritical production
742 of extract from poplar containing bioactive substances - An economic analysis. *Industrial
743 Crops and Products* 2022; 184. 10.1016/j.indcrop.2022.115094

744 Omolayo Y, Feingold BJ, Neff RA, Romeiko XX. Life cycle assessment of food loss and waste in the
745 food supply chain. *Resources, Conservation and Recycling* 2021; 164: 105119.
746 10.1016/j.resconrec.2020.105119

747 Omran N, Sharaai AH, Hashim AH. Visualization of the Sustainability Level of Crude Palm Oil
748 Production: A Life Cycle Approach. *Sustainability* 2021; 13. 10.3390/su13041607

749 Onat NC, Kucukvar M, Halog A, Cloutier S. Systems Thinking for Life Cycle Sustainability Assessment:
750 A Review of Recent Developments, Applications, and Future Perspectives. *Sustainability*
751 2017; 9: 706. 10.3390/su9050706

752 Pari L, Alexopoulou E, Stefanoni W, Latterini F, Cavalaris C, Palmieri N. The Eco-Efficiency of Castor
753 Supply Chain: A Greek Case Study. *Agriculture-Basel* 2022; 12. 10.3390/agriculture12020206

754 Pari L, Suardi A, Stefanoni W, Latterini F, Palmieri N. Environmental and Economic Assessment of
755 Castor Oil Supply Chain: A Case Study. *Sustainability* 2020; 12. 10.3390/su12166339
756 Pattanaik L, Padhi SK, Hariprasad P, Naik SN. Life cycle cost analysis of natural indigo dye production
757 from *Indigofera tinctoria* L. plant biomass: a case study of India. *Clean Technologies and*
758 *Environmental Policy* 2020; 22: 1639-1654. 10.1007/s10098-020-01914-y
759 Peña A, Rovira-Val MR. A longitudinal literature review of life cycle costing applied to urban
760 agriculture. *International Journal of Life Cycle Assessment* 2020. 10.1007/s11367-020-01768-
761 y
762 Pergola M, D'Amico M, Celano G, Palese AM, Scuderi A, Di Vita G, et al. Sustainability evaluation of
763 Sicily's lemon and orange production: An energy, economic and environmental analysis.
764 *Journal of Environmental Management* 2013; 128: 674-682. 10.1016/j.jenvman.2013.06.007
765 Pexas G, Mackenzie SG, Wallace M, Kyriazakis I. Accounting for spatial variability in life cycle cost-
766 effectiveness assessments of environmental impact abatement measures. *International*
767 *Journal of Life Cycle Assessment* 2021; 26: 1236-1253. 10.1007/s11367-021-01915-z
768 Pizzol M, Weidema B, Brandão M, Osset P. Monetary valuation in Life Cycle Assessment: a review.
769 *Journal of Cleaner Production* 2015; 86: 170-179. 10.1016/j.jclepro.2014.08.007
770 Rahmah DM, Putra AS, Ishizaki R, Noguchi R, Ahamed T. A Life Cycle Assessment of Organic and
771 Chemical Fertilizers for Coffee Production to Evaluate Sustainability toward the Energy-
772 Environment-Economic Nexus in Indonesia. *Sustainability* 2022; 14. 10.3390/su14073912
773 Ribeiro I, Sobral P, Pecas P, Henriques E. A sustainable business model to fight food waste. *Journal of*
774 *Cleaner Production* 2018; 177: 262-275. 10.1016/j.jclepro.2017.12.200
775 Rivera XCS, Azapagic A. Life cycle costs and environmental impacts of production and consumption of
776 ready and home-made meals. *Journal of Cleaner Production* 2016; 112: 214-228.
777 10.1016/j.jclepro.2015.07.111
778 Roffeis M, Wakefield ME, Almeida J, Valada TRA, Devic E, Kone N, et al. Life cycle cost assessment of
779 insect based feed production in West Africa. *Journal of Cleaner Production* 2018; 199: 792-
780 806. 10.1016/j.jclepro.2018.07.179
781 Roos A, Ahlgren S. Consequential life cycle assessment of bioenergy systems – A literature review.
782 *Journal of Cleaner Production* 2018; 189: 358-373. 10.1016/j.jclepro.2018.03.233
783 Roselli L, Casieri A, de Gennaro BC, Sardaro R, Russo G. Environmental and Economic Sustainability of
784 Table Grape Production in Italy. *Sustainability* 2020; 12. 10.3390/su12093670
785 Ruviaro CF, de Leis CM, Florindo TJ, de Medeiros Florindo GIB, da Costa JS, Tang WZ, et al. Life cycle
786 cost analysis of dairy production systems in Southern Brazil. *Science of the Total Environment*
787 2020; 741. 10.1016/j.scitotenv.2020.140273
788 Saber Z, Esmaeili M, Pirdashti H, Motevali A, Nabavi-Pelesaraei A. Exergoenvironmental-Life cycle
789 cost analysis for conventional, low external input and organic systems of rice paddy
790 production. *Journal of Cleaner Production* 2020; 263. 10.1016/j.jclepro.2020.121529
791 Sala S, Farioli F, Zamagni A. Life cycle sustainability assessment in the context of sustainability science
792 progress (part 2). *The International Journal of Life Cycle Assessment* 2013a; 18: 1686-1697.
793 10.1007/s11367-012-0509-5
794 Sala S, Farioli F, Zamagni A. Progress in sustainability science: lessons learnt from current
795 methodologies for sustainability assessment: Part 1. *The International Journal of Life Cycle*
796 *Assessment* 2013b; 18: 1653-1672. 10.1007/s11367-012-0508-6
797 Sanyé-Mengual E, Gasperi D, Michelon N, Orsini F, Ponchia G, Gianquinto G. Eco-efficiency
798 assessment and food security potential of home gardening: A case study in Padua, Italy.
799 *Sustainability (Switzerland)* 2018; 10. 10.3390/su10072124
800 Sanyé-Mengual E, Oliver-Solà J, Montero JJ, Rieradevall J. An environmental and economic life cycle
801 assessment of rooftop greenhouse (RTG) implementation in Barcelona, Spain. Assessing new
802 forms of urban agriculture from the greenhouse structure to the final product level.
803 *International Journal of Life Cycle Assessment* 2015; 20: 350-366. 10.1007/s11367-014-0836-
804 9

805 Schau EM, Fet AM. LCA studies of food products as background for environmental product
806 declarations. *The International Journal of Life Cycle Assessment* 2008; 13: 255-264.
807 10.1065/lca2007.12.372

808 Schulte M, Lewandowski I, Pude R, Wagner M. Comparative life cycle assessment of bio-based
809 insulation materials: Environmental and economic performances. *Global Change Biology*
810 *Bioenergy* 2021; 13: 979-998. 10.1111/gcbb.12825

811 Silva VL, Sanjuán N. Opening up the black box: A systematic literature review of life cycle assessment
812 in alternative food processing technologies. *Journal of Food Engineering* 2019; 250: 33-45.
813 10.1016/j.jfoodeng.2019.01.010

814 Smith M, Lal P. Environmental and economic assessment of hard apple cider using an integrated LCA-
815 LCC approach. *Sustainable Production and Consumption* 2022; 32: 282-295.
816 10.1016/j.spc.2022.04.026

817 Soldatos P. Economic Aspects of Bioenergy Production from Perennial Grasses in Marginal Lands of
818 South Europe. *Bioenergy Research* 2015; 8: 1562-1573. 10.1007/s12155-015-9678-y

819 Stillitano T, De Luca AI, Falcone G, Spada E, Gulisano G, Strano A. Economic profitability assessment
820 of mediterranean olive growing systems. *Bulgarian Journal of Agricultural Science* 2016; 22:
821 517-526.

822 Stillitano T, Falcone G, De Luca AI, Piga A, Conte P, Strano A, et al. Innovative technologies in EVO oil
823 extraction: an economic and environmental impact analysis. *Rivista Italiana Delle Sostanze*
824 *Grasse* 2019; 96: 223-230.

825 Strano A, Falcone G, Nicolo BF, Stillitano T, De Luca AI, Nesci FS, et al. Eco-profiles and economic
826 performances of a high-value fruit crop in southern Italy: a case study of bergamot (*Citrus*
827 *bergamia* Risso). *Agroecology and Sustainable Food Systems* 2017; 41: 1124-1145.
828 10.1080/21683565.2017.1357064

829 Strano A, Stillitano T, De Luca AI, Falcone G, Gulisano G. Profitability analysis of small-scale
830 beekeeping firms by using life cycle costing (LCC) methodology. *American Journal of*
831 *Agricultural and Biological Science* 2015; 10: 116-127. 10.3844/ajabssp.2015.116.127

832 Styles D, Thorne F, Jones MB. Energy crops in Ireland: An economic comparison of willow and
833 *Miscanthus* production with conventional farming systems. *Biomass & Bioenergy* 2008; 32:
834 407-421. 10.1016/j.biombioe.2007.10.012

835 Sutherland CS, Yukich J, Goeree R, Tediosi F. A literature review of economic evaluations for a
836 neglected tropical disease: human African trypanosomiasis ("sleeping sickness"). *PLoS*
837 *neglected tropical diseases* 2015; 9: e0003397-e0003397. 10.1371/journal.pntd.0003397

838 Swarr TE, Hunkeler D, Klöpffer W, Pesonen H-L, Ciroth A, Brent AC, et al. Environmental life-cycle
839 costing: a code of practice. *The International Journal of Life Cycle Assessment* 2011; 16: 389-
840 391. 10.1007/s11367-011-0287-5

841 Tamburini E, Pedrini P, Marchetti MG, Fano EA, Castaldelli G. Life Cycle Based Evaluation of
842 Environmental and Economic Impacts of Agricultural Productions in the Mediterranean Area.
843 *Sustainability* 2015; 7: 2915-2935. 10.3390/su7032915

844 Tziolas E, Bournaris T. Economic and Environmental Assessment of Agro-Energy Districts in Northern
845 Greece: a Life Cycle Assessment Approach. *Bioenergy Research* 2019; 12: 1145-1162.
846 10.1007/s12155-019-10020-x

847 Valente C, Moller H, Johnsen FM, Saxegard S, Brunson ER, Alvseike OA. Life cycle sustainability
848 assessment of a novel slaughter concept. *Journal of Cleaner Production* 2020; 272.
849 10.1016/j.jclepro.2020.122651

850 Venanzi S, Pezzolla D, Cecchini L, Pauselli M, Ricci A, Sordi A, et al. Use of agricultural by-products in
851 the development of an agro-energy chain: A case study from the Umbria region. *Science of*
852 *the Total Environment* 2018; 627: 494-505. 10.1016/j.scitotenv.2018.01.176

853 Verduna T, Blanc S, Merlino VM, Cornale P, Battaglini LM. Sustainability of Four Dairy Farming
854 Scenarios in an Alpine Environment: The Case Study of Toma di Lanzo Cheese. *Frontiers in*
855 *Veterinary Science* 2020; 7. 10.3389/fvets.2020.569167

856 Wagner M, Kamp L, Graeff-Honninger S, Lewandowski I. Environmental and Economic Performance
857 of Yacon (*Smallanthus sonchifolius*) Cultivated for Fructooligosaccharide Production.
858 Sustainability 2019a; 11. 10.3390/su11174581
859 Wagner M, Mangold A, Lask J, Petig E, Kiesel A, Lewandowski I. Economic and environmental
860 performance of miscanthus cultivated on marginal land for biogas production. Global Change
861 Biology Bioenergy 2019b; 11: 34-49. 10.1111/gcbb.12567
862 Wohner B, Gabriel VH, Krenn B, Krauter V, Tacker M. Environmental and economic assessment of
863 food-packaging systems with a focus on food waste. Case study on tomato ketchup. Science
864 of the Total Environment 2020; 738. 10.1016/j.scitotenv.2020.139846
865 World Business Council for Sustainable Development. Eco-efficiency: Learning Module. Conches-
866 Geneva Switzerland, 2006.
867 Yang N, Li F, Liu Y, Dai T, Wang Q, Zhang J, et al. Environmental and Economic Life-Cycle Assessments
868 of Household Food Waste Management Systems: A Comparative Review of Methodology and
869 Research Progress. Sustainability 2022; 14: 7533. 10.3390/su14137533
870 Zhen H, Gao W, Jia L, Qiao Y, Ju X. Environmental and economic life cycle assessment of alternative
871 greenhouse vegetable production farms in peri-urban Beijing, China. Journal of Cleaner
872 Production 2020; 269. 10.1016/j.jclepro.2020.122380
873 Zira S, Rydhmer L, Ivarsson E, Hoffmann R, Roos E. A life cycle sustainability assessment of organic
874 and conventional pork supply chains in Sweden. Sustainable Production and Consumption
875 2021; 28: 21-38. 10.1016/j.spc.2021.03.028
876 Zortea RB, Maciel VG, Passuello A. Sustainability assessment of soybean production in Southern
877 Brazil: A life cycle approach. Sustainable Production and Consumption 2018; 13: 102-112.
878 10.1016/j.spc.2017.11.002
879
880